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iChip increases the success of cultivation of TBT-resistant and TBT-degrading bacteria from estuarine sediment.

Polrot, A., Kirby, J.R.¹, Olorunniji F.J., Birkett, J.W. and Sharples, G.P.

¹Liverpool John Moores University, Byrom Street, Liverpool, L3 3AF. Corresponding author: J.R.Kirby@ljmu.ac.uk

Abstract:

Standard methods of microbial cultivation only enable the isolation of a fraction of the total environmental bacteria. Numerous techniques have been developed to increase the success of isolation and cultivation in the laboratory, some of which derive from diffusion chambers. In a diffusion chamber, environmental bacteria in agar medium are put back in the environment to grow as close to their natural conditions as possible, only separated from the environment by semi-permeable membranes. In this study, the iChip, a device that possesses hundreds of mini diffusion chambers, was used to isolate tributyltin (TBT) resistant and degrading bacteria. IChip was shown to be efficient at increasing the number of cultivable bacteria compared to standard methods. TBT-resistant strains belonging to *Oceanisphaera sp.*, *Pseudomonas sp.*, *Bacillus sp.* and *Shewanella sp.* were identified from Liverpool Dock sediment. Among the isolates in the present study, only members of *Pseudomonas sp.* were able to use TBT as a sole carbon source. It is the first time that members of the genus *Oceanisphaera* have been shown to be TBT-resistant. Although iChip has been used in the search for molecules of biomedical interest here we demonstrate its promising application in bioremediation.

Keypoints:

- iChip is efficient to improve the success of isolation and cultivation of sediment bacteria
- iChip can be used to isolate microorganisms of interest in bioremediation
- Members of *Oceanisphaera* are resistant to TBT but cannot use it as sole carbon source

27 **Keywords:**

28 Bacterial isolation, bioremediation, high throughput isolation, isolation chip, tributyltin

29 **Introduction:**

30 Tributyltin is an organotin compound that has been used widely as a biocide in antifouling paints. It is therefore
31 highly toxic and has been shown to be a major threat to aquatic ecosystems. Due to its toxicity, it was subjected
32 to a global ban in 2008 (Sonak et al. 2009). However, TBT is still a major concern in many locations around the
33 world (Filipkowska and Kowalewska 2019). In fact, it is still authorised in a small number of countries (Turner
34 and Glegg 2014), and it is suspected to be used illegally in many others because of its high efficiency (Egardt et
35 al. 2017). The main concern is its high persistence in anoxic sediments and as such, is a pernicious legacy
36 contaminant. Indeed, TBT is hydrophobic and strongly binds to organic matter and sediment where it can remain
37 for decades (Langston et al. 2015). Sediment therefore acts as a secondary source of contamination during
38 resuspension events, causing more disturbance to aquatic ecosystems. There is therefore a need to remediate
39 sediment contaminated with TBT.

40 Traditional remediation techniques such as incineration (Song et al. 2005) or electrochemical oxidation
41 (Beuselinck and Valle 2008) are usually regarded as efficient but costly. In addition, they can cause environmental
42 issues as they involve the excavation of sediment, which causes problems of contaminant spreading and further
43 pollution due to carbon emissions during transportation (Manap and Voulvoulis 2015). The more environmentally
44 sustainable approach is bioremediation, where contaminants are broken down by the activity of biological
45 organisms. In particular, *in situ* bioremediation removes the need for excavation plus the associated cost and
46 environmental issues linked to it (Polrot et al. 2021). Bioremediation can be further subdivided into
47 phytoremediation, when using plants (Pilon-Smits 2005), or biodegradation, when using microorganisms (Adams
48 et al. 2015). The latter is especially pertinent for *in situ* bioremediation of port sediment. Biodegradation includes
49 natural attenuation, biostimulation and bioaugmentation (Tyagi et al. 2011; Adams et al. 2015). Natural attenuation
50 consists of using the native microbial community to naturally degrade harmful contaminants (Lofrano et al. 2017).
51 Biostimulation aims at boosting the degrading activity of the microbial community by providing more favourable
52 conditions, for example by the addition of nutrients (Adams et al. 2015), or through oxygenation (Scow and Hicks
53 2005). Finally, bioaugmentation consists of adding specific microorganisms to decontaminate the material (Tyagi

54 et al. 2011; Adams et al. 2015). The added microorganisms are selected for their exceptional abilities to efficiently
55 degrade the contaminants of interest.

56 The use of bioremediation requires a comprehensive understanding of the degradation pathways and kinetics, the
57 microbial communities involved in the degradation as well as the most favourable conditions for the growth and
58 degrading activity of the microorganisms involved. A first step towards this objective is to proceed with the
59 isolation and cultivation of the microbial degraders. Thus, research has been carried out to isolate and characterise
60 TBT-resistant and degrading microorganisms (Cruz et al. 2015). Among the identified microbes include *Chlorella*
61 species (Tsang et al. 1999; Jin et al. 2011) and fungi such as *Cunninghamella elegans* or *Cochliobolus lunatus*
62 (Bernat and Długoński 2002; Bernat et al. 2013). In addition, many bacteria have been studied for their TBT
63 degradation ability, such as *Aeromonas molluscorum*, *Enterobacter cloacae* and numerous species of
64 *Pseudomonas* (Finnegan et al. 2018).

65 Despite this, it is well-known that only a small proportion of microbes have been discovered so far. Indeed, it is
66 estimated that more than 99% of bacteria remain unknown (Locey and Lennon 2016). The main reason for this is
67 our inability to cultivate them in the laboratory. Classic methods of isolation and cultivation, that were used for
68 the isolation of TBT-degrading bacteria so far, failed to provide the appropriate conditions for the growth of the
69 majority of the environmental bacteria and are biased towards the same species. Nevertheless, some techniques
70 have been developed to improve the success of cultivation of novel species, usually by mimicking as accurately
71 as possible the natural environment (Hahn et al. 2019; Bodor et al. 2020). Among these, the diffusion chamber
72 concept was of special interest. In diffusion chambers, microorganisms are trapped in agar while in contact with
73 their natural environment with semipermeable membranes. The membranes ensure that cells cannot move in or
74 out of the diffusion chamber but small molecules that may be necessary for microbial growth can enter the chamber
75 (Kaeberlein et al. 2002). On the basis of this concept, iChip was created, acting like hundreds of mini diffusion
76 chambers and therefore allowing the high-throughput isolation of bacteria (Nichols et al. 2010). IChip allowed the
77 cultivation of different species of bacteria than standard plating methods (Nichols et al. 2010).

78 The first aim of this study was to evaluate the beneficial potential of using iChip for the isolation of bacteria of
79 interest in the field of bioremediation and more specifically for TBT biodegradation. A second aim was to advance
80 the knowledge on TBT biodegradation in estuarine sediment with the isolation of TBT-resistant and TBT-
81 degrading bacteria. To fulfil these objectives, a comparison of the standard plating and iChip techniques was

82 performed by measuring the difference in culturability of sediment bacteria using the two techniques. TBT-
83 resistant/degrading bacteria were then screened among the obtained isolates.

84 **Material and Methods:**

85 Sediment sampling and preparation

86 Sediment samples (textural class ‘slightly sandy mud’ (Flemming 2000), comprising 14.3 % clay, 79.5 % silt, 6.2
87 % sand) were taken from Liverpool Brocklebank Dock. The samples had a pH of 7.8, salinity of 27 psu, total
88 nitrogen content of 0.26 %, total carbon content of 3.92% and total organic carbon (TOC) content of 3.12 %.
89 Sediment from Liverpool port was chosen for this study because TBT hotspots are usually concentrated around
90 ports and harbours (Filipkowska and Kowalewska 2019). Sampling locations in the docks were chosen according
91 to TBT contamination data from 2010 (data provided by Peel Ports). Organotin measurement revealed that the
92 contamination in these samples was below detection level at the time of the sampling. This supports the hypothesis
93 that the local microbial community is capable of TBT biodegradation and those samples were therefore selected
94 for the present study. One sample remained untouched in a cold room, stored in the dark at a temperature of 4°C.
95 For microcosm experiments measuring TBT biodegradation in different environmental scenarios (Polrot, 2022),
96 another sample was sieved at 2 mm and spiked with 10 µg TBTCI / g dw sediment (concentration corresponding
97 to a heavy contamination scenario and constrained by the detection limit of the organotin measurement method
98 used) and thoroughly mixed by hand before being put back in the cold store for 4 weeks as an equilibration step.
99 After that equilibration step, the mud was incubated at 20°C for 3 months. At the end of this incubation period, the
100 sample was used for the present study and is referred as “prepared sediment” for the rest of this paper. When using
101 sediment stored directly after sampling and not processed further, the term “untouched sediment” is used.

102 Sediment dilution and standard plating

103 Serial dilutions of the two types of sediment were plated on Tryptic Soy Agar (TSA) and TSA + 1mM TBT in
104 order to calculate the abundance of bacteria capable of growth in standard laboratory conditions. After inoculation
105 of different sediment dilutions in triplicates, the agar plates were incubated at room temperature for 3 to 5 days
106 before the enumeration of colonies was performed.

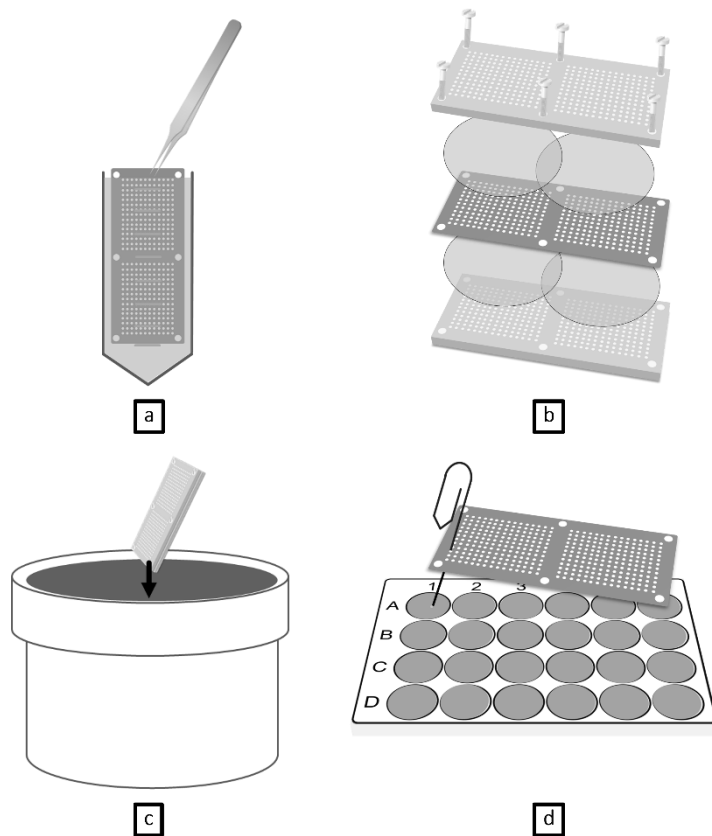
107 The result of this enumeration was used to calculate the appropriate dilution for the inoculation of one “cultivable”
108 bacterial cell in 10% of the iChip through-holes (10^2 bacteria per mL).

109 iChip assembly and incubation:

110 IChips were manufactured in the general engineering workshops of Liverpool John Moores University using the
111 instructions provided by Nichols et al. (2010). Figure 1b indicates all of the components of an iChip, the central
112 plate and the two external ones, which are pierced with a multitude of through-holes arranged in two arrays, in this
113 case two arrays of 192 through-holes. Before assemblage, all the components were sterilized by immersion in 70%
114 ethanol for 15 minutes. They were then allowed to dry under a sterile hood after which the central plate was
115 immersed in molten agar (**Fig. 1a**) containing the appropriate sediment dilution as a means to load one cultivable
116 bacterial cell in 10% of the through-holes (10^2 cultivable cells/mL). Once the agar solidified on the central plate,
117 the excess was removed using a sterile microscope slide and 8 sterile polycarbonate membranes disks of 27cm
118 diameter with 0.03 μ m diameter pores were placed on each side. The external plates were finally mounted at the
119 bottom and top of the central plates and the whole assemblage was screwed together (**Fig. 1b**). To avoid any
120 leaking from the sides, petroleum jelly was applied to seal the edges of the iChip, which was then protected with
121 a fine band of parafilm. After assemblage, the iChips were immersed in a bucket of sediment and stored at 20°C
122 for a week (**Fig. 1c**).

123

124



125

126 **Fig. 1: Steps to sediment bacteria isolation and cultivation using an iChip**

127 The central plate is loaded with fusion agarose medium inoculated with sediment bacterial dilution (a). The iChip is then
 128 assembled with 0.03 μm polycarbonate membranes and the external plate, screwed together (b), and immersed in a bucket of
 129 muddy sediment for 2 weeks (c). After incubation, the iChip is thoroughly rinsed with sterile water, disassembled and sterile
 130 gauge clips are used to deposit each agar plug in a well of a 24-well plate filled up with TSA (d).

131

132 Isolate recovery:

133 After the incubation period, the iChips were thoroughly rinsed in sterile distilled water and disassembled. About
 134 one hundred random cores were retrieved from each iChip using a sterile and unbound gauge paper clip and gently
 135 crushed on the surface of TSA medium in 24-well plates (**Fig. 1d**). The 24-well plates were incubated for several
 136 weeks at room temperature in the dark. The percentage of positive wells at this step was used to calculate the
 137 difference in cultivability between iChip and standard plating.

138 Screening for TBT resistance and use as sole carbon source:

139 Each isolate that could be grown on the 24 well plates containing TSA from the iChip cores were subcultured on
 140 TSA + 1 mM TBT to screen for the resistance phenotype.

141 The isolates that could grow on TSA + 1 mM TBT were further subcultured on Minimal Salt Medium (MSM)
142 containing 1 mM TBT as the sole carbon source. MSM was prepared with the following compounds per litre of
143 distilled water: 0.06g ferrous sulphate; 12.6g dipotassium hydrogen orthophosphate; 3.64g potassium dihydrogen
144 orthophosphate; 2g ammonium nitrate; 0.2g magnesium sulphate; 0.0012g sodium molybdate; 0.0012g manganese
145 sulphate; 0.15g calcium chloride; 15g agar. 1L of medium containing only agar and the phosphate buffer was
146 autoclaved, all the other elements were prepared in solution separately, filter sterilized and added to the fusion
147 medium after autoclaving and before pouring into petri dishes.

148 Genus identification of the isolates:

149 DNA extraction

150 24 colonies growing on TSA + 1 mM TBT were selected to be further identified by 16S rRNA gene sequencing.
151 20 isolates coming from the isolation through iChip, and 4 isolates obtained using the classic method of isolation.
152 Freshly grown colonies were resuspended in 30 µL of sterile water and heated at 95°C for 10 minutes to extract
153 their DNA. The suspensions were then spun down for 2 minutes in a benchtop centrifuge at maximum speed and
154 the supernatant was used as template DNA for the PCR steps.

155 DNA amplification

156 The amplification was performed using the following universal primers: 27F (AGAGTTTGATCATGGCTCA)
157 and 1492R (TACGGTTACCTTGTTACGACTT). The reaction was prepared in a volume of 50 µL in total, with
158 25 µL of ReadyMix™ (Sigma), 1 µL of 10 pM of each primer and 2µL of DNA. Reactions were then performed
159 in a thermocycler with the following program: 94°C for 2 min of initial denaturation followed by 35 cycles at 94°C
160 for 1 min, 58°C for 30 sec and 72°C for 1 min, finishing with a final extension at 72°C for 10 min. The
161 amplification of the samples was detected along with a DNA molecular weight standard (1 kb+, Invitrogen) by
162 electrophoresis in a 2% agarose gel stained with SYBR Safe (Invitrogen) and visualized by transillumination by
163 UV light.

164 The DNA concentration was then measured using a Nanodrop. As all the concentrations were too low, the samples
165 were evaporated and resuspended in the appropriate volume to obtain 25 ng/µL. 5 µL of each sample was then
166 added to 5 µL of primer at 5pmol/µL. 24 tubes were prepared with the forward primer 27F and 24 others with the
167 reverse primer 1492R. The 48 tubes were barcoded using the LightRun barcodes from Eurofins Genomics and
168 sent to the company for Sanger sequencing.

169 Sequence analyses

170 The ab1 files received from Sanger sequencing were checked for quality and the sequences appropriately corrected.
171 The forward and reverse sequences of the same isolates were aligned and reassembled using BioEdit and the
172 resulting FASTA sequences were analysed by BLAST using the total database, excluding
173 uncultured/environmental sample sequences.

174 Nucleotide sequence accession number

175 The sequences were deposited in GenBank and their accession numbers are detailed in **Table 1**.

176 Statistical analyses:

177 All statistical analyses were performed using R Studio. Significant differences in the cultivability of bacteria using
178 the two methods were calculated with a Student's t-test. Statistical significance was assumed when the p-value
179 was below or equal to 0.05.

180 **Results:**

181 Abundance of cultivable bacteria on TSA medium:

182 The abundance of cultivable bacteria increased significantly (prepared sediment: p-value = 0.003, native sediment:
183 p-value = 0.007) when using one round of culturing in iChip compared to standard plating on TSA plates (**Fig. 1**).
184 The number of CFU increased by a factor of 5.5 and 9.5 for the experiment involving untouched sediment and
185 prepared sediment respectively (**Fig. 2**). A higher abundance of cultivable bacteria was also observed for the
186 method using prepared sediment compared to untouched sediment.

187 Proportion of TBT-resistant bacteria cultivated using iChip compared to standard plating

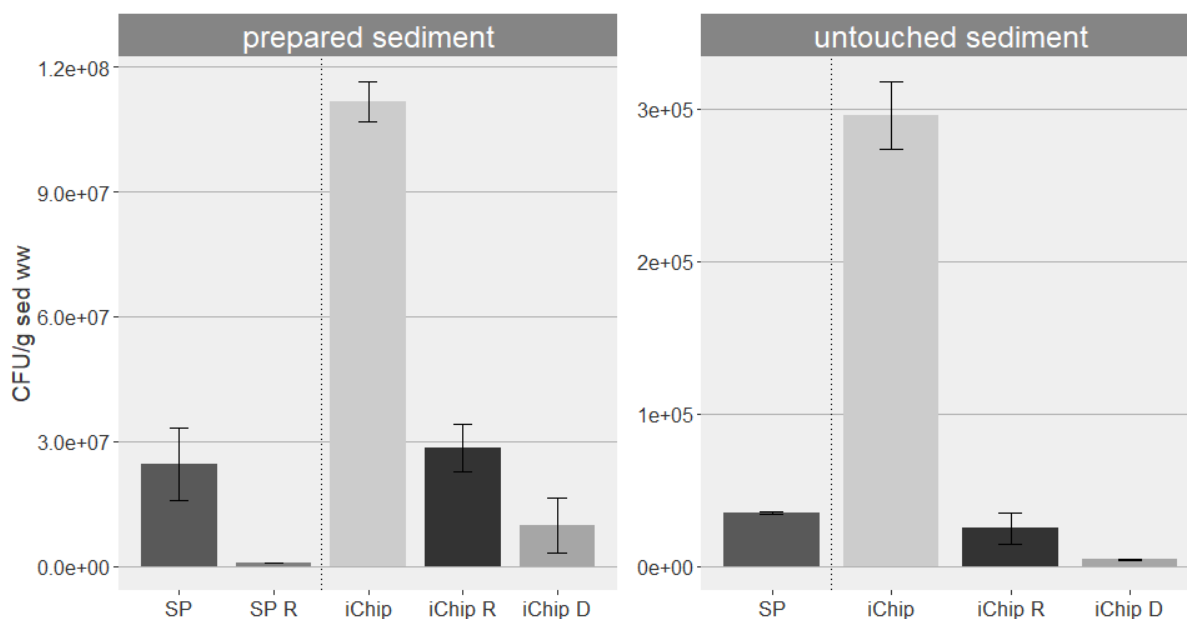
188 From untouched sediment, no TBT-resistant bacteria could be grown using the standard method of plating
189 sediment dilutions on TSA in petri dishes. However, the proportion of TBT-resistant bacteria among the isolates
190 firstly grown on TSA without TBT was not checked. Note as well that no isolates could be obtained when preparing
191 an iChip using TSA containing 1 mM TBT.

192 Nevertheless, TBT-resistant proportions among iChip isolates on TSA using the two different types of sediment
193 can be compared. A higher proportion of TBT-resistant bacteria was found for prepared sediment (p-value= 0.038),

194 with 38.2% of TBT-resistant isolates obtained from prepared sediment compared to 16.3% for untouched sediment
195 (Fig. 2).

196 Proportion of bacteria using TBT as sole carbon source

197 Although the mean number of bacteria capable of using TBT as the sole carbon source appeared higher for prepared
198 sediment compared with untouched sediment (9.3% and 2.0% of the isolates respectively), no statistical difference
199 could be detected due to the high variability within the triplicates (Fig. 2, p-value = 0.36).



200 **Fig. 2: Difference in cultivability between standard plating and iChip method using prepared or untouched sediment**
201 SP: CFU numbers obtained by Standard Plating; SP R: TBT-resistant CFU numbers obtained by Standard Plating on TSA + 1
202 mM TBT; iChip: CFU obtained after one round of iChip and subculturing on TSA; iChip R: TBT-resistant CFU numbers from
203 the subculturing of isolates coming from iChip; iChip D: CFU numbers for cells able to use TBT as sole carbon source from
204 the subculturing of isolates coming from iChip. Results shown represent the mean of triplicates and the error bars are the
205 standard deviations.
206

207

208 Identification of the isolates through 16S rRNA genes Sanger sequencing

209 After several rounds of cultivation, some isolates could not be recovered. The remaining isolates growing on TSA
210 after four rounds of cultivation were therefore further identified. 18 of them came from iChip experiment using
211 prepared sediment, two came from iChip experiment using untouched sediment and four came from standard
212 plating isolation (Table 1).

213 As shown in Table 1, after the Sanger sequencing of 16S rRNA genes, four distinctive genera were identified by
214 BLAST analyses: *Pseudomonas sp.*; *Shewanella sp.*; *Bacillus sp.* and *Oceanisphaera sp.* All of them were able to
215 grow on TSA +1mM TBT for at least 4 subculturing attempts. However, some of them stopped growing after this
216 4th step, but they could still grow on TSA without TBT.

217 The four isolates coming from standard plating were identified as *Pseudomonas sp.* In the names of the isolates,
218 the first letter represents the label of an iChip (α , β , γ , Y). A correlation seems to be observed between the isolate's
219 genera and the iChip experiment.

220 **Discussion:**

221 IChip increases the abundance of culturable bacteria

222 The period of culturing in iChip constitutes a good adaptation step prior to growth of bacteria on synthetic media.
223 While a bacterium is trapped in TSA in an iChip buried in sediment, molecules that may be necessary for their
224 growth can diffuse across the polycarbonate membranes and into the medium. As the growing conditions are closer
225 to those of the natural environment, it is not surprising greater cultivation success is achieved. The real benefit of
226 using the technique is the fact that after sub-culturing iChip agar plugs on TSA in full laboratory conditions, a
227 greater variety of bacteria are able to grow compared to the attempts at isolation without using the intermediate
228 step in iChip.

229 The mechanisms behind this adaptation are unclear. It is also important to note that among the initial isolates which
230 could grow after the direct subculturing from iChip, a number of others failed to grow after a couple of subculturing
231 attempts. As our interest was focused on TBT resistant bacteria, only these were subcultured. Failure to maintain
232 bacterial isolates after subculturing is often described but there are a lack of explanations for this issue (Overmann
233 et al. 2017; Hahn et al. 2019). As the subculturing was performed on TSA + 1 mM TBTCl, some hypotheses can
234 be proposed to explain this lack of growth, in addition to an unknown cause. First, the subculturing may have been
235 delayed, and the bacteria could not be recovered after being kept in the fridge for a few weeks. Second, during the
236 subculturing, a very small quantity of key molecules necessary for the growth of some isolates may have been
237 utilised during the initial subculturing stages but eventually became depleted. Finally, given the selectivity of the
238 medium used, the bacteria could simply have lost their ability to grow in the presence of TBT. This explanation
239 was confirmed for some of the isolates, which after the fourth subculturing stage could be grown on TSA but not
240 on TSA + 1 mM TBTCl. This loss of resistance is most likely to occur through the loss of a plasmid, therefore
241 suggesting that the resistance genes are located on a plasmid for at least some of these strains. Plasmid loss is a
242 well-studied phenomenon due to the wide use of plasmids in research but our understanding remains incomplete
243 (Carroll and Wong 2018). Plasmids are usually well maintained in the presence of a selective pressure, here TBT,
244 but if the isolation plates are kept long enough for TBT degradation to occur, the selective pressure could be
245 reduced around the isolates, which would increase the chance of plasmid loss (Hanak and Cranenburgh 2001).

246 A higher proportion of TBT-resistant bacteria are found among isolates obtained from prepared
247 sediment

248 In the literature, bacteria are usually called resistant when growing on a medium containing a biocide concentration
249 that kills 90% of the population (Cruz et al. 2015). For the purpose of this study, however, TBT-resistant bacteria
250 are those bacteria that grow on a medium containing 1 mM TBTCl. Observing a higher proportion of TBT-resistant
251 bacteria among the isolates obtained from prepared sediment compared to the ones obtained from untouched
252 sediment is to be expected.

253 Different mechanisms can lead to bacterial resistance to TBT. There are at least four theoretical categories of
254 resistance mechanism: 1) TBT exclusion/efflux from the cell; 2) TBT degradation into DBT, MBT and inorganic
255 tin; 3) TBT metabolization and use as a carbon source and 4) bioaccumulation using metallothionein-like proteins
256 (Cruz et al. 2015). Determining the resistance mechanism used by the bacteria isolated in this study would require
257 further testing. Previous studies of TBT-resistant bacteria have been able to identify some genes and molecules
258 involved in the resistance mechanisms. Transcriptomic studies have looked at the difference in gene expression in
259 the presence of TBTCl. Bernat et al. (2014) reported a clear change in membrane phospholipid composition as
260 well as production of peroxidase. The peroxidase could have a protective role against the generation of reactive
261 oxygen species that have been reported to play a critical role in TBTCl toxicity. Efflux pumps have been identified
262 as a basis of the resistance in two bacterial species, coded by the operon *tbtABM* in some *Pseudomonas stutzeri*
263 strains (Jude et al. 2004) and coded by the gene *SugE* in *Aeromonas molluscorum* (Cruz et al. 2013).

264 For a bioremediation purpose, the mechanism of most interest is the degradation of the compound. A quick way
265 of checking for degradation ability is to provide TBT as sole carbon source in the growth medium. Therefore,
266 further tests were carried out to identify this type of TBT-degrader among the isolated strains.

267 Some of the isolates are able to use TBT as the sole carbon source

268 As a straightforward way of screening TBT-degrading bacteria, the TBT-resistant isolates were subcultured on a
269 medium containing TBT as the sole carbon source. Growth on this medium demonstrates the ability of the bacteria
270 to use TBT as a sole carbon source.

271 The high variability of the results prevented the detection of a statistical difference between the proportion of
272 isolates able to use TBT as sole carbon source in prepared sediment and untouched sediment. A higher number of
273 bacteria using TBT as the sole carbon source in the prepared sediment would be an expected result as the presence
274 of TBT will have favoured a population of bacteria that was adapted to the presence of such a biocide. TBT

275 degradation and its use as a carbon source is thought to happen through sequential debutylation but the enzymes
276 responsible for this degradation have never been clearly identified (Cruz et al. 2015). In Hassan (2017), the author
277 suggests a role of the protein sugE in TBT degradation as its overexpression enhanced TBT degradation, but the
278 addition of the gene *sugE* alone could not provide the degradation phenotype in *E.coli*. In parallel, siderophores
279 produced by *Pseudomonas chlororaphis* have been shown to be responsible for Tin-C cleavage using triphenyltin
280 (TPT), diphenyltin (DPT) and dibutyltin (DBT) as the substrates and may have the same effect on TBT (Inoue et
281 al. 2003). For siderophores, as well as enzymatic degradation, however, TBT may not be the intended target and
282 its degradation could result from co-metabolism. It is important to emphasise that bacteria, which are not able to
283 use TBT as the sole carbon source could still have the ability to degrade it. Further tests would be necessary to
284 resolve this.

285 iChip reveals members of *Oceanisphaera*, *Bacillus*, *Shewanella* and *Pseudomonas* as TBT-
286 resistant bacteria, and members of *Pseudomonas* as TBT-degrading bacteria

287 The loss of the resistance ability for some of the isolates after a couple rounds of subculturing on TSA + 1mM
288 TBT would suggest a plasmidic location of the resistance genes. These include the only *Bacillus* sp. isolate, the
289 two *Shewanella* sp. isolates, some of the *Pseudomonas* sp. and *Oceanisphaera* sp. isolates.

290 The remaining *Oceanisphaera* sp. isolates were still maintained on TSA + 1mM TBT but could not grow on MSM
291 + 1mM TBT, which means that they were not capable of using TBT as the sole carbon source. At this stage it
292 cannot be determined if they are still capable of TBT degradation by another mechanism. TBT could be degraded
293 by an adverse reaction of enzymes secreted by the bacteria without utilisation of the degradation product.
294 Nevertheless, this is the first time that members of the genus *Oceanisphaera* have been shown to be capable of
295 TBT resistance. *Oceanisphaera* members have been repeatedly isolated from coastal and marine sediment
296 (Romanenko et al. 2003; Shin et al. 2012; Zhou et al. 2015; Cho and Lee 2016), the present study therefore shows
297 their presence in estuarine sediment too.

298 Finally, many of the isolates belonging to *Pseudomonas* sp. were able to use TBT as the sole carbon source. This
299 result is not surprising as *Pseudomonas* members have often been reported as TBT-resistant and as TBT-degraders
300 (Roy et al. 2004; Khanolkar et al. 2015; Yáñez et al. 2015; Ebah et al. 2016). In addition they are also known to
301 degrade a wide range of other sediment contaminants (Wasi et al. 2013).

302 It is interesting to note that all of the isolates coming from the same iChip experiments belong to the same genera,
303 although the small numbers of representatives for some iChips prevent statistically significant conclusions to be
304 made.

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Table 1: Details of the isolates identified by Sanger sequencing of the 16S rRNA genes

This table describes the different isolates and the techniques used for their obtention as well as their growth capacities when the identification was performed and the result of the identification. All of these isolates could grow on TSA + 1 mM at the 1st subculturing.

Accession number of 16S rRNA gene sequence	Isolate	Isolation technique used	Growth on the following medium after 4 th subculturing			Identification
			TSA	TSA + 1 mM TBT	MSM + 1 mM TBT	
OM158192	β2A3	iChip - prepared sediment	+	+	+	<i>Pseudomonas sp.</i>
OM158193	β2B2	iChip - prepared sediment	+	+	+	<i>Pseudomonas sp.</i>
OM158197	β5A5	iChip - prepared sediment	+	+	+	<i>Pseudomonas sp.</i>
OM158198	β5A6	iChip - prepared sediment	+	+	+	<i>Pseudomonas sp.</i>
OM158201	β5C4	iChip - prepared sediment	+	+	+	<i>Pseudomonas sp.</i>
OM158202	β5C5	iChip - prepared sediment	+	+	+	<i>Pseudomonas sp.</i>
OM158200	β5C3	iChip - prepared sediment	+	+	+	<i>Pseudomonas sp.</i>
OM158183	3A1	standard plating	+	+	+	<i>Pseudomonas sp.</i>
OM158184	3A2	standard plating	+	+	+	<i>Pseudomonas sp.</i>
OM158203	I13b	standard plating	+	+	+	<i>Pseudomonas sp.</i>
OM158191	α4D6	iChip - prepared sediment	+	+	-	<i>Oceanisphaera sp.</i>
OM158190	α4A2	iChip - prepared sediment	+	+	-	<i>Oceanisphaera sp.</i>
OM158189	α3D4	iChip - prepared sediment	+	+	-	<i>Oceanisphaera sp.</i>
OM158187	α1C3	iChip - prepared sediment	+	+	-	<i>Oceanisphaera sp.</i>
OM158185	7A	standard plating	+	+	-	<i>Pseudomonas sp.</i>
OM158186	α1B6	iChip - prepared sediment	+	-	-	<i>Oceanisphaera sp.</i>
OM158188	α1D5	iChip - prepared sediment	+	-	-	<i>Oceanisphaera sp.</i>
OM158195	β2C5	iChip - prepared sediment	+	-	-	<i>Pseudomonas sp.</i>
OM158194	β2B6	iChip - prepared sediment	+	-	-	<i>Pseudomonas sp.</i>
OM158196	β2D5	iChip - prepared sediment	+	-	-	<i>Pseudomonas sp.</i>
OM158199	β5B5	iChip - prepared sediment	+	-	-	<i>Pseudomonas sp.</i>
OM158204	γ1D4	iChip - prepared sediment	+	-	-	<i>Bacillus sp.</i>
OM158206	Z3D5b	iChip - untouched sediment	+	-	-	<i>Shewanella sp.</i>
OM158205	Z3D5a	iChip - untouched sediment	+	-	-	<i>Shewanella sp.</i>

329

330 Discussion on the use of iChip for the isolation of uncultured bacteria

331 Owing to its design, iChips are useful tools for the high throughput isolation of bacteria from a wide range of
332 environments. In iChip, bacterial cells can easily be isolated from one another, and their growth is facilitated by
333 the close proximity to the environment. One of the issues stated for the cultivation of unknown bacteria is that the
334 fast-growing species outcompete the slow growing or rare species on the culture plates but in iChips, each bacterial
335 cell occupies one of the many through holes, giving more chance for these species to successfully develop. IChip,
336 however, will not solve every issue. For example, the subculturing is later done in full laboratory conditions, and
337 as this paper shows, not all the bacteria that have been able to grow in iChip are adapted for further growth on
338 synthetic medium. Ideally, a coupling of iChip and the use of alternative media and growth conditions could lead
339 to the best results. The need for key growth factors that are normally not present in the classic incubation media
340 may persist after subculturing out of the iChip, and media supplemented with different types of molecules would
341 still be useful. On the contrary, the nutrient-rich media classically used have sometimes been pointed out as
342 inhibitory to some types of bacteria referred to as ‘oligophilic’ which would only develop on nutrient-poor media
343 (Watve et al. 2000). Lowering the temperature of incubation is also usually suggested and this was done in the
344 present study where all the incubation steps were performed at 20°C.

345 Conclusion

346 iChip was previously shown to successfully increase the success of cultivation of bacteria producing metabolites
347 of medical interest (Piddock 2015) and here we demonstrated its efficiency in increasing the abundance of
348 culturable bacteria of interest in the field of bioremediation and more specifically TBT biodegradation. Further
349 effort is however required in order to maintain most of these isolates in full laboratory conditions after the steps
350 of growth in iChips. After identification of the isolates obtained by iChip, members of the genus *Oceanisphaera*
351 were found associated with TBT resistance for the first time.

352

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481 Author Contribution

482 PA conceived and designed research. PA conducted the experiments. OF contributed new reagents. PA
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485 Data availability

486 The datasets generated during and/or analysed during the current study are available from the
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488 Ethics approval

489 This article does not contain any studies with human participants or animals performed by any of the
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491