



LJMU Research Online

Gould, E, Fraser, H, Parker, T, Nakagawa, S, Griffith, S, Macleod, R and et al,

Same data, different analysts: variation in effect sizes due to analytical decisions in ecology and evolutionary biology

<http://researchonline.ljmu.ac.uk/id/eprint/25570/>

Article

Citation (please note it is advisable to refer to the publisher's version if you intend to cite from this work)

Gould, E, Fraser, H, Parker, T, Nakagawa, S, Griffith, S, Macleod, R and et al, (2025) Same data, different analysts: variation in effect sizes due to analytical decisions in ecology and evolutionary biology. BMC Biology, 23 (1). ISSN 1741-7007

LJMU has developed **LJMU Research Online** for users to access the research output of the University more effectively. Copyright © and Moral Rights for the papers on this site are retained by the individual authors and/or other copyright owners. Users may download and/or print one copy of any article(s) in LJMU Research Online to facilitate their private study or for non-commercial research. You may not engage in further distribution of the material or use it for any profit-making activities or any commercial gain.

The version presented here may differ from the published version or from the version of the record. Please see the repository URL above for details on accessing the published version and note that access may require a subscription.

For more information please contact researchonline@ljmu.ac.uk

<http://researchonline.ljmu.ac.uk/>

REGISTERED REPORT

Open Access



Same data, different analysts: variation in effect sizes due to analytical decisions in ecology and evolutionary biology

Elliot Gould¹ , Hannah S. Fraser² , Timothy H. Parker^{3*} , Shinichi Nakagawa⁴ , Simon C. Griffith⁵ , Peter A. Vesik¹ , Fiona Fidler² , Daniel G. Hamilton⁶ , Robin N. Abbey-Lee⁷ , Jessica K. Abbott⁸, Luis A. Aguirre⁹ , Carles Alcaraz¹⁰ , Irith Aloni¹¹ , Drew Altschul¹² , Kunal Arekar¹³ , Jeff W. Atkins¹⁴ , Joe Atkinson¹⁵ , Christopher M. Baker¹⁶ , Meghan Barrett¹⁷, Kristian Bell¹⁸ , Suleiman Kehinde Bello¹⁹ , Iván Beltrán²⁰ , Bernd J. Berauer²¹ , Michael Grant Bertram²² , Peter D. Billman²³ , Charlie K. Blake²⁴ , Shannon Blake²⁵, Louis Bliard²⁶ , Andrea Bonisoli-Alquati²⁷ , Timothée Bonnet²⁸ , Camille Nina Marion Bordes²⁹ , Aneesh P. H. Bose²² , Thomas Botterill-James³⁰ , Melissa Anna Boyd³¹ , Sarah A. Boyle³² , Tom Bradfer-Lawrence³³ , Jennifer Bradham³⁴, Jack A. Brand²² , Martin I. Brengdah³⁵ , Martin Bulla³⁶ , Luc Bussière³⁷ , Ettore Camerlenghi³⁸ , Sara E. Campbell³⁹ , Leonardo L. F. Campos⁴⁰ , Anthony Caravaggi⁴¹ , Pedro Cardoso⁴² , Charles J. W. Carroll⁴³, Therese A. Catanach⁴⁴ , Xuan Chen⁴⁵ , Heung Ying Janet Chik⁴⁶ , Emily Sarah Choy⁴⁷ , Alec Philip Christie⁴⁸ , Angela Chuang⁴⁹ , Amanda J. Chunco⁵⁰ , Bethany L. Clark⁵¹ , Andrea Contina⁵² , Garth A. Covernton⁵³ , Murray P. Cox⁵⁴ , Kimberly A. Cressman⁵⁵, Marco Crotti⁵⁶ , Connor Davidson Crouch⁵⁷ , Pietro B. D'Amelio⁵⁸ , Alexandra Allison de Sousa⁵⁹ , Timm Fabian Döbert⁶⁰ , Ralph Dobler⁶¹, Adam J. Dobson⁶² , Tim S. Doherty⁶³, Szymon Marian Drobniak⁶⁴ , Alexandra Grace Duffy⁶⁵ , Alison B. Duncan⁶⁶ , Robert P. Dunn⁶⁷ , Jamie Dunning⁶⁸, Trishna Dutta⁶⁹ , Luke Eberhart-Hertel⁷⁰ , Jared Alan Elmore⁷¹ , Mahmoud Medhat Elsherif⁷² , Holly M. English⁷³ , David C. Ensminger⁷⁴ , Ulrich Rainer Ernst⁷⁵ , Stephen M. Ferguson⁷⁶ , Esteban Fernandez-Juricic⁷⁷ , Thalita Ferreira-Arruda⁷⁸ , John Fieberg⁷⁹ , Elizabeth A. Finch⁸⁰ , Evan A. Fiorenza⁸¹ , David N. Fisher⁸² , Amélie Fontaine⁸³, Wolfgang Forstmeier⁷⁰ , Yoan Fourcade⁸⁴ , Graham S. Frank⁸⁵ , Cathryn A. Freund⁸⁶ , Eduardo Fuentes-Lillo⁸⁷ , Sara L. Gandy⁸⁸ , Dustin G. Gannon⁸⁹ , Ana I. García-Cervigón⁹⁰ , Alexis C. Garretson⁹¹ , Xuezhen Ge⁹² , William L. Geary⁹³ , Charly Géron⁹⁴ , Marc Gilles⁹⁵ , Antje Girndt⁹⁶ , Daniel Glikzman⁹⁷, Harrison B. Goldspiel⁹⁸ , Dylan G. E. Gomes⁹⁹ , Megan Kate Good¹⁰⁰ , Sarah C. Goslee¹⁰¹ , J. Stephen Gosnell¹⁰² , Eliza M. Grames¹⁰³ , Paolo Gratton¹⁰⁴ , Nicholas M. Grebe¹⁰⁵ , Skye M. Greenler¹⁰⁶ , Maaïke Griffioen¹⁰⁷ , Daniel M. Griffith¹⁰⁸ , Frances J. Griffith¹⁰⁹ , Jake J. Grossman¹¹⁰ , Ali Güncan¹¹¹ , Stef Haesen¹¹² , James G. Hagan¹¹³ , Heather A. Hager¹¹⁴ , Jonathan Philo Harris¹¹⁵ , Natasha Dean Harrison¹¹⁶ , Sarah Syedia Hasnain¹¹⁷ , Justin Chase Havird¹¹⁸ , Andrew J. Heaton¹¹⁹ , María Laura Herrera-Chaustre¹²⁰ , Tanner J. Howard¹ , Bin-Yan Hsu¹²¹ , Fabiola Iannarilli⁷⁹ , Esperanza C. Iranzo¹²² , Erik N. K. Iverson¹²³

*Correspondence:

Timothy H. Parker
parkerth@whitman.edu

Full list of author information is available at the end of the article



© The Author(s) 2025. **Open Access** This article is licensed under a Creative Commons Attribution-NonCommercial-NoDerivatives 4.0 International License, which permits any non-commercial use, sharing, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons licence, and indicate if you modified the licensed material. You do not have permission under this licence to share adapted material derived from this article or parts of it. The images or other third party material in this article are included in the article's Creative Commons licence, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons licence and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this licence, visit <http://creativecommons.org/licenses/by-nc-nd/4.0/>.

Saheed Olaide Jimoh¹²⁴ , Douglas H. Johnson⁷⁹ , Martin Johnsson¹²⁵ , Jesse Jorna¹²⁶, Tommaso Jucker¹²⁷ , Martin Jung¹²⁸ , Ineta Kačergytė¹²⁹ , Oliver Kaltz¹³⁰, Alison Ke¹³¹ , Clint D. Kelly¹³² , Katharine Keogan¹³³ , Friedrich Wolfgang Keppeler¹³⁴ , Alexander K. Killion¹³⁵ , Dongmin Kim¹³⁶ , David P. Kochan¹³⁷ , Peter Korsten¹³⁸ , Shan Kothari¹³⁹ , Jonas Kuppler¹⁴⁰ , Jillian M. Kusch¹⁴¹ , Malgorzata Lagisz¹⁴² , Kristen Marianne Lalla⁸³ , Daniel J. Larkin⁷⁹ , Courtney L. Larson¹⁴³ , Katherine S. Lauck¹³¹ , M. Elise Lauterbur¹⁴⁴ , Alan Law¹⁴⁵ , Don-Jean Léandri-Breton⁸³ , Jonas J. Lembrechts¹⁴⁶ , Kiara L'Herpinier⁵ , Eva J. P. Lievens¹⁴⁷ , Daniela Oliveira de Lima¹⁴⁸ , Shane Lindsay¹⁴⁹, Martin Luquet¹⁵⁰ , Ross MacLeod¹⁵¹ , Kirsty H. Macphie¹⁵² , Kit Magellan¹⁵³, Magdalena M. Mair¹⁵⁴ , Lisa E. Malm¹⁵⁵ , Stefano Mammola¹⁵⁶ , Caitlin P. Mandeville¹⁵⁷ , Michael Manhart¹⁵⁸ , Laura Milena Manrique-Garzon¹⁵⁹ , Elina Mäntylä¹²¹ , Philippe Marchand¹⁶⁰ , Benjamin Michael Marshall¹⁴⁵ , Charles A. Martin¹⁶¹ , Dominic Andreas Martin¹⁶² , Jake Mitchell Martin²² , April Robin Martinig¹⁶³ , Erin S. McCallum²² , Mark McCauley¹⁶⁴ , Sabrina M. McNew¹⁴⁴ , Scott J. Meiners¹⁶⁵ , Thomas Merkling¹⁶⁶ , Marcus Michelangeli²² , Maria Moiron¹⁶⁷ , Bruno Moreira¹⁶⁸ , Jennifer Mortensen¹⁶⁹, Benjamin Mos¹⁷⁰ , Taofeek Olatunbosun Muraina¹⁷¹ , Penelope Wrenn Murphy¹⁷² , Luca Nelli⁵⁶ , Petri Niemelä¹⁷³, Josh Nightingale¹⁷⁴ , Gustav Nilsson¹⁷⁵ , Sergio Nolzco³⁸ , Sabine S. Nooten¹⁷⁶ , Jessie Lanterman Novotny¹⁷⁷ , Agnes Birgitta Olin¹⁷⁸ , Chris L. Organ¹⁷⁹, Kate L. Ostevik¹⁸⁰ , Facundo Xavier Palacio¹⁸¹ , Matthieu Paquet¹²⁹ , Darren James Parker¹⁸² , David J. Pascall¹⁸³ , Valerie J. Pasquarella¹⁸⁴ , John Harold Paterson¹¹³, Ana Payo-Payo¹⁸⁵ , Karen Marie Pedersen¹⁸⁶ , Grégoire Perez¹⁸⁷ , Kayla I. Perry¹⁸⁸ , Patrice Pottier¹⁴² , Michael J. Proulx¹⁸⁹ , Raphaël Proulx¹⁹⁰ , Jessica L. Pruett¹⁹¹, Veronarindra Ramananjato¹⁹² , Finaritra Tolotra Randimbiarison¹⁹³, Onja H. Razafindratsima¹⁹⁴ , Diana J. Rennison¹⁹⁵ , Federico Riva¹⁹⁶ , Sepand Riyahi¹⁹⁷ , Michael James Roast¹⁹⁸ , Felipe Pereira Rocha¹⁹⁹ , Dominique G. Roche²⁰⁰ , Cristian Román-Palacios²⁰¹ , Michael S. Rosenberg²⁰² , Jessica Ross²⁰³ , Freya E. Rowland²⁰⁴ , Deusdedith Rugemalila²⁰⁵ , Avery L. Russell²⁰⁶ , Suvi Ruuskanen²⁰⁷ , Patrick Saccone²⁰⁸ , Asaf Sadeh²⁰⁹ , Stephen M. Salazar²¹⁰ , Kris Sales²¹¹ , Pablo Salmón²¹² , Alfredo Sánchez-Tójar²¹³ , Leticia Pereira Santos²¹⁴, Francesca Santostefano²¹⁵ , Hayden T. Schilling²¹⁶ , Marcus Schmidt²¹⁷ , Tim Schmoll²¹³ , Adam C. Schneider²¹⁸ , Allie E. Schrock²¹⁹ , Julia Schroeder⁶⁸ , Nicolas Schtickzelle²²⁰ , Nick L. Schultz²²¹ , Drew A. Scott²²² , Michael Peter Scroggie²²³ , Julie Teresa Shapiro²²⁴ , Nitika Sharma²²⁵ , Caroline L. Shearer²¹⁹ , Diego Simón²²⁶ , Michael I. Sitvarin²²⁷ , Fabrício Luiz Skupien²²⁸ , Heather Lea Slinn²²⁹, Grania Polly Smith²³⁰, Jeremy A. Smith²³¹ , Rahel Sollmann⁸⁹ , Kaitlin Stack Whitney²³² , Shannon Michael Still²³³ , Erica F. Stuber²³⁴ , Guy F. Sutton²³⁵ , Ben Swallow²³⁶ , Conor Claverie Taff²³⁷ , Elina Takola²³⁸ , Andrew J. Tanentzap²³⁹ , Rocío Tarjuelo²⁴⁰ , Richard J. Telford²⁴¹ , Christopher J. Thawley²⁴² , Hugo Thierry²⁴³, Jacqueline Thomson⁹², Svenja Tidau²⁴⁴ , Emily M. Tompkins²⁴⁵ , Claire Marie Tortorelli²⁴⁶ , Andrew Trlica²⁴⁷ , Biz R. Turnell²⁴⁸ , Lara Urban²⁴⁹ , Stijn Van de Vondel¹⁴⁶ , Jessica Eva Megan van der Wal²⁵⁰ , Jens Van Eeckhoven²⁵¹ , Francis van Oordt⁸³ , K. Michelle Vanderwel²⁵², Mark C. Vanderwel²⁵³, Karen J. Vanderwolf²⁵⁴ , Juliana Vélez⁷⁹ , Diana Carolina Vergara-Florez²⁵⁵ , Brian C. Verrelli¹⁴⁶ , Marcus Vinícius Vieira²⁵⁶ , Nora Villamil²⁵⁷ , Valerio Vitali²⁵⁸ , Julien Vollering²⁵⁹ , Jeffrey Walker²⁶⁰ , Xanthe J. Walker²⁶¹ , Jonathan A. Walter²⁶² , Pawel Waryszak²⁶³ , Ryan J. Weaver²⁶⁴ , Ronja E. M. Wedegärtner²⁶⁵ , Daniel L. Weller²⁶⁶ , Shannon Whelan⁸³ , Rachel Louise White²⁶⁷ , David William Wolfson⁷⁹ , Andrew Wood²⁶⁸ , Scott W. Yanco²⁶⁹ , Jian D. L. Yen²²³ , Casey Youngflesh²⁷⁰ , Giacomo Zilio²⁷¹ , Cédric Zimmer²⁷² , Gregory Mark Zimmerman²⁷³ and Rachel A. Zitomer⁸⁵ 

Abstract

Although variation in effect sizes and predicted values among studies of similar phenomena is inevitable, such variation far exceeds what might be produced by sampling error alone. One possible explanation for variation among results is differences among researchers in the decisions they make regarding statistical analyses. A growing array of studies has explored this analytical variability in different fields and has found substantial variability among results despite analysts having the same data and research question. Many of these studies have been in the social sciences, but one small “many analyst” study found similar variability in ecology. We expanded the scope of this prior work by implementing a large-scale empirical exploration of the variation in effect sizes and model predictions generated by the analytical decisions of different researchers in ecology and evolutionary biology. We used two unpublished datasets, one from evolutionary ecology (blue tit, *Cyanistes caeruleus*, to compare sibling number and nestling growth) and one from conservation ecology (*Eucalyptus*, to compare grass cover and tree seedling recruitment). The project leaders recruited 174 analyst teams, comprising 246 analysts, to investigate the answers to prespecified research questions. Analyses conducted by these teams yielded 141 usable effects (compatible with our meta-analyses and with all necessary information provided) for the blue tit dataset, and 85 usable effects for the *Eucalyptus* dataset. We found substantial heterogeneity among results for both datasets, although the patterns of variation differed between them. For the blue tit analyses, the average effect was convincingly negative, with less growth for nestlings living with more siblings, but there was near continuous variation in effect size from large negative effects to effects near zero, and even effects crossing the traditional threshold of statistical significance in the opposite direction. In contrast, the average relationship between grass cover and *Eucalyptus* seedling number was only slightly negative and not convincingly different from zero, and most effects ranged from weakly negative to weakly positive, with about a third of effects crossing the traditional threshold of significance in one direction or the other. However, there were also several striking outliers in the *Eucalyptus* dataset, with effects far from zero. For both datasets, we found substantial variation in the variable selection and random effects structures among analyses, as well as in the ratings of the analytical methods by peer reviewers, but we found no strong relationship between any of these and deviation from the meta-analytic mean. In other words, analyses with results that were far from the mean were no more or less likely to have dissimilar variable sets, use random effects in their models, or receive poor peer reviews than those analyses that found results that were close to the mean. The existence of substantial variability among analysis outcomes raises important questions about how ecologists and evolutionary biologists should interpret published results, and how they should conduct analyses in the future.

Keywords Analytical heterogeneity, Metascience, Many-analyst, Replication crisis, Reproducibility

Introduction

One value of science derives from its production of replicable, and thus reliable, results. When we repeat a study using the original methods, we should be able to expect a similar result. However, perfect replicability is not a reasonable goal. Effect sizes will vary, and even reverse in sign, by chance alone [37]. Observed patterns can differ for other reasons as well. It could be that we do not sufficiently understand the conditions that led to the original result so when we seek to replicate it, the conditions differ due to some “hidden moderator”. This hidden moderator hypothesis is described by meta-analysts in ecology and evolutionary biology as “true biological heterogeneity” [93]. This idea of true heterogeneity is popular in ecology and evolutionary biology, and there are good reasons to expect it in the complex systems in which we work [94]. However, despite similar expectations in psychology, recent evidence in that discipline contradicts the hypothesis that moderators are common obstacles to replicability, as variability in results in a large “many labs” collaboration was mostly unrelated to commonly

hypothesized moderators such as the conditions under which the studies were administered [50]. Another possible explanation for variation in effect sizes is that researchers often present biased samples of results, thus reducing the likelihood that later studies will produce similar effect sizes [33, 34, 80, 82, 83]. It also may be that although researchers did successfully replicate the conditions, the experiment, and measured variables, analytical decisions differed sufficiently among studies to create divergent results [96, 99].

Analytical decisions vary among studies because researchers have many options. Researchers need to decide how to exclude possibly anomalous or unreliable data, how to construct variables, which variables to include in their models, and which statistical methods to use. Depending on the dataset, this short list of choices could encompass thousands or millions of possible alternative specifications [100]. However, researchers making these decisions presumably do so with the goal of doing the best possible analysis, or at least the best analysis within their current skill set. Thus, it seems

likely that some specification options are more probable than others, possibly because they have previously been shown (or claimed) to be better, or because they are more well known. Of course, some of these different analyses (maybe many of them) may be equally valid alternatives. Regardless, on probably any topic in ecology and evolutionary biology, we can encounter differences in choices of data analysis. The extent of these differences in analyses and the degree to which these differences influence the outcomes of analyses and therefore studies' conclusions are important empirical questions. These questions are especially important given that many papers draw conclusions after applying a single method, or even a single statistical model, to analyze a dataset.

The possibility that different analytical choices could lead to different outcomes has long been recognized [36], and various efforts to address this possibility have been pursued in the literature. For instance, one common method in ecology and evolutionary biology involves creating a set of candidate models, each consisting of a different (though often similar) set of predictor variables, and then, for the predictor variable of interest, averaging the slope across all models (i.e. model averaging) [20, 40]. This method reduces the chance that a conclusion is contingent upon a single model specification, though use and interpretation of this method is not without challenges [40]. Further, the models compared to each other typically differ only in the inclusion or exclusion of certain predictor variables and not in other important ways, such as methods of parameter estimation. More explicit examination of outcomes of differences in model structure, model type, data exclusion, or other analytical choices can be implemented through sensitivity analyses (e.g., [78]). Sensitivity analyses, however, are typically rather narrow in scope and are designed to assess the sensitivity of analytical outcomes to a particular analytical choice rather than to a large universe of choices. Recently, however, analysts in the social sciences have proposed extremely thorough sensitivity analysis, including 'multiverse analysis' [104] and the 'specification curve' [99], as a means of increasing the reliability of results. With these methods, researchers identify relevant decision points encountered during analysis and conduct the analysis many times to incorporate many plausible decisions made at each of these points. The study's conclusions are then based on a broad set of the possible analyses and so allow the analyst to distinguish between robust conclusions and those that are highly contingent on particular model specifications. These are useful outcomes, but specifying a universe of possible modelling decisions is not a trivial undertaking. Further, the analyst's knowledge and biases will influence decisions about the boundaries of that universe, and so there will always be room

for disagreement among analysts about what to include. Including more specifications is not necessarily better. Some analytical decisions are better justified than others, and including biologically implausible specifications may undermine this process. Regardless, these powerful methods have yet to be adopted, and even the more limited forms of sensitivity analyses are not particularly widespread. Most studies publish a small set of analyses and so the existing literature does not provide much insight into the degree to which published results are contingent on analytical decisions.

Despite the potential major impacts of analytical decisions on variance in results, the outcomes of different individuals' data analysis choices have only recently begun to receive much empirical attention. The only formal exploration of this that we were aware of when we submitted our Stage 1 manuscript were (1) an analysis in social science that asked whether male professional football (soccer) players with darker skin tone were more likely to be issued red cards (ejection from the game for rule violation) than players with lighter skin tone [96] and (2) an analysis in neuroimaging which evaluated nine separate hypotheses involving the neurological responses detected with fMRI in 108 participants divided between two treatments in a decision making task [15]. Several others have been published since [16, 23, 44, 92], and we recently learned of an earlier small study in ecology [103]. In the red card study, 29 teams designed and implemented analyses of a dataset provided by the study coordinators [96]. Analyses were peer reviewed (results blind) by at least two other participating analysts, a level of scrutiny consistent with standard pre-publication peer review. Among the final 29 analyses, odds ratios varied from 0.89 to 2.93, meaning point estimates varied from having players with lighter skin tones receive more red cards (odds ratio < 1) to a strong effect of players with darker skin tones receiving more red cards (odds ratio > 1). Twenty of the 29 teams found a statistically significant effect in the predicted direction of players with darker skin tones being issued more red cards. This degree of variation in peer-reviewed analyses from identical data is striking, but the generality of this finding has only just begun to be formally investigated [16, 23, 44, 92].

In the neuroimaging study, 70 teams evaluated each of the nine different hypotheses with the available fMRI data [15]. These 70 teams followed a divergent set of workflows that produced a wide range of results. The rate of reporting of statistically significant support for the nine hypotheses ranged from 21 to 84%, and for each hypothesis on average, 20% of research teams observed effects that differed substantially from the majority of other teams. Some of the variability in results among studies could be explained by analytical decisions such

as choice of software package, smoothing function, and parametric versus non-parametric corrections for multiple comparisons. However, substantial variability among analyses remained unexplained, and presumably emerged from the many different decisions each analyst made in their long workflows. Such variability in results among analyses from this dataset and from the very different red-card dataset suggests that sensitivity of analytical outcome to analytical choices may characterize many distinct fields, as several more recent many-analyst studies also suggest [16, 44, 92].

To further develop the empirical understanding of the effects of analytical decisions on study outcomes, we chose to estimate the extent to which researchers' data analysis choices drive differences in effect sizes, model predictions, and qualitative conclusions in ecology and evolutionary biology. This is an important extension of the meta-research agenda of evaluating factors influencing replicability in ecology, evolutionary biology, and beyond [31]. To examine the effects of analytical decisions, we used two different datasets and recruited researchers to analyze one or the other of these datasets to answer a question we defined. The first question was "To what extent is the growth of nestling blue tits (*Cyanistes caeruleus*) influenced by competition with siblings?" To answer this question, we provided a dataset that includes brood size manipulations from 332 broods conducted over 3 years at Wytham Wood, UK. The second question was "How does grass cover influence *Eucalyptus* spp. seedling recruitment?" For this question, analysts used a dataset that includes, among other variables, number of seedlings in different size classes, percentage cover of different life forms, tree canopy cover, and distance from canopy edge from 351 quadrats spread among 18 sites in Victoria, Australia.

We explored the impacts of data analysts' choices with descriptive statistics and with a series of tests to attempt to explain the variation among effect sizes and predicted values of the dependent variable produced by the different analysis teams for both datasets separately. To describe the variability, we present forest plots of the standardized effect sizes and predicted values produced by each of the analysis teams, estimate heterogeneity (both absolute, τ^2 , and proportional, I^2) in effect size and predicted values among the results produced by these different teams, and calculate a similarity index that quantifies variability among the predictor variables selected for the different statistical models constructed by the different analysis teams. These descriptive statistics provide the first estimates of the extent to which explanatory statistical models and their outcomes in ecology and evolutionary biology vary based on the decisions of different data analysts. We then quantified

the degree to which the variability in effect size and predicted values could be explained by (1) variation in the quality of analyses as rated by peer reviewers and (2) the similarity of the choices of predictor variables between individual analyses.

Methods

This project involved a series of steps (1–6) that began with identifying datasets for analyses and continued through recruiting independent groups of scientists to analyze the data, allowing the scientists to analyze the data as they saw fit, generating peer review ratings of the analyses (based on methods, not results), evaluating the variation in effects among the different analyses, and producing the final manuscript.

Step 1: Select datasets

We used two previously unpublished datasets, one from evolutionary ecology and the other from ecology and conservation.

Evolutionary ecology

Our evolutionary ecology dataset is relevant to a sub-discipline of life-history research which focuses on identifying costs and trade-offs associated with different phenotypic conditions. These data were derived from a brood-size manipulation experiment imposed on wild birds nesting in boxes provided by researchers in an intensively studied population. Understanding how the growth of nestlings is influenced by the numbers of siblings in the nest can give researchers insights into factors such as the evolution of clutch size, determination of provisioning rates by parents, and optimal levels of sibling competition [25, 77, 89, 110, 112]. Data analysts were provided this dataset and instructed to answer the following question: "To what extent is the growth of nestling blue tits (*Cyanistes caeruleus*) influenced by competition with siblings?"

Researchers conducted brood size manipulations and population monitoring of blue tits at Wytham Wood, a 380-ha woodland in Oxfordshire, UK (1° 20' W, 51° 47' N). Researchers regularly checked approximately 1100 artificial nest boxes at the site and monitored the 330 to 450 blue tit pairs occupying those boxes in 2001–2003 during the experiment. Nearly all birds made only one breeding attempt during the April to June study period in a given year. At each blue tit nest, researchers recorded the date the first egg appeared, clutch size, and hatching date. For all chicks alive at age 14 days, researchers measured mass and tarsus length and fitted a uniquely numbered, British Trust for Ornithology (BTO) aluminum leg ring.

Researchers attempted to capture all adults at their nests between day 6 and day 14 of the chick-rearing period. For these captured adults, researchers measured mass, tarsus length, and wing length and fitted a uniquely numbered BTO leg ring. During the 2001–2003 breeding seasons, researchers manipulated brood sizes using cross fostering. They matched broods for hatching date and brood size and moved chicks between these paired nests 1 or 2 days after hatching. They sought to either enlarge or reduce all manipulated broods by approximately one fourth. To control for effects of being moved, each reduced brood had a portion of its brood replaced by chicks from the paired increased brood, and vice versa. Net manipulations varied from plus or minus four chicks in broods of 12 to 16 to plus or minus one chick in broods of 4 or 5. Researchers left approximately one third of all broods unmanipulated. These unmanipulated broods were not selected systematically to match manipulated broods in clutch size or laying date. We have mass and tarsus length data from 3720 individual chicks divided among 167 experimentally enlarged broods, 165 experimentally reduced broods, and 120 unmanipulated broods. The full list of variables included in the dataset is publicly available (<https://osf.io/hdv8m>), along with the data (<https://osf.io/qjzby>).

Ecology and conservation

Additional Explanation:

Shortly after beginning to recruit analysts, several analysts noted a small set of related errors in the blue tit dataset. We corrected the errors, replaced the dataset on our OSF site, and emailed the analysts on 19 April 2020 to instruct them to use the revised data. The email to analysts is available here (<https://osf.io/4h53z>). The errors are explained in that email.

Our ecology and conservation dataset is relevant to a sub-discipline of conservation research which focuses on investigating how best to revegetate private land in agricultural landscapes. These data were collected on private land under the Bush Returns program, an incentive system where participants entered into a contract with the Goulburn Broken Catchment Management Authority and received annual payments if they executed predetermined restoration activities. This particular dataset is based on a passive regeneration initiative, where livestock grazing was removed from the property in the hopes that the *Eucalyptus* spp. overstorey would regenerate without active (and expensive) planting. Analyses of some related data have been published [67, 113] but those analyses do not address the question analysts answered in our study. Data analysts were provided this dataset and instructed

to answer the following question: “How does grass cover influence *Eucalyptus* spp. seedling recruitment?”

Researchers conducted three rounds of surveys at 18 sites across the Goulburn Broken catchment in northern Victoria, Australia, in winter and spring 2006 and autumn 2007. In each survey period, a different set of 15 × 15 m quadrats were randomly allocated across each site within 60 m of existing tree canopies. The number of quadrats at each site depended on the size of the site, ranging from four at smaller sites to 11 at larger sites. The total number of quadrats surveyed across all sites and seasons was 351. The number of *Eucalyptus* spp. seedlings was recorded in each quadrat along with information on the GPS location, aspect, tree canopy cover, distance to tree canopy, and position in the landscape. Ground layer plant species composition was recorded in three 0.5 × 0.5 m sub-quadrats within each quadrat. Subjective cover estimates of each species as well as bare ground, litter, rock and moss/lichen/soil crusts were recorded. Subsequently, this was augmented with information about the precipitation and solar radiation at each GPS location. The full list of variables included in the dataset is publicly available (<https://osf.io/r5gbn>), along with the data (<https://osf.io/qz5cu>).

Step 2: Recruitment and initial survey of analysts

The lead team (TP, HE, SN, EG, SG, PV, DH, FF) created a publicly available document providing a general description of the project (<https://osf.io/mn5aj/>). The project was advertised at conferences, via Twitter, using mailing lists for ecological societies (including *Ecology*, *Evoldir*, and lists for *the Environmental Decisions Group*, and *Transparency in Ecology and Evolution*), and via word of mouth. The target population was active ecology, conservation, or evolutionary biology researchers with a graduate degree (or currently studying for a graduate degree) in a relevant discipline. Researchers could choose to work independently or in a small team. For the sake of simplicity, we refer to these as “analysis teams” though some comprised one individual. We aimed for a minimum of 12 analysis teams independently evaluating each dataset (see sample size justification below). We simultaneously recruited volunteers to peer review the analyses conducted by the other volunteers through the same channels. Our goal was to recruit a similar number of peer reviewers and analysts, and to ask each peer reviewer to review a minimum of four analyses. If we were unable to recruit at least half the number of reviewers as analysis teams, we planned to ask analysts to serve also as reviewers (after they had completed their analyses), but this was unnecessary. Therefore, no data analysts peer reviewed analyses of the dataset they had analyzed. All analysts and reviewers were offered the

opportunity to share co-authorship on this manuscript and we planned to invite them to participate in the collaborative process of producing the final manuscript. All analysts signed [digitally] a consent (ethics) document (<https://osf.io/xyp68/>) approved by the Whitman College Institutional Review Board prior to being allowed to participate.

Preregistration Deviation:

Due to the large number of recruited analysts and reviewers and the anticipated challenges of receiving and integrating feedback from so many authors, we limited analyst and reviewer participation in the production of the final manuscript to an invitation to call attention to serious problems with the manuscript draft.

We identified our minimum number of analysts per dataset by considering the number of effects needed in a meta-analysis to generate an estimate of heterogeneity (τ^2) with a 95% confidence interval that does not encompass zero. This minimum sample size is invariant regardless of τ^2 . This is because the same t -statistic value will be obtained by the same sample size regardless of variance (τ^2). We see this by first examining the formula for the standard error, SE for variance, (τ^2) or ($SE\tau^2$) assuming normality in an underlying distribution of effect sizes [51]:

$$SE(\tau^2) = \sqrt{\frac{2\tau^4}{n-1}}$$

and then rearranging the above formula to show how the t -statistic is independent of τ^2 , as seen below.

$$t = \frac{\tau^2}{SE(\tau^2)} = \sqrt{\frac{n-1}{2}}$$

We then find a minimum $n = 12$ according to this formula.

Step 3: Primary data analyses

Analysis teams registered and answered a demographic and expertise survey (<https://osf.io/seqzy/>). We then provided them with the dataset of their choice and requested that they answer a specific research question. For the evolutionary ecology dataset that question was “To what extent is the growth of nestling blue tits (*Cyanistes caeruleus*) influenced by competition with siblings?” and for the conservation ecology dataset it was “How does grass cover influence *Eucalyptus* spp. seedling recruitment?” Once their analysis was complete, they answered a structured survey (<https://osf.io/neyc7/>), providing analysis technique, explanations of their analytical choices, quantitative results, and a statement describing their conclusions. They also were asked to upload their analysis files (including the dataset as they formatted it for analysis

and their analysis code [if applicable]) and a detailed journal-ready statistical methods section.

Additional Information:

As is common in many studies in ecology and evolutionary biology, the datasets we provided contained many variables, and the research questions we provided could be addressed by our datasets in many different ways. For instance, volunteer analysts had to choose the dependent (response) variable and the independent variable, and make numerous other decisions about which variables and data to use and how to structure their model.

Preregistration Deviation:

We originally planned to have analysts complete a single survey (<https://osf.io/neyc7/>), but after we evaluated the results of that survey, we realized we would need a second survey (<https://osf.io/8w3v5/>) to adequately collect the information we needed to evaluate heterogeneity of results (step 5). We provided a set of detailed instructions with the follow-up survey, and these instructions are publicly available and can be found within the following files (blue tit: <https://osf.io/kr2g9>, *Eucalyptus*: <https://osf.io/dfvym>).

Step 4: Peer reviews of analyses

At minimum, each analysis was evaluated by four different reviewers, and each volunteer peer reviewer was randomly assigned methods sections from at least four analyst teams (the exact number varied). Each peer reviewer registered and answered a demographic and expertise survey identical to that asked of the analysts, except we did not ask about “team name” since reviewers did not work in teams. Reviewers evaluated the methods of each of their assigned analyses one at a time in a sequence determined by the project leaders. We systematically assigned the sequence so that, if possible, each analysis was allocated to each position in the sequence for at least one reviewer. For instance, if each reviewer were assigned four analyses to review, then each analysis would be the first analysis assigned to at least one reviewer, the second analysis assigned to another reviewer, the third analysis assigned to yet another reviewer, and the fourth analysis assigned to a fourth reviewer. Balancing the order in which reviewers saw the analyses controls for order effects, e.g. a reviewer might be less critical of the first methods section they read than the last.

The process for a single reviewer was as follows. First, the reviewer received a description of the methods of a single analysis. This included the narrative methods section, the analysis team’s answers to our survey questions regarding their methods, including analysis code, and the dataset. The reviewer was then asked, in an online survey (<https://osf.io/4t36u/>), to rate that analysis on a scale of 0–100 based on this prompt: “Rate the overall appropriateness of this analysis to answer

the research question (*one of the two research questions inserted here*) with the available data". To help you calibrate your rating, please consider the following guidelines:

- 100. A perfect analysis with no conceivable improvements from the reviewer
- 75. An imperfect analysis but the needed changes are unlikely to dramatically alter outcomes
- 50. A flawed analysis likely to produce either an unreliable estimate of the relationship or an over-precise estimate of uncertainty
- 25. A flawed analysis likely to produce an unreliable estimate of the relationship and an over-precise estimate of uncertainty
- 0. A dangerously misleading analysis, certain to produce both an estimate that is wrong and a substantially over-precise estimate of uncertainty that places undue confidence in the incorrect estimate.

*Please note that these values are meant to calibrate your ratings. We welcome ratings of any number between 0 and 100.

After providing this rating, the reviewer was presented with this prompt, in multiple-choice format: "Would the analytical methods presented produce an analysis that is (a) publishable as is, (b) publishable with minor revision, (c) publishable with major revision, (d) deeply flawed and unpublishable?" The reviewer was then provided with a series of text boxes and the following prompts: "Please explain your ratings of this analysis. Please evaluate the choice of statistical analysis type. Please evaluate the process of choosing variables for and structuring the statistical model. Please evaluate the suitability of the variables included in (or excluded from) the statistical model. Please evaluate the suitability of the structure of the statistical model. Please evaluate choices to exclude or not exclude subsets of the data. Please evaluate any choices to transform data (or, if there were no transformations, but you think there should have been, please discuss that choice)." After submitting this review, a methods section from a second analysis was then made available to the reviewer. This same sequence was followed until all analyses allocated to a given reviewer were provided and reviewed. After providing the final review, the reviewer was simultaneously provided with all four (or more) methods sections the reviewer had just completed reviewing, the option to revise their original ratings, and a text box to provide an explanation. The invitation to revise the original ratings was as

follows: "If, now that you have seen all the analyses you are reviewing, you wish to revise your ratings of any of these analyses, you may do so now." The text box was prefaced with this prompt: "Please explain your choice to revise (or not to revise) your ratings."

Additional Information: unregistered analysis

To determine how consistent peer reviewers were in their ratings, we assessed inter-rater reliability among reviewers for both the categorical and quantitative ratings combining blue tit and *Eucalyptus* data using Krippendorff's alpha for ordinal and continuous data respectively. This provides a value that is between -1 (total disagreement between reviewers) and 1 (total agreement between reviewers).

Step 5: Evaluate variation

Additional Information: analysis schematic

The lead team conducted a range of preregistered and exploratory analyses to understand variation between analyses and their results. Figure 1 is intended to clarify the analyses described below.

The lead team conducted the analyses outlined in this section. We described the variation in model specification in several ways. We calculated summary statistics describing variation among analyses, including mean, SD, and range of number of variables per model included as fixed effects, the number of interaction terms, the number of random effects, and the mean, SD, and range of sample sizes. We also present the number of analyses in which each variable was included. We summarized the variability in standardized effect sizes and predicted values of dependent variables among the individual analyses using standard random effects meta-analytic techniques. First, we derived standardized effect sizes from each individual analysis. We did this for all linear models or generalized linear models by converting the t value and the degree of freedom (df) associated with regression coefficients (e.g. the effect of the number of siblings [predictor] on growth [response] or the effect of grass cover [predictor] on seedling recruitment [response]) to the correlation coefficient, r , using the following:

$$r = \frac{t^2}{(t^2 + df)}$$

This formula can only be applied if t and df values originate from linear or generalized linear models [72]. If, instead, linear mixed-effects models (LMMs) or generalized linear mixed-effects models (GLMMs) were used by a given analysis, the exact df cannot be estimated. However, adjusted df can be estimated,

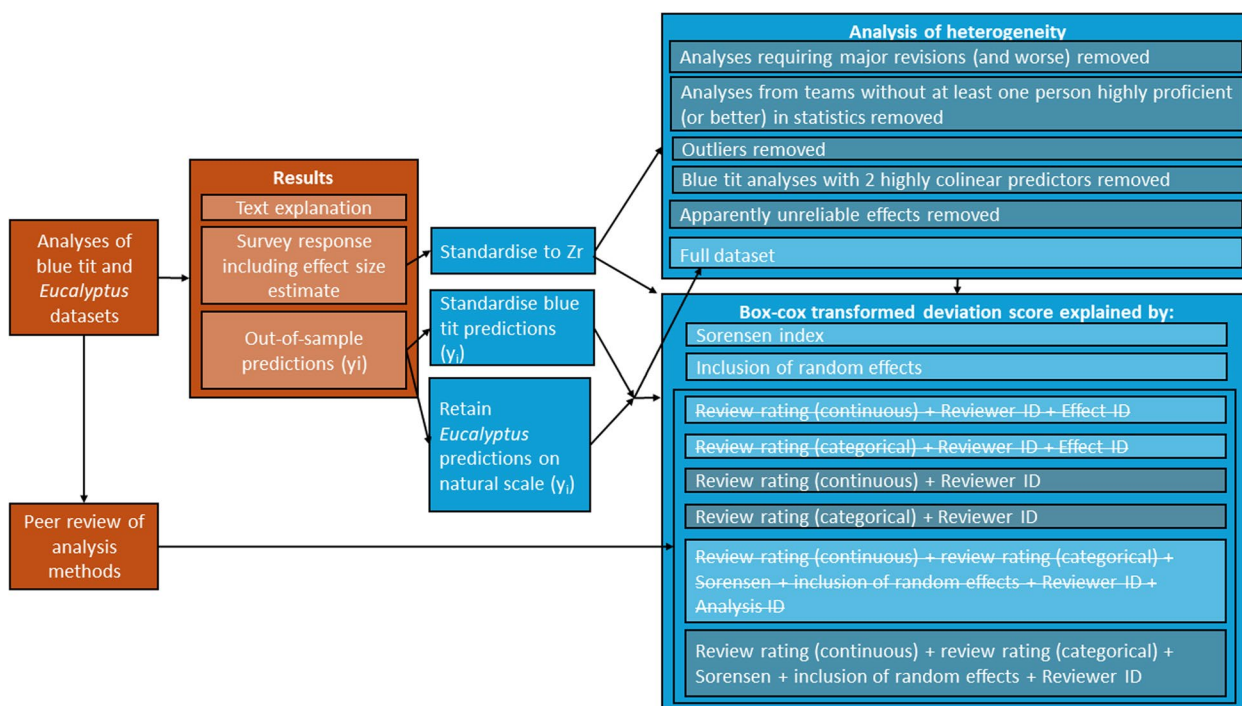


Fig. 1 Schematic of research process showing recruited analyst and reviewer contributions in orange and core team contributions in blue. Items that are crossed out were preregistered but could not be conducted. Items with a greyed background were added as exploratory analyses after preregistration

for example, using the Satterthwaite approximation of df , df_s , [note that SAS uses this approximation to obtain df for LMMs and GLMMs; [63]. For analyses using either LMMs or GLMMs that do not produce df_s we planned to obtain df_s by rerunning the same (G)LMMs using the `lmer()` or `glmer()` function in the `lmerTest` package in R [56, 87].

Preregistration Deviation:

Rather than re-run these analyses ourselves, we sent a follow-up survey (referenced above under “Primary data analyses”) to analysts and asked them to follow our instructions for producing this information. The instructions are publicly available and can be found within the following files (blue tit: <https://osf.io/kr2g9>, *Eucalyptus*: <https://osf.io/dfvym>).

We then used the t values and df_s from the models to obtain r as per the formula above. All r and accompanying df (or df_s) were converted to Fisher’s Z_r .

$$Z_r = \frac{1}{2} \ln \left(\frac{1+r}{1-r} \right)$$

and its sampling variance; $1/(n-3)$ where $n=df+1$. Any analyses from which we could not derive a signed Z_r ,

for instance one with a quadratic function in which the slope changed sign, were considered unusable for analyses of Z_r . We expected such analyses would be rare. In fact, most submitted analyses excluded from our meta-analysis of Z_r , were excluded because of a lack of sufficient information provided by the analyst team rather than due to the use of effects that could not be converted to Z_r . Regardless, as we describe below, we generated a second set of standardized effects (predicted values) that could (in principle) be derived from any explanatory model produced by these data.

Besides Z_r , which describes the strength of a relationship based on the amount of variation in a dependent variable explained by variation in an independent variable, we also examined differences in the shape of the relationship between the independent and dependent variables. To accomplish this, we derived a point estimate (out-of-sample predicted value) for the dependent variable of interest for each of three values of our primary independent variable. We originally described these three values as associated with the 25th percentile, median, and 75th percentile of the independent variable and any covariates.

Preregistration Deviation:

The original description of the out-of-sample specifications did not account for the facts that (a) some variables are not distributed in a way that allowed division in percentiles and that (b) variables could be either positively or negatively correlated with the dependent variable. We provide a more thorough description here:

We derived three point-estimates (out-of-sample predicted values) for the dependent variable of interest; one for each of three values of our primary independent variable that we specified. We also specified values for all other variables that could have been included as independent variables in analysts' models so that we could derive the predicted values from a fully specified version of any model produced by analysts. For all potential independent variables, we selected three values or categories. Of the three we selected, one was associated with small, one with intermediate, and one with large values of one typical dependent variable (day 14 chick weight for the blue tit data and total number of seedlings for the *Eucalyptus* data; analysts could select other variables as their dependent variable, but the others typically correlated with the two identified here). For continuous variables, this means we identified the 25th percentile, median, and 75th percentile and, if the slope of the linear relationship between this variable and the typical dependent variable was positive, we left the quartiles ordered as is. If, instead, the slope was negative, we reversed the order of the independent variable quartiles so that the 'lower' quartile value was the one associated with the lower value for the dependent variable. In the case of categorical variables, we identified categories associated with the 25th percentile, median, and 75th percentile values of the typical dependent variable after averaging the values for each category. However, for some continuous and categorical predictors, we also made selections based on the principle of internal consistency between certain related variables, and we fixed a few categorical variables as identical across all three levels where doing so would simplify the modelling process (specification tables available: blue tit: <https://osf.io/86akx>; *Eucalyptus*: <https://osf.io/jh7g5>).

We used the 25th and 75th percentiles rather than minimum and maximum values to reduce the chance of occupying unrealistic parameter space. We planned to derive these predicted values from the model information provided by the individual analysts. All values (predictions) were first transformed to the original scale along with their standard errors (SE); we used the delta method [111] for the transformation of SE. We used the square of the SE associated with predicted values as the sampling variance in the meta-analyses described below, and we planned to analyze these predicted values in exactly the same ways as we analyzed Z_r in the following analyses.

Preregistration Deviation:

1. Standardizing blue tit out-of-sample predictions (y_i)

Because analysts of blue tit data chose different dependent variables on different scales, after transforming out-of-sample values to the original scales, we standardized all values as z scores ('standard scores') to put all dependent variables on the same scale and make them comparable. This involved taking each relevant value on the original scale (whether a predicted point estimate or a SE associated with that estimate) and subtracting the value in question from the mean value of that dependent variable derived from the full dataset and then dividing this difference by the standard deviation, SD, corresponding to the mean from the full dataset (Supplementary Material B, Equation B.1).

Note that we were unable to standardise some analyst-constructed variables, so these analyses were excluded from the final out-of-sample estimates meta-analysis, see Supplementary Material B, section B.1.2.1 for details and explanation.

2. Log-transforming *Eucalyptus* out-of-sample predictions y_i

All analyses of the *Eucalyptus* data chose dependent variables that were on the same scale, that is, *Eucalyptus* seedling counts. Although analysts may have used different size-classes of *Eucalyptus* seedlings for their dependent variable, we considered these choices to be akin to sub-setting, rather than as different response variables, since changing the size-class of the dependent variable ultimately results in observations being omitted or included. Consequently, we did not standardise *Eucalyptus* out-of-sample predictions.

We were unable to fit quasi-Poisson or Poisson meta-regressions, as desired [79], because available meta-analysis packages (e.g. metafor:: and metainc::) do not provide implementation for outcomes as estimates-only, methods are only provided for outcomes as ratios or rate-differences between two groups. Consequently, we log-transformed the out-of-sample predictions for the *Eucalyptus* data and use the mean estimate for each prediction scenario as the dependent variable in our meta-analysis with the associated SE as the sampling variance in the meta-analysis [74, 75]. Table 2.

We plotted individual effect size estimates (Z_r) and predicted values of the dependent variable (y_i) and their corresponding 95% confidence / credible intervals in forest plots to allow visualization of the range and precision of effect size and predicted values. Further, we included these estimates in random effects meta-analyses [14, 43] using the *metafor* package in R [87, 114]:

$$Z_r \sim 1 + (1|Effect\ ID)$$

$$y_i \sim 1 + (1|Effect\ ID)$$

where y_i is the predicted value for the dependent variable at the 25th percentile, median, or 75th percentile of the independent variables. The individual Z_r effect sizes were weighted with the inverse of sampling variance for Z_r . The individual predicted values for dependent variable (y_i) were weighted by the inverse of the associated SE^2 (original registration omitted "inverse of the" in error). These analyses provided an average Z_r score (\bar{Z}_r) or an average y_i (\bar{y}_i) with corresponding 95% confidence interval and allowed us to estimate two heterogeneity indices, τ^2 and I^2 . The former, τ^2 , is the absolute measure of heterogeneity or the between-study variance (in our case, between-effect variance) whereas I^2 is a relative measure of heterogeneity. We obtained the estimate of relative heterogeneity (I^2) by dividing the between-effect variance by the sum of between-effect and within-effect variance (sampling error variance). I^2 is thus, in a standard meta-analysis, the proportion of variance that is due to heterogeneity as opposed to sampling error. When calculating I^2 , within-study variance is amalgamated across studies to create a "typical" within-study variance which serves as the sampling error variance [14, 43]. Our goal

here was to visualize and quantify the degree of variation among analyses in effect size estimates [72]. We did not test for statistical significance.

Additional explanation:

Our use of I^2 to quantify heterogeneity violates an important assumption, but this violation does not invalidate our use of I^2 as a metric of how much heterogeneity can derive from analytical decisions. In standard meta-analysis, the statistic I^2 quantifies the proportion of variance that is greater than we would expect if differences among estimates were due to sampling error alone [88]. However, it is clear that this interpretation does not apply to our value of I^2 because I^2 assumes that each estimate is based on an independent sample (although these analyses can account for non-independence via hierarchical modelling), whereas all our effects were derived from largely or entirely overlapping subsets of the same dataset. Despite this, we believe that I^2 remains a useful statistic for our purposes. This is because, in calculating I^2 , we are still setting a benchmark of expected variation due to sampling error based on the variance associated with each separate effect size estimate, and we are assessing how much (if at all) the variability among our effect sizes exceeds what would be expected had our effect sizes been based on independent data. In other words, our estimates can tell us how much proportional heterogeneity is possible from analytical decisions alone when sample sizes (and therefore meta-analytic within-estimate variance) are similar to the ones in our analyses. Among other implications, our violation of the independent sample assumption means that we (dramatically) over-estimate the variance expected due to sampling error, and because I^2 is a proportional estimate, we thus underestimate the actual proportion of variance due to differences among analyses other than sampling error. However, correcting this underestimation would create a trivial value since we designed the study so that much of the variance would derive from analytic decisions as opposed to differences in sampled data. Instead, retaining the I^2 value as typically calculated provides a useful comparison to I^2 values from typical meta-analyses.

Interpretation of τ^2 also differs somewhat from traditional meta-analysis, and we discuss this further in the Results.

Finally, we assessed the extent to which deviations from the meta-analytic mean by individual effect sizes (Z_r) or the predicted values of the dependent variable (y_i) were explained by the peer rating of each analysis team's method section, by a measurement of the distinctiveness of the set of predictor variables included in each analysis, and by the choice of whether or not to include random effects in the model. The deviation score, which served as the dependent variable in these analyses, is the absolute value of the difference between the meta-analytic mean (\bar{Z}_r or \bar{y}_i) and the individual Z_r (or y_i) estimate for each analysis. We used the Box-Cox transformation on the absolute values of deviation scores to achieve an approximately normal distribution [28, 29]. We described variation in this dependent variable with both a series of univariate analyses and a multivariate analysis. All these analyses were general linear (mixed) models. These analyses were secondary to our estimation of variation in effect sizes described above. We wished to quantify relationships among variables, but we had no a priori expectation of effect size and made no dichotomous decisions about statistical significance.

When examining the extent to which reviewer ratings (on a scale from 0 to 100) explained deviation from the average effect (or predicted value), each analysis had been rated by multiple peer reviewers, so for each reviewer score to be included, we include each deviation score in the analysis multiple times. To account for the non-independence of multiple ratings of the same analysis, we planned to include analysis identity as a random effect in our general linear mixed model in the *lme4* package in R [11, 87]. To account for potential differences among reviewers in their scoring of analyses, we also planned to include reviewer identity as a random effect:

$$\text{DeviationScore}_j = \text{BoxCox}(\text{DeviationFromMean}_j)$$

$$\text{DeviationScore}_{ij} \sim \text{Rating}_{ij} + \text{ReviewerID}_i + \text{EffectID}_j$$

$$\text{ReviewerID}_i \sim N(0, \sigma_i^2)$$

$$\text{EffectID}_j \sim N(0, \sigma_j^2)$$

where $\text{DeviationFromMean}_j$ is the deviation from the meta-analytic mean for the j th analysis, ReviewerID_i is the random intercept assigned to each i reviewer, and EffectID_j is the random intercept assigned to each j analysis, both of which are assumed to be normally distributed with a mean of 0 and a variance of σ^2 . Absolute deviation scores were Box-Cox transformed using the `step_box_cox()` function from the *timetk* package in R [24, 87].

Additional explanation:

In our meta-analyses based on Box-Cox transformed deviation scores, we leave these deviation scores unweighted. This is consistent with our registration, which did not mention weighting these scores. However, the fact that we did not mention weighting the scores was actually an error: we had intended to weight them, as is standard in meta-analysis, using the inverse variance of the Box-Cox transformed deviation scores Supplementary Material C, equation C.1. Unfortunately, when we did conduct the weighted analyses, they produced results in which some weighted estimates differed radically from the unweighted estimate because the weights were invalid. Such invalid weights can sometimes occur when the variance (upon which the weights depend) is partly a function of the effect size, as in our Box-Cox transformed deviation scores [73]. In the case of the *Eucalyptus* analyses, the most extreme outlier was weighted much more heavily (by close to two orders of magnitude) than any other effect sizes because the effect size was, itself, so high. Therefore, we made the decision to avoid weighting by inverse variance in all analyses of the Box-Cox transformed deviation scores. This was further justified because (a) most analyses have at least some moderately unreliable weights, and (b) the sample sizes were mostly very similar to each other across submitted analyses, and so meta-analytic weights are not particularly important here [19]. We systematically investigated the impact of different weighting schemes and random effects on model convergence and results, see Supplementary Material C, section C.8 for more details.

We conducted a similar analysis with the four categories of reviewer ratings ((1) deeply flawed and unpublishable, (2) publishable with major revision, (3) publishable with minor revision, (4) publishable as is) set as ordinal predictors numbered as shown here. As with the analyses above, we planned for these analyses to also include random effects of analysis identity and reviewer identity. Both of these analyses (1: 1–100 ratings as the fixed effect, 2: categorical ratings as the fixed effects) were planned to be conducted eight times for each dataset. Each of the four responses (Z_r , Y_{25} , Y_{50} , Y_{75}) were to be compared once to the initial ratings provided by the peer reviewers, and again based on the revised ratings provided by the peer reviewers.

Preregistration deviation:

1. We planned to include random effects of both analysis identity and reviewer identity in these models comparing reviewer ratings with deviation scores. However, after we received the analyses, we discovered that a subset of analyst teams had either conducted multiple analyses and/or identified multiple effects per analysis as answering the target question. We therefore faced an even more complex potential set of random effects. We decided that including Team ID and Effect ID along with Reviewer ID as random effects in the same model would almost certainly lead to model fit problems, and so we started with simpler models including just Effect ID and Reviewer ID. However, even with this simpler structure, our dataset was sparse, with reviewers rating a small number of analyses, resulting in models with singular fit (Supplementary Material C, section C.2). Removing one of the random effects was necessary for the models to converge. For both models of deviation from the meta-analytic mean explained by categorical or continuous reviewer ratings, we removed the random effect of Effect ID, leaving Reviewer ID as the only random effect.
 2. We conducted analyses only with the final peer ratings after the opportunity for revision, not with the initial ratings. This was because when we recorded the final ratings, the initial ratings were over-written, therefore we did not have access to those initial values.
-

The next set of univariate analyses sought to explain deviations from the mean effects based on a measure of the distinctiveness of the set of variables included in each analysis. As a ‘distinctiveness’ score, we used Sorensen’s Similarity Index (an index typically used to compare species composition across sites), treating variables as species and individual analyses as sites. To generate an individual Sorensen’s value for each analysis required calculating the pairwise Sorensen’s value for all pairs of analyses (of the same dataset), and then taking the average across these Sorensen’s values for each analysis. We calculated the Sorensen’s index values using the *beta-part* package [10] in R:

$$\beta_{\text{Sorensen}} = \frac{b + c}{2a + b + c}$$

where a is the number of variables common to both analyses, b is the number of variables that occur in the first analysis but not in the second and c is the number of variables that occur in the second analysis. We then used the per-model average Sorensen’s index value as an independent variable to predict the deviation score in a general linear model, and included no random effect since each analysis is included only once, in R [87]:

Additional explanation:

When we planned this analysis, we anticipated that analysts would identify a single primary effect from each model, so that each model would appear in the analysis only once. Our expectation was incorrect because some analysts identified >1 effect per analysis, but we still chose to specify our model as registered and not use a random effect. This is because most models produced only one effect and so we expected that specifying a random effect to account for the few cases where >1 effect was included for a given model would prevent model convergence.

Note that this analysis contrasts with the analyses in which we used reviewer ratings as predictors because in the analyses with reviewer ratings, each effect appeared in the analysis approximately four times due to multiple reviews of each analysis, and so it was much more important to account for that variance through a random effect.

$$\text{DeviationScore}_j \sim \beta_{\text{Sorensen}_j}$$

Next, we assessed the relationship between the inclusion of random effects in the analysis and the deviation from the mean effect size. We anticipated that most analysts would use random effects in a mixed model framework, but if we were wrong, we wanted to evaluate the differences in outcomes when using random effects versus not using random effects. Thus, if there were at least 5 analyses that did and 5 analyses that did not include random effects, we would add a binary predictor variable “random effects included (yes/no)” to our set of univariate analyses and would add this predictor variable to our multivariate model described below. This standard was only met for the *Eucalyptus* analyses, and so we only examined inclusion of random effects as a predictor variable in meta-analysis of this set to analyses.

Finally, we conducted a multivariate analysis with the five predictors described above (peer ratings 0–100 and peer ratings of publishability 1–4; both original and revised and Sorensen’s index, plus a sixth for *Eucalyptus*, presence / absence of random effects) with random effects of analysis identity and reviewer identity in the *lme4* package in R [11, 87]. We had stated here in the text that we would use only the revised (final) peer ratings in this analysis, so the absence of the initial ratings is not a deviation from our plan:

$$\text{DeviationScore}_j = \text{BoxCox}(\text{DeviationFromMean}_j)$$

$$\text{DeviationScore}_{ij} \sim \text{RatingContinuous}_{ij} + \text{RatingCategorical}_{ij} \\ + \beta \text{Sorensen}_j + \text{ReviewerID}_i + \text{Effect ID}_j$$

$$\text{ReviewerID}_i \sim N(0, \sigma_i^2)$$

$$\text{EffectID}_j \sim N(0, \sigma_j^2)$$

We conducted all the analyses described above eight times; for each of the four responses (Z_r , Y_{25} , Y_{50} , Y_{75}) one time for each of the two datasets.

We have publicly archived all relevant data, code, and materials on the Open Science Framework (<https://osf.io/mn5aj/>). Archived data includes the original datasets distributed to all analysts, any edited versions of the data analyzed by individual groups, and the data we analyzed with our meta-analyses, which include the effect sizes derived from separate analyses, the statistics describing variation in model structure among analyst groups, and the anonymized answers to our surveys of analysts and peer reviewers. Similarly, we have archived both the analysis code used for each individual analysis (where available) and the code from our meta-analyses. We have also archived copies of our survey instruments from analysts and peer reviewers.

Our rules for excluding data from our study were as follows. We excluded from our synthesis any individual analysis submitted after we had completed peer review or those unaccompanied by analysis files that allow us to understand what the analysts did. We also excluded any individual analysis that did not produce an outcome that could be interpreted as an answer to our primary question (as posed above) for the respective dataset. For instance, this means that in the case of the data on blue tit chick growth, we excluded any analysis that did not include something that can be interpreted as growth or size as a dependent (response) variable, and in the case of the *Eucalyptus* establishment data, we excluded any analysis that did not include a measure of grass cover among the independent (predictor) variables. Also, as described above, any analysis that could not produce an effect that could be converted to a signed Z_r was excluded from analyses of Z_r .

Preregistration Deviation:

Some analysts had difficulty implementing our instructions to derive the out-of-sample predictions, and in some cases (especially for the *Eucalyptus* data), they submitted predictions with implausibly extreme values. We believed these values were incorrect and thus made the conservative decision to exclude out-of-sample predictions where the estimates were > 3 standard deviations from the mean value from the full dataset provided to teams for analysis.

Additional explanation: Best practices in many-analysts research

After we initiated our project, a paper was published outlining best practices in many-analysts studies [1]. Although we did not have access to this document when we implemented our project, our study complies with these practices nearly completely. The one exception is that although we requested analysis code from analysts, we did not require submission of code.

Additional explanation: unregistered analyses

1. Evaluating model fit.

We evaluated all fitted models using the `performance::performance()` function from the *performance* package [60] and the `glance()` function from the *broom.mixed* package [13]. For all models, we calculated the square root of the residual variance (Sigma) and the root mean squared error (RMSE). For GLMMs `performance::performance()` calculates the marginal and conditional R^2 values as well as the contribution of random effects (ICC), based on [76]. The conditional R^2 accounts for both the fixed and random effects, while the marginal R^2 considers only the variance of the fixed effects. The contribution of random effects is obtained by subtracting the marginal R^2 from the conditional R^2 .

2. Exploring outliers and analysis quality.

After seeing the forest plots of Z_r values and noticing the existence of a small number of extreme outliers, especially from the *Eucalyptus* analyses, we wanted to understand the degree to which our heterogeneity estimates were influenced by these outliers. To explore this question, we removed the highest two and lowest two values of Z_r in each dataset and re-calculated our heterogeneity estimates.

To help understand the possible role of the quality of analyses in driving the heterogeneity we observed among estimates of Z_r , we created forest plots and recalculated our heterogeneity estimates after removing all effects from analysis teams that had received at least one rating of “deeply flawed and unpublishable” and then again after removing all effects from analysis teams with at least one rating of either “deeply flawed and unpublishable” or “publishable with major revisions”. We also used self-identified levels of statistical expertise to examine heterogeneity when we retained analyses only from analysis teams that contained at least one member who rated themselves as “highly proficient” or “expert” (rather than “novice” or “moderately proficient”) in conducting statistical analyses in their research area in our intake survey.

Additionally, to assess potential impacts of highly collinear predictor variables on estimates of Z_r in blue tit analyses, we created forest plots (Supplementary Material B, Figure B.5) and recalculated our heterogeneity estimates after we removed analyses that contained the brood count after manipulation and the highly correlated (correlation of 0.89, Supplementary Material D, Figure D.2) brood count at day 14. This removal included the one effect based on a model that contained both these variables and a third highly correlated variable, the estimate of number of chicks fledged (the only model that included the estimate of number of chicks fledged). We did not conduct a similar analysis for the *Eucalyptus* dataset because there were no variables highly collinear with the primary predictors (grass cover variables) in that dataset (Supplementary Material D, Figure D.1).

3. Exploring possible impacts of lower quality estimates of degrees of freedom.

Our meta-analyses of variation in Z_r required variance estimates derived from estimates of the degrees of freedom in original analyses from which Z_r estimates were derived. While processing the estimates of degrees of freedom submitted by analysts, we identified a subset of these estimates in which we had lower confidence because two or more effects from the same analysis were submitted with identical degrees of freedom. We therefore conducted a second set of (more conservative) meta-analyses that excluded these Z_r estimates with identical estimates of degrees of freedom and we present these analyses in the supplement.

Step 6: Facilitated discussion and collaborative write-up of manuscript

We planned for analysts and initiating authors to discuss the limitations, results, and implications of the study and collaborate on writing the final manuscript for review as a stage-2 Registered Report.

Preregistration deviation:

As described above, due to the large number of recruited analysts and reviewers and the anticipated challenges of receiving and integrating feedback from so many authors, we limited analyst and reviewer participation in the production of the final manuscript to an invitation to call attention to serious problems with the manuscript draft.

We built an R package, `ManyEcoEvo`: to conduct the analyses described in this study [38], which can be downloaded from <https://github.com/egouldo/ManyEcoEvo/> to reproduce our analyses or replicate the analyses described here using alternate datasets. Data cleaning and preparation of analysis data, as well as the analysis, is conducted in R [87] reproducibly using the `targets` package ([57]). This data and analysis pipeline is stored in the `ManyEcoEvo`: package repository and its outputs are made available to users of the package when the library is loaded.

The full manuscript, including further analysis and presentation of results is written in Quarto [2]. The source code to reproduce the manuscript is hosted at <https://github.com/egouldo/ManyAnalysts/> [39], and the rendered version of the source code may be viewed at <https://egouldo.github.io/ManyAnalysts/>. All R packages and their versions used in the production of the manuscript are listed in Table 7 at the end of this paper.

Results

Summary statistics

In total, 173 analyst teams, comprising 246 analysts, contributed 182 usable analyses (compatible with our meta-analyses and provided with all information needed for inclusion) of the two datasets examined in this study which yielded 215 effects. Analysts produced 134 distinct effects that met our criteria for inclusion in at least one of our meta-analyses for the blue tit dataset. Analysts produced 81 distinct effects meeting our criteria for inclusion for the *Eucalyptus* dataset. Excluded analyses and effects either did not answer our specified biological questions, were submitted with insufficient information for inclusion in our meta-analyses, or were incompatible with production of our effect size(s). We expected cases of this final scenario (incompatible analyses), for instance we cannot extract a Z_r from random forest models, which is why we analyzed two distinct types of effects, Z_r and out-of-sample predictions. Some effects only provided sufficient information for a subset of analyses and were

only included in that subset. For both datasets, most submitted analyses incorporated mixed effects. Submitted analyses of the blue tit dataset typically specified normal error and analyses of the *Eucalyptus* dataset typically specified a non-normal error distribution (Supplementary Material A, Table A.1).

For both datasets, the composition of models varied substantially in regards to the number of fixed and random effects, interaction terms, and the number of data points used, and these patterns differed somewhat between the blue tit and *Eucalyptus* analyses (See Supplementary Material A Table A. 2). Focusing on the models included in the Z_r analyses (because this is the larger sample), blue tit models included a similar number of fixed effects on average (mean 5.2 ± 2.92 SD, range 1 to 19) as *Eucalyptus* models (mean 5.01 ± 3.83 SD, range 1 to 13), but the standard deviation in the number of fixed effects was somewhat larger in the *Eucalyptus* models. The average number of interaction terms was much larger for the blue tit models (mean 0.44 ± 1.11 SD, range 0 to 10) than for the *Eucalyptus* models (mean 0.16 ± 0.65 SD, range 0 to 5), but still under 0.5 for both, indicating that most models did not contain interaction terms. Blue tit models also contained more random effects (mean 3.53 ± 2.08 SD, range 0 to 10) than *Eucalyptus* models (mean 1.41 ± 1.09 SD, range 0 to 4). The maximum possible sample size in the blue tit dataset (3720 nestlings) was an order of magnitude larger than the maximum possible in the *Eucalyptus* dataset (351 plots), and the means and standard deviations of the sample size used to derive the effects eligible for our study were also an order of magnitude greater for the blue tit dataset (mean 2611.09 ± 937.48 SD, range 76 to 76) relative to the *Eucalyptus* models (mean 298.43 ± 106.25 SD, range 18 to 351). However, the standard deviation in sample size from the *Eucalyptus* models was heavily influenced by a few cases of dramatic sub-setting (described below). Approximately three quarters of *Eucalyptus* models used sample sizes within 3% of the maximum. In contrast, fewer than 20% of blue tit models relied on sample sizes within 3% of the maximum, and approximately 50% of blue tit models relied on sample sizes 29% or more below the maximum.

Analysts provided qualitative descriptions of the conclusions of their analyses. Each analysis team provided one conclusion per dataset. These conclusions could take into account the results of any formal analyses completed by the team as well as exploratory and visual analyses of the data. Here we summarize all qualitative responses, regardless of whether we had sufficient information to use the corresponding model results in our quantitative analyses below. We classified these conclusions into the categories summarized below (Table 1):

- *Mixed*: some evidence supporting a positive effect, some evidence supporting a negative effect
- *Conclusive negative*: negative relationship described without caveat
- *Qualified negative*: negative relationship but only in certain circumstances or where analysts express uncertainty in their result
- *Conclusive none*: analysts interpret the results as conclusive of no effect
- *Qualified none*: analysts describe finding no evidence of a relationship but they describe the potential for an undetected effect
- *Qualified positive*: positive relationship described but only in certain circumstances or where analysts express uncertainty in their result
- *Conclusive positive*: positive relationship described without caveat

For the blue tit dataset, most analysts concluded that there was negative relationship between measures of sibling competition and nestling growth, though half the teams expressed qualifications or described effects as mixed or absent. No analysts concluded that there was a positive relationship even though some individual effect sizes were positive, apparently because all analysts who produced effects indicating positive relationships also produced effects indicating negative relationships and therefore described their results as qualified, mixed, or absent. For the *Eucalyptus* dataset, there was a broader spread of conclusions with at least one analyst team providing conclusions consistent with each conclusion category. The most common conclusion for the *Eucalyptus* dataset was that there was no relationship between grass cover and *Eucalyptus* recruitment (either conclusive or qualified description of no relationship), but more than half the teams concluded that there were effects; negative, positive, or mixed.

Distribution of effects

Effect sizes (Z_r)

Although the majority (118 of 131) of the usable Z_r effects from the blue tit dataset found nestling growth decreased with sibling competition, and the meta-analytic mean \bar{Z}_r (Fisher's transformation of the correlation coefficient) was convincingly negative (-0.35 ± 0.06 95%CI), there was substantial variability in the strength and the direction of this effect. Z_r ranged from -1.55 to 0.38 , and approximately continuously from -0.93 to 0.19 (Fig. 2a and Table 4), and of the 118 effects with negative slopes, 93 had confidence intervals excluding 0. Of the 13 with positive slopes indicating increased nestling growth in the presence of more siblings, 2 had confidence intervals excluding zero (Fig. 2a).

Meta-analysis of the *Eucalyptus* dataset also showed substantial variability in the strength of effects as measured by Z_r , and unlike with the blue tits, a notable lack of consistency in the direction of effects (Fig. 2b, Table 4). Z_r ranged from -4.47 (Supplementary Material A, Figure A.2), indicating a strong tendency for reduced *Eucalyptus* seedling success as grass cover increased, to 0.39 , indicating the opposite. Although the range of reported effects skewed strongly negative, this was due to a small number of substantial outliers. Most values of Z_r were relatively small with values $<|0.2|$ and the meta-analytic mean effect size was close to zero (-0.09 ± 0.12 95%CI). Of the 79 effects, fifty-three had confidence intervals overlapping zero, approximately a quarter (fifteen) crossed the traditional threshold of statistical significance indicating a negative relationship between grass cover and seedling success, and eleven crossed the significance threshold indicating a positive relationship between grass cover and seedling success (Fig. 2b).

Out-of-sample predictions (y_i)

As with the effect size Z_r , we observed substantial variability in the size of out-of-sample predictions derived from the analysts' models. Blue tit predictions (Fig. 3a), which were z -score-standardized to accommodate the use of different response variables, always ranged far in excess of one standard deviation. In the y_{25} scenario, model predictions ranged from -1.84 to 0.42 (a range of 2.68 standard deviations), in the y_{50} they ranged from -0.52 to 1.08 (a range of 1.63 standard deviations), and in the y_{75} scenario they ranged from -0.03 to 1.59 (a range of 1.9 standard deviations). As should be expected given the existence of both negative and positive Z_r values, all three out-of-sample scenarios produced both negative and positive predictions, although as with the Z_r values, there is a clear trend for scenarios with more siblings to be associated with smaller nestlings. This is supported by the meta-analytic means of these three sets of predictions which were -0.66 (95%CI -0.82 – 0.5) for the y_{25} , 0.34 (95%CI 0.2 – 0.48) for the y_{50} , and 0.67 (95%CI 0.57 – 0.77) for the y_{75} .

Eucalyptus out-of-sample predictions also varied substantially (Fig. 3b), but because they were not z -score-standardized and are instead on the original count scale, the types of interpretations we can make differ. The predicted *Eucalyptus* seedling counts per 15×15 m plot for the y_{25} scenario ranged from 0.04 to 26.99 , for the y_{50} scenario ranged from 0.04 to 44.34 , and for the y_{75} scenario they ranged from 0.03 to 61.34 . The meta-analytic mean predictions for these three scenarios were similar; 1.27 (95%CI 0.59 – 2.3) for the y_{25} , 2.92 (95%CI 0.98 – 3.89) for the y_{50} , and 2.92 (95%CI 1.59 – 4.9) for the y_{75} scenarios respectively.

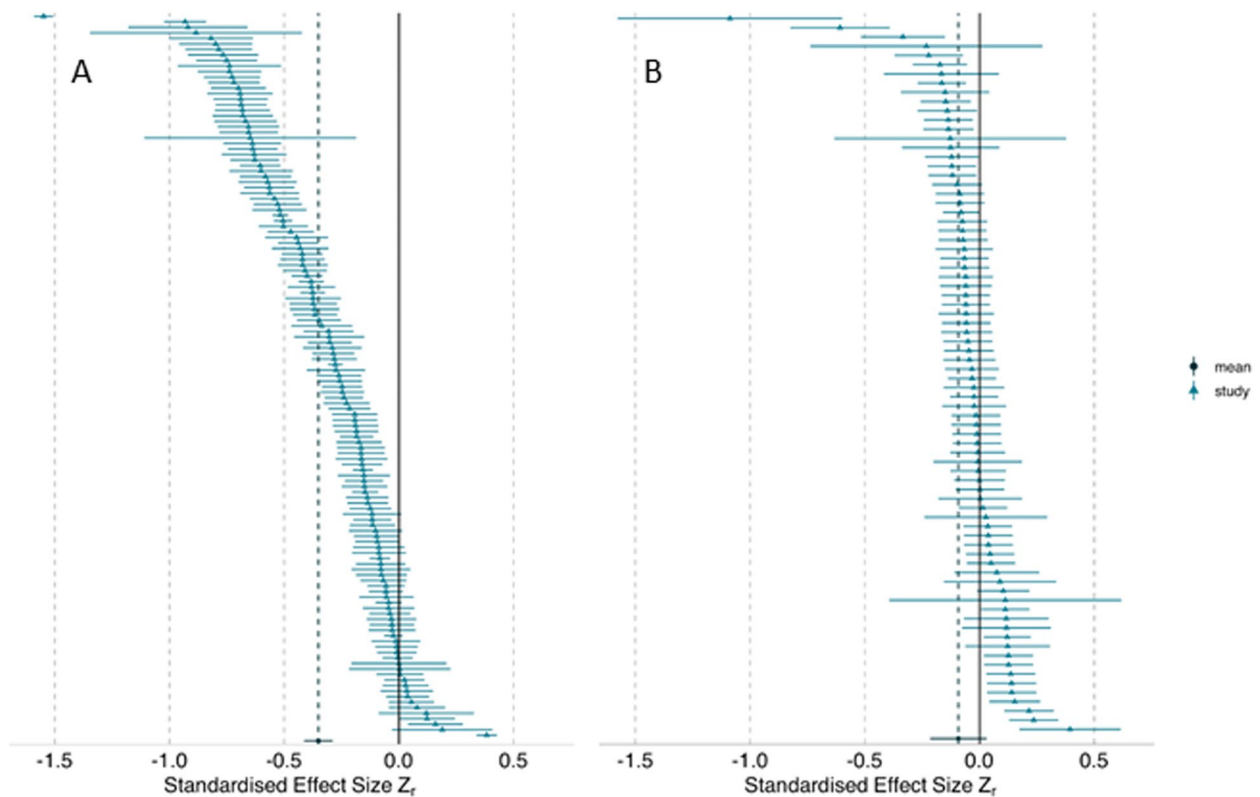


Fig. 2 Forest plots of meta-analytic estimated standardized effect sizes (Z_r , blue triangles) and their 95% confidence intervals for each effect size included in the meta-analysis model. **A** Blue tit analyses: Points where Z_r are less than 0 indicate analyses that found a negative relationship between sibling number and nestling growth. **B** *Eucalyptus* analyses: Points where Z_r are less than 0 indicate a negative relationship between grass cover and *Eucalyptus* seedling success. The meta-analytic mean effect size is denoted by a black circle and a dashed vertical line, with error bars also representing the 95% confidence interval. The solid black vertical line demarcates effect size of 0, indicating no relationship between the test variable and the response variable. Note that the *Eucalyptus* plot omits one extreme outlier with the value of -4.47 (Supplementary Material A, Figure A.2) in order to standardize the x-axes on these two panels

Table 1 Tallies of analysts’ qualitative answers to the research questions addressed by their analyses

Dataset	Mixed	Negative Conclusive	Negative Qualified	None Conclusive	None Qualified	Positive Qualified	Positive Conclusive
blue tit	5	37	27	4	1	0	0
<i>Eucalyptus</i>	8	6	12	19	12	4	2

Quantifying heterogeneity

Effect sizes (Z_r)

We quantified both absolute (τ^2) and relative (I^2) heterogeneity resulting from analytical variation. Both measures suggest that substantial variability among effect sizes was attributable to the analytical decisions of analysts.

The total absolute level of variance beyond what would typically be expected due to sampling error, τ^2 (Table 2), among all usable blue tit effects was 0.08 and for *Eucalyptus* effects was 0.27. This is similar to or exceeding the median value (0.105) of τ^2 found across 31 recent

meta-analyses (calculated from the data in [121]). The similarity of our observed values to values from meta-analyses of different studies based on different data suggests the potential for a large portion of heterogeneity to arise from analytical decisions. For further discussion of interpretation of τ^2 in our study, please consult discussion of post hoc analyses below.

In our analyses, I^2 is a plausible index of how much more variability among effect sizes we have observed, as a proportion, than we would have observed if sampling error were driving variability. We discuss our

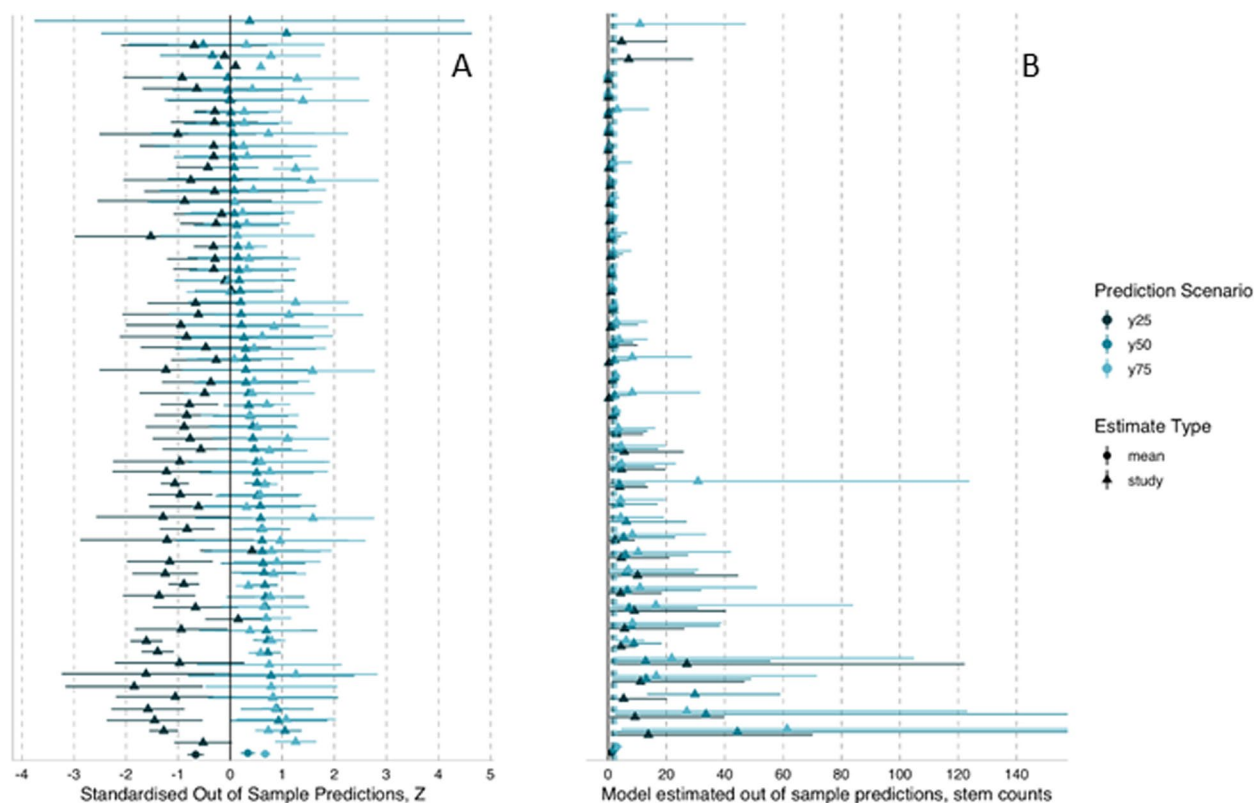


Fig. 3 Forest plot of meta-analytic estimated out-of-sample predictions. **A** Standardized (z-score) blue tit out-of-sample predictions, Z_r . **B** Response-scale (stem counts) *Eucalyptus* out-of-sample predictions. Triangles represent individual estimates. Circles represent the meta-analytic mean for each prediction scenario. Dark-blue points correspond to y_{25} scenario, medium-blue points correspond to the y_{50} scenario, while light blue points correspond to the y_{75} scenario. Error bars are 95% confidence intervals. Note that, for the *Eucalyptus* analysis, outliers (observations more than 3 SD above the mean) have been removed prior to model fitting and do not appear on this figure. The x-axis is truncated to approximately 140, and thus some error bars are incomplete. See Supplementary Material B, Figure B.6 for full figure

interpretation of I^2 further in the methods, but in short, it is a useful metric for comparison to values from published meta-analyses and provides a plausible value for how much heterogeneity could arise in a normal meta-analysis with similar sample sizes due to analytical variability alone. In our study, total I^2 for the blue tit Z_r estimates was extremely large, at 97.61%, as was the *Eucalyptus* estimate (98.59% Table 2).

Although the overall I^2 values were similar for both *Eucalyptus* and blue tit analyses, the relative composition of that heterogeneity differed. For both datasets, the majority of heterogeneity in Z_r was driven by differences among effects as opposed to differences among teams, though this was more prominent for the *Eucalyptus* dataset, where nearly all of the total heterogeneity was driven by differences among effects (91.7%) as opposed to differences among teams (6.89%) (Table 2).

Out-of-sample predictions (y_i)

We observed substantial heterogeneity among out-of-sample estimates, but the pattern differed somewhat

from the Z_r values (Table 3). Among the blue tit predictions, I^2 ranged from medium-high for the y_{25} scenario (68.54) to low (27.9) for the y_{75} scenario. Among the *Eucalyptus* predictions, I^2 values were uniformly high (>82%). For both datasets, most of the existing heterogeneity among predicted values was attributable to among-team differences, with the exception of the y_{50} analysis of the *Eucalyptus* dataset. We are limited in our interpretation of τ^2 for these estimates because, unlike for the Z_r estimates, we have no benchmark for comparison with other meta-analyses.

Post hoc analysis: Exploring outlier characteristics and the effect of outlier removal on heterogeneity Effect sizes (Z_r)

The outlier *Eucalyptus* Z_r values were striking and merited special examination. The three negative outliers had very low sample sizes that were based on either small subsets of the dataset or, in one case, extreme aggregation of data. The outliers associated with small subsets had sample sizes ($n=117, 90, 18$) that were less than

Table 2 Heterogeneity in the estimated effects Z_i for meta-analyses of: the full dataset, as well as from post hoc analyses wherein analyses with outliers are removed, analyses with effects from analysis teams with at least one “unpublishable” rating are excluded, analyses receiving at least one “major revisions” rating or worse excluded, analyses from teams with at least one analyst self-rated as “highly proficient” or “expert” in statistical analysis are included, and (blue tit only) analyses that did not included the pair of highly collinear predictors together. τ^2_{Team} is the absolute heterogeneity for the random effect Team. $\tau^2_{\text{Effect ID}}$ is the absolute heterogeneity for the random effect Effect ID nested under Team. Effect ID is the unique identifier assigned to each individual statistical effect submitted by an analysis team. We nested Effect ID within analysis team identity (Team) because analysis teams often submitted >1 statistical effect, either because they considered >1 model or because they derived >1 effect per model, especially when a model contained a factor with multiple levels that produced >1 contrast. τ^2_{Total} is the total absolute heterogeneity. I^2_{Total} is the proportional heterogeneity; the proportion of the variance among effects not attributable to sampling error, I^2_{Team} is the subset of the proportional heterogeneity due to differences among Teams and $I^2_{\text{Team, EffectID}}$ is subset of the proportional heterogeneity attributable to among-Effect ID differences

Dataset	N _{Obs}	τ^2_{Total}	τ^2_{Team}	τ^2_{EffectID}	I^2_{Total}	I^2_{Team}	$I^2_{\text{Team, EffectID}}$
All Analyses							
<i>Eucalyptus</i>	79	0.27	0.02	0.25	98.59%	6.89%	91.70%
blue tit	131	0.08	0.03	0.05	97.61%	36.71%	60.90%
Blue tit analyses containing highly collinear predictors removed							
blue tit	117	0.07	0.04	0.03	96.92%	58.18%	38.75%
All analyses, outliers removed							
<i>Eucalyptus</i>	75	0.01	0.00	0.01	66.19%	19.25%	46.94%
blue tit	127	0.07	0.04	0.02	96.84%	64.63%	32.21%
Analyses receiving at least one “Unpublishable” rating removed							
<i>Eucalyptus</i>	55	0.01	0.01	0.01	79.74%	28.31%	51.43%
blue tit	109	0.08	0.03	0.05	97.52%	35.68%	61.84%
Analyses receiving at least one “Unpublishable” and or “Major Revisions” rating removed							
<i>Eucalyptus</i>	13	0.03	0.03	0.00	88.91%	88.91%	0.00%
blue tit	32	0.14	0.01	0.13	98.72%	5.17%	93.55%
Analyses from teams with highly proficient or expert data analysts							
<i>Eucalyptus</i>	34	0.58	0.02	0.56	99.41%	3.47%	95.94%
blue tit	89	0.09	0.03	0.06	97.91%	31.43%	66.49%

half of the total possible sample size of 351. The case of extreme aggregation involved averaging all values within each of the 351 sites in the dataset.

Surprisingly, both the largest and smallest effect sizes in the blue tit analyses (Fig. 2a) come from the same analyst (anonymous ID: “Adelong”), with identical models in terms of the explanatory variable structure, but with different response variables. However, the radical change in effect was primarily due to collinearity with covariates. The primary predictor variable (brood count after manipulation) was accompanied by several collinear variables, including the highly collinear (correlation of 0.89 Supplementary Material D, Figure D.2) covariate (brood count at day 14) in both analyses. In the analysis of nestling weight, brood count after manipulation showed a strong positive partial correlation with weight after controlling for brood count at day 14 and treatment category (increased, decreased, unmanipulated). In that same analysis, the most collinear covariate (the day 14 count) had a negative partial correlation with weight. In the analysis with tarsus length as the response variable,

these partial correlations were almost identical in absolute magnitude, but reversed in sign and so brood count after manipulation was now the collinear predictor with the negative relationship. The two models were therefore very similar, but the two collinear predictors simply switched roles, presumably because a subtle difference in the distribution of weight and tarsus length data.

When we dropped the *Eucalyptus* outliers, I^2 decreased from high (98.59 %), using [43] suggested benchmark, to between moderate and high (66.19 %, Table 2). However, more notably, τ^2 dropped from 0.27 to 0.01, indicating that, once outliers were excluded, the observed variation in effects was similar to what we would expect if sampling error were driving the differences among effects (since τ^2 is the variance beyond that driven by sampling error). The interpretation of this value of τ^2 in the context of our many-analyst study is somewhat different than a typical meta-analysis, however, since in our study (especially for *Eucalyptus*, where most analyses used almost exactly the same data points), there is almost no role for sampling error in driving the observed differences among

Table 3 Heterogeneity among the out-of-sample predictions y_i for both blue tit and *Eucalyptus* datasets. τ^2_{Team} is the absolute heterogeneity for the random effect Team. $T^2_{EffectID}$ is the absolute heterogeneity for the random effect Effect ID nested under Team. Effect ID is the unique identifier assigned to each individual statistical effect submitted by an analysis team. We nested Effect ID within analysis team identity (Team) because analysis teams often submitted >1 statistical effect, either because they considered >1 model or because they derived >1 effect per model, especially when a model contained a factor with multiple levels that produced >1 contrast. τ^2_{Total} is the total absolute heterogeneity. I^2_{Total} is the proportional heterogeneity; the proportion of the variance among effects not attributable to sampling error, I^2_{Team} is the subset of the proportional heterogeneity due to differences among Teams and $I^2_{Team,EffectID}$ is subset of the proportional heterogeneity attributable to among-Effect ID differences

Prediction Scenario	N_{Obs}	T_{Total}	T^2_{Team}	$T^2_{EffectID}$	I^2_{Total}	I^2_{Team}	$I^2_{Team,EffectID}$
blue tit							
y25	63	0.23	0.11	0.03	68.54%	53.43%	15.11%
y50	60	0.23	0.06	0.00	50%	46.29%	3.71%
y75	63	0.23	0.02	0.00	27.9%	27.89%	0.01%
<i>Eucalyptus</i>							
y25	38	5.75	1.48	0.68	86.93%	59.54%	27.39%
y50	38	5.75	1.32	0.83	89.63%	55%	34.64%
y75	38	5.75	1.03	0.41	80.19%	57.41%	22.78%

Table 4 Estimated mean value of the standardized correlation coefficient, \bar{Z}_r , along with its standard error and 95% confidence intervals. We re-computed the meta-analysis for different post hoc subsets of the data: All eligible effects, removal of effects from blue tit analyses that contained a pair of highly collinear predictor variables, removal of effects from analysis teams that received at least one peer rating of “deeply flawed and unpublishable”, removal of any effects from analysis teams that received at least one peer rating of either “deeply flawed and unpublishable” or “publishable with major revisions”, inclusion of only effects from analysis teams that included at least one member who rated themselves as “highly proficient” or “expert” at conducting statistical analyses in their research area

Dataset	$\hat{\mu}$	$SE[\hat{\mu}]$	95% CI	Statistic	p
All analyses					
<i>Eucalyptus</i>	-0.09	0.06	[-0.22,0.03]	-1.47	0.14
blue tit	-0.35	0.03	[-0.41,-0.29]	-11.02	<0.001
Blue tit analyses containing highly collinear predictors removed					
blue tit	-0.36	0.03	[-0.42,-0.29]	-10.97	<0.001
All analyses, outliers removed					
<i>Eucalyptus</i>	-0.03	0.01	[-0.06,0.00]	-2.23	0.026
blue tit	-0.36	0.03	[-0.42,-0.30]	-11.48	<0.001
Analyses receiving at least one “Unpublishable” rating removed					
<i>Eucalyptus</i>	-0.02	0.02	[-0.07,0.02]	-1.15	0.3
blue tit	-0.36	0.03	[-0.43,-0.30]	-10.82	<0.001
Analyses receiving at least one “Unpublishable” and or “Major Revisions” rating removed					
<i>Eucalyptus</i>	-0.04	0.05	[-0.15,0.07]	-0.77	0.4
blue tit	-0.37	0.07	[-0.51,-0.23]	-5.34	<0.001
Analyses from teams with highly proficient or expert data analysts					
<i>Eucalyptus</i>	-0.17	0.13	[-0.43,0.10]	-1.24	0.2
blue tit	-0.36	0.04	[-0.44,-0.28]	-8.93	<0.001

the estimates. Thus, rather than concluding that the variability we observed among estimates (after removing outliers) was due only to sampling error (because τ^2 became small: 10% of the median from [121]), we instead conclude that the observed variability, which must be due to the divergent choices of analysts rather than sampling error, is approximately of the same magnitude as what we would have expected if, instead, sampling error, and not analytical heterogeneity, were at work. Conversely, dropping outliers from the set of blue tit effects did not meaningfully reduce I^2 , and only modestly reduced τ^2 (Table 2). Thus, effects at the extremes of the distribution were much stronger contributors to total heterogeneity for effects from analyses of the *Eucalyptus* than for the blue tit dataset.

Out-of-sample predictions (y_i)

We did not conduct these post hoc analyses on the out-of-sample predictions as the number of eligible effects was smaller and the pattern of outliers differed.

Post hoc analysis: Exploring the effect of removing analyses with poor peer ratings on heterogeneity Effect sizes (Z_r)

Removing poorly rated analyses had limited impact on the meta-analytic means (Supplementary Material B, Figure B.3). For the *Eucalyptus* dataset, the meta-analytic mean shifted from -0.09 to -0.02 when effects from analyses rated as unpublishable were removed, and to -0.04 when effects from analyses rated, at least once, as unpublishable or requiring major revisions were removed. Further, the confidence intervals for all of these means overlapped each of the other means (Table 4). We

saw similar patterns for the blue tit dataset, with only small shifts in the meta-analytic mean, and confidence intervals of all three means overlapping each other mean (Table 4). Refitting the meta-analysis with a fixed effect for categorical ratings also showed no indication of differences in group meta-analytic means due to peer ratings (Supplementary Material B, Figure B.1).

For the blue tit dataset, removing poorly rated analyses led to only negligible changes in I^2_{Total} and relatively minor impacts on τ^2 . However, for the *Eucalyptus* dataset, removing poorly rated analyses led to notable reductions in I^2_{Total} and substantial reductions in τ^2 . When including all analyses, the *Eucalyptus* I^2_{Total} was 98.59% and τ^2 was 0.27, but eliminating analyses with ratings of “unpublishable” reduced I^2_{Total} to 79.74% and τ^2 to 0.01, and removing also those analyses “needing major revisions” left I^2_{Total} at 88.91% and τ^2 at 0.03 (Table 2). Additionally, the allocations of I^2 to the team versus individual effect were altered for both blue tit and *Eucalyptus* meta-analyses by removing poorly rated analyses, but in different ways. For blue tit meta-analysis, between a third and two-thirds of the total I^2 was attributable to among-team variance in most analyses until both analyses rated “unpublishable” and analyses rated in need of “major revision” were eliminated, in which case almost all remaining heterogeneity was attributable to among-effect differences. In contrast, for *Eucalyptus* meta-analysis, the among-team component of I^2 was less than third until both analyses rated “unpublishable” and analyses rated in need of “major revision” were eliminated, in which case almost 90% of heterogeneity was attributable to differences among teams.

Out-of-sample predictions (yi)

We did not conduct these post hoc analyses on the out-of-sample predictions as the number of eligible effects was smaller and our ability to interpret heterogeneity values for these analyses was limited

Post hoc analysis: exploring the effect of including only analyses conducted by analysis teams with at least one member self-rated as “highly proficient” or “expert” in conducting statistical analyses in their research area

Effect sizes (Z_r)

Including only analyses conducted by teams that contained at least one member who rated themselves as “highly proficient” or “expert” in conducting the relevant statistical methods had negligible impacts on the meta-analytic means (Table 4), the distribution of Z_r effects (Supplementary Material B, Figure B.4), or heterogeneity estimates (Table 2), which remained extremely high.

Out-of-sample predictions (yi)

We did not conduct these post hoc analyses on the out-of-sample predictions as the number of eligible effects was smaller.

Post hoc analysis: exploring the effect of excluding estimates of Z_r in which we had reduced confidence

As described in our addendum to the methods, we identified a subset of estimates of Z_r in which we had less confidence because of features of the submitted degrees of freedom. Excluding these effects in which we had lower confidence had minimal impact on the meta-analytic mean and the estimates of total I^2 and τ^2 for both blue tit and *Eucalyptus* meta-analyses, regardless of whether outliers were also excluded (Supplementary Material B, Table B. 1).

Post hoc analysis: exploring the effect of excluding effects from blue tit models that contained two highly collinear predictors

Effect sizes (Z_r)

Excluding effects from blue tit models that contained the two highly collinear predictors (brood count after manipulation and brood count at day 14) had negligible impacts on the meta-analytic means (Table 4), the distribution of Z_r effects (Supplementary Material B, Figure B.5), or heterogeneity estimates (Table 2), which remained high.

Out-of-sample predictions

Inclusion of collinear predictors does not harm model prediction, and so we did not conduct these post hoc analyses.

Explaining variation in deviation scores

None of the pre-registered predictors explained substantial variation in deviation among submitted statistical effects from the meta-analytic mean (Tables 5 and 6).

Deviation scores as explained by reviewer ratings

Effect sizes (Z_r)

We obtained reviews from 153 reviewers who reviewed analyses for a mean of 3.27 (range 1–11) analysis teams. Analyses of the blue tit dataset received a total of 240 reviews, each was reviewed by a mean of 3.87 (SD 0.71, range 3–5) reviewers. Analyses of the *Eucalyptus* dataset received a total of 178 reviews, each was reviewed by a mean of 4.24 (SD 0.79, range 3–6) reviewers. We tested for inter-rater-reliability (IRR) to examine how similarly reviewers reviewed each analysis and found approximately no agreement among reviewers. When

considering continuous ratings, IRR was 0.01, and for categorical ratings, IRR was -0.14 .

Many of the models of deviation as a function of peer ratings faced issues of failure to converge or singularity due to sparse design matrices with our pre-registered random effects (Effect ID and Reviewer ID) (see Supplementary Material C). These issues persisted after increasing the tolerance and changing the optimizer. For both *Eucalyptus* and blue tit datasets, models with continuous ratings as a predictor were singular when both pre-registered random effects were included.

When using both categorical and continuous ratings as predictors, only models converged and allowed 95% confidence intervals to be calculated when specifying Reviewer ID as a random effect. The categorical ratings model had a R^2_C of 0.09 and a R^2_M of 0.01, the continuous ratings model had a R^2_C of 0.09 and a R^2_M of 0.01 for the blue tit dataset and a R^2_C of 0.12 and a R^2_M of 0.01 for the *Eucalyptus* dataset. Neither continuous nor categorical reviewer ratings of the analyses meaningfully predicted deviance from the meta-analytic mean (Table 6, Fig. 4). We re-ran the multi-level meta-analysis with a fixed effect for the categorical publishability ratings and found no difference in mean standardized effect sizes among publishability ratings (Supplementary Material B, Figure B.1).

Out-of-sample predictions (y_i)

Some models of the influence of reviewer ratings on out-of-sample predictions (y_i) had issues with convergence and singularity of fit (see Supplementary Material C, Table C.3) and those models that converged and were not singular showed no strong relationship (Supplementary Material C, Figure C.2, Supplementary Material C, Figure C.3), as with the Z_r analyses.

Deviation scores as explained by the distinctiveness of variables in each analysis

Effect sizes (Z_r)

We employed Sorensen's index to calculate the distinctiveness of the set of predictor variables used in each model (Fig. 5). The mean Sorensen's score for blue tit analyses was 0.59 (SD 0.1, range 0.43–0.86), and for *Eucalyptus* analyses was 0.69 (SD 0.08, range 0.55–0.98).

We found no meaningful relationship between distinctiveness of variables selected and deviation from the meta-analytic mean (Table 6, Fig. 5) for either blue tit (mean 0.42, 95%CI $-0.49, 1.34$) or *Eucalyptus* effects (mean 0.18, 95%CI $-2.78, 3.14$).

Out-of-sample predictions (y_i)

As with the Z_r estimates, we did not observe any convincing relationships between deviation scores of

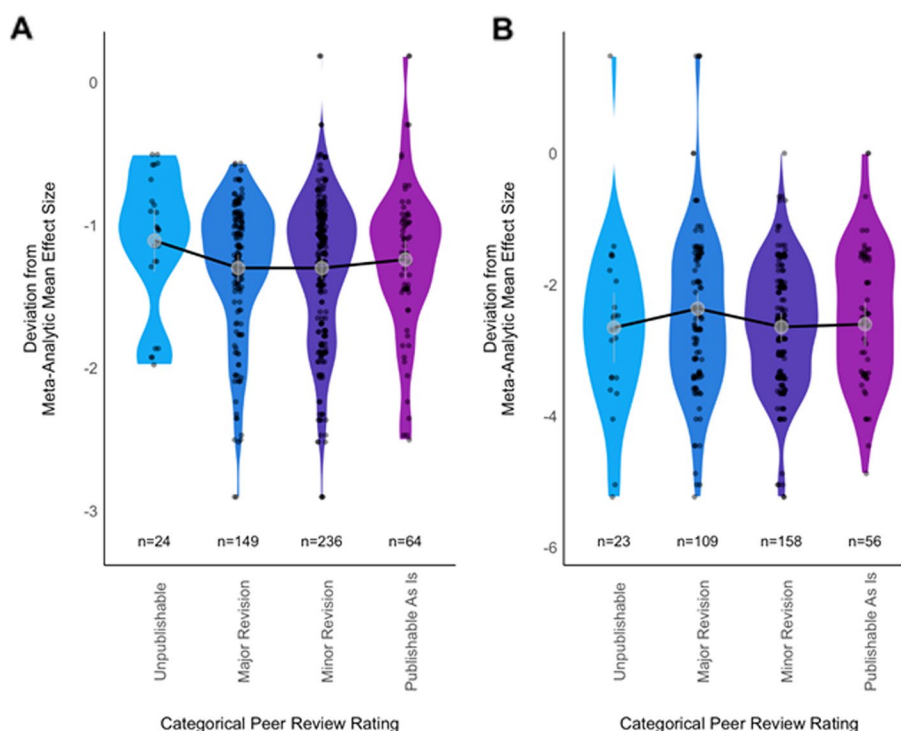


Fig. 4 Violin plot of Box-Cox transformed deviation from meta-analytic mean \bar{Z}_r as a function of categorical peer rating. Grey points for each rating group denote model-estimated marginal mean deviation, and error bars denote 95%CI of the estimate. **A** Blue tit dataset. **B** *Eucalyptus* dataset

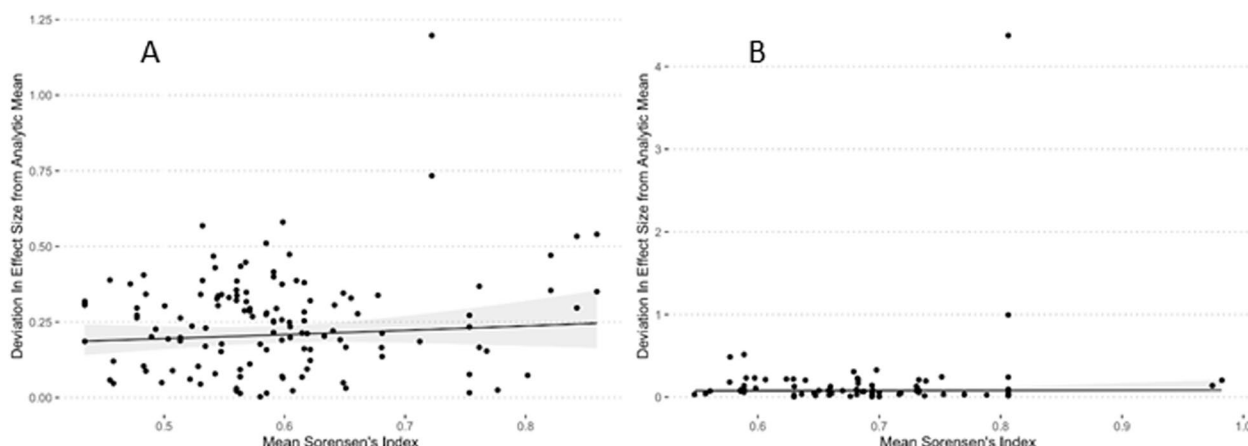


Fig. 5 Fitted model of the Box-Cox-transformed deviation score (deviation in effect size from meta-analytic mean) as a function of the mean Sorensen's index showing distinctiveness of the set of predictor variables. Grey ribbons on predicted values are 95% CIs. **A** blue tit dataset. **B** *Eucalyptus* dataset

out-of-sample predictions and Sorensen's index values (see Supplementary Material C4.1).

inclusion and deviation from meta-analytic mean among the *Eucalyptus* analyses (Tables 6, Fig. 6).

Deviation scores as explained by the inclusion of random effects

Effect sizes (Z_r)

There were only three blue tit analyses that did not include random effects, which is below the pre-registered threshold for fitting a model of the Box-Cox transformed deviation from the meta-analytic mean as a function of whether the analysis included random-effects. However, 17 *Eucalyptus* analyses included only fixed effects, which crossed our pre-registered threshold. Consequently, we performed this analysis for the *Eucalyptus* dataset only. There was no relationship between random-effect

Out-of-sample predictions (y_i)

As with the Z_r estimates, we did not examine the possibility of a relationship between the inclusion of random effects and the deviation scores of the blue tit out-of-sample predictions. When we examined the possibility of this relationship for the *Eucalyptus* effects, we found consistent evidence of somewhat higher Box-Cox-transformed deviation values for models including a random effect, meaning the models including random effects averaged slightly higher deviation from the meta-analytic means (Supplementary Material C, Figure C.5).

Table 5 Summary metrics for registered models seeking to explain deviation (Box-Cox transformed absolute deviation scores) from \bar{Z}_r as a function of Sorensen's Index, categorical peer ratings, and continuous peer ratings for blue tit and *Eucalyptus* analyses, and as a function of the presence or absence of random effects (in the analyst's models) for *Eucalyptus* analyses. We report coefficient of determination, R^2 , for our models including only fixed effects as predictors of deviation, and we report $R^2_{\text{Conditional}}$, R^2_{Marginal} and the intra-class correlation (ICC) from our models that included both fixed and random effects. For all our models, we calculated the residual standard deviation σ and root mean squared error (RMSE)

Dataset	NObs	R^2	$R^2_{\text{Conditional}}$	R^2_{Marginal}	ICC	σ	RMSE
Deviation explained by categorical ratings							
<i>Eucalyptus</i>	346		0.13	0.01	0.12	1.06	1.02
blue tit	473		0.09	7.47×10^{-3}	0.08	0.5	0.48
Deviation explained by continuous ratings							
<i>Eucalyptus</i>	346		0.12	7.44×10^{-3}	0.11	1.06	1.03
blue tit	473		0.09	3.44×10^{-3}	0.09	0.5	0.48
Deviation explained by Sorensen's index							
<i>Eucalyptus</i>	79	1.84×10^{-4}				1.12	1.1
blue tit	131	6.32×10^{-3}				0.51	0.51
Deviation explained by inclusion of random effects							
<i>Eucalyptus</i>	79	8.75×10^{-8}				1.12	1.1

Table 6 Parameter estimates from models of Box-Cox transformed deviation scores from \bar{Z}_r as a function of continuous and categorical peer ratings, Sorensen scores, and the inclusion of random effects. Standard errors (SE) and 95% confidence intervals (95% CI) are reported for all estimates, while t values, degrees of freedom and p -values are presented for fixed-effects. Note that positive parameter estimates mean that as the predictor variable increases, so does the absolute value of the deviation from the meta-analytic mean

Parameter	Random effect	Coefficient	SE	95% CI	t	df	p
Deviation explained by inclusion of random effects - <i>Eucalyptus</i>							
(Intercept)		-2.53	0.27	[-3.06, -1.99]	-9.31	77	<0.001
Mixed model		0.00	0.31	[-0.60, 0.60]	0.00	77	>0.9
Deviation explained by Sorensen's index - <i>Eucalyptus</i>							
(Intercept)		-2.65	1.05	[-4.70, -0.60]	-2.53	77	0.011
Mean Sorensen's index		0.18	1.51	[-2.78, 3.14]	0.12	77	>0.9
Deviation explained by Sorensen's index - blue tit							
(Intercept)		-1.53	0.28	[-2.08, -0.98]	-5.42	129	<0.001
Mean Sorensen's index		0.42	0.47	[-0.49, 1.34]	0.91	129	0.4
Deviation explained by continuous ratings - <i>Eucalyptus</i>							
(Intercept)		-2.23	0.23	[-2.69, -1.78]	-9.65	342	<0.001
RateAnalysis		-0.004	0	[-0.011, 0]	-1.44	342	0.15
SD(Intercept)	Reviewer ID	0.37	0.09	[0.24, 0.60]			
SD(Observations)	Residual	1.06	0.04	[0.98, 1.15]			
Deviation explained by continuous ratings - blue tit							
(Intercept)		-1.16	0.11	[-1.37, -0.94]	-10.60	469	<0.001
RateAnalysis		-0.002	0	[-0.004, 0]	-1.22	469	0.2
SD(Intercept)	Reviewer ID	0.16	0.03	[0.10, 0.24]			
SD(Observations)	Residual	0.5	0.02	[0.46, 0.53]			
Deviation explained by categorical ratings - <i>Eucalyptus</i>							
(Intercept)		-2.66	0.27	[-3.18, -2.13]	-9.97	340	<0.001
Publishable with major revision		0.29	0.29	[-0.27, 0.85]	1.02	340	0.3
Publishable with minor revision		0.01	0.28	[-0.54, 0.56]	0.04	340	>0.9
Publishable as is		0.05	0.31	[-0.55, 0.66]	0.17	340	0.9
SD(Intercept)	Reviewer ID	0.39	0.09	[0.25, 0.61]			
SD(Observations)	Residual	1.06	0.04	[0.98, 1.15]			
Deviation explained by categorical ratings - blue tit							
(Intercept)		-1.11	0.11	[-1.33, -0.89]	-9.91	467	<0.001
Publishable with major revision		-0.19	0.12	[-0.42, 0.04]	-1.62	467	0.10
Publishable with minor revision		-0.19	0.12	[-0.42, 0.04]	-1.65	467	0.10
Publishable as is		-0.13	0.13	[-0.39, 0.12]	-1.02	467	0.3
SD(Intercept)	Reviewer ID	0.15	0.04	[0.10, 0.24]			
SD(Observations)	Residual	0.5	0.02	[0.46, 0.53]			

Multivariate analysis effect size (Z_r) and out-of-sample predictions (y_i)

Like the univariate models, the multivariate models did a poor job of explaining deviations from the meta-analytic mean. Because we pre-registered a multivariate model that contained collinear predictors that produce results which are not readily interpretable, we present these models in the supplement. We also had difficulty with convergence and singularity for multivariate

models of out-of-sample (y_i) result and had to adjust which random effects we included (Supplementary Material C, Table C.8). However, no multivariate analyses of *Eucalyptus* out-of-sample results avoided problems of convergence or singularity, no matter which random effects we included (Supplementary Material C, Table C.8). We therefore present no multivariate *Eucalyptus* y_i models. We present parameter estimates from multivariate Z_r models for both datasets (Supplementary

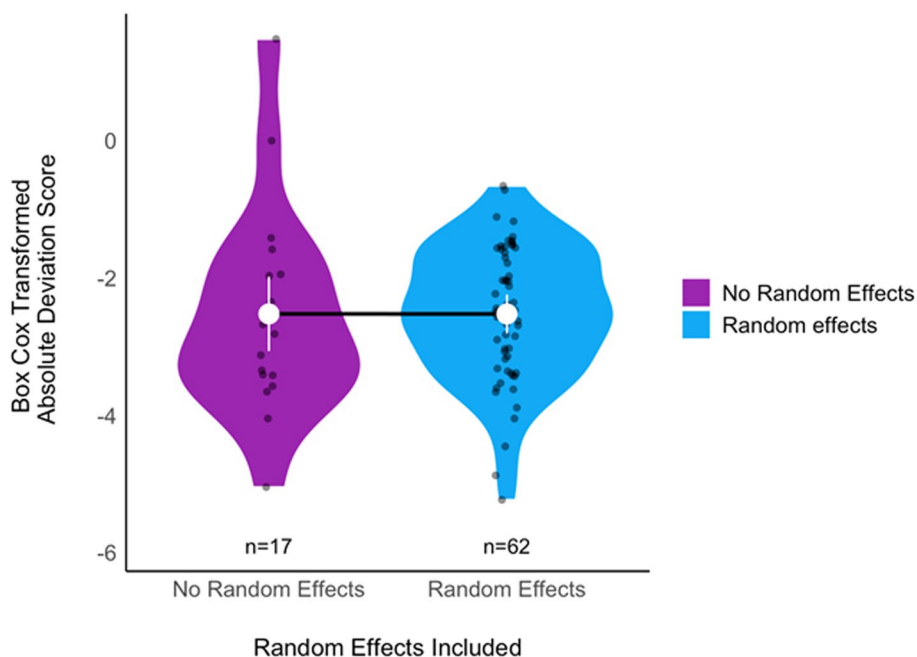


Fig. 6 Violin plot of mean Box-Cox transformed deviation from meta-analytic mean as a function of random-effects inclusion in *Eucalyptus* analyses. White point for each group of analyses denotes model-estimated marginal mean deviation, and error bars denote 95% CI of the estimate

Material C, Table 6, Table 7) and from y_i models from the blue tit dataset (Supplementary Material C, Table C.10, Table C.9). We include interpretation of the results from these models in the supplement, but the results do not change the interpretations we present above based on the univariate analyses.

Discussion

When a large pool of ecologists and evolutionary biologists analyzed the same two datasets to answer the corresponding two research questions, they produced substantially heterogeneous sets of answers. Although the variability in analytical outcomes was high for both datasets, the patterns of this variability differed distinctly between them. For the blue tit dataset, there was nearly continuous variability across a wide range of Z_r values. In contrast, for the *Eucalyptus* dataset, there was less variability across most of the range, but more striking outliers at the tails. Among out-of-sample predictions, there was again almost continuous variation across a wide range (2 SD) among blue tit estimates. For *Eucalyptus*, out-of-sample predictions were also notably variable, with about half the predicted stem count values at <2 but the other half being much larger, and ranging to nearly

40 stems per 15 m × 15 m plot. We investigated several hypotheses for drivers of this variability within datasets, but found little support for any of these. Most notably, even when we excluded analyses that had received one or more poor peer reviews, the heterogeneity in results largely persisted. Regardless of what drives the variability, the existence of such dramatically heterogeneous results when ecologists and evolutionary biologists seek to answer the same questions with the same data should trigger conversations about how ecologists and evolutionary biologists analyze data and interpret the results of their own analyses and those of others in the literature [8, 16, 96, 100].

Our observation of substantial heterogeneity due to analytical decisions is consistent with a small earlier study in ecology [103] and a growing body of work from the quantitative social sciences [15, 16, 23, 44, 92, 96]. In these studies, when volunteers from the discipline analyzed the same data, they produced a worryingly diverse set of answers to a pre-set question. This diversity included a wide range of effect sizes, and in most cases, even involved effects in opposite directions. Thus, our result should not be viewed as an anomalous outcome from two particular datasets, but instead as evidence

from additional disciplines regarding the heterogeneity that can emerge from analyses of complex datasets to answer questions in probabilistic science. Not only is our major observation consistent with other studies, it is, itself, robust because it derived primarily from simple forest plots that we produced based on a small set of decisions that were mostly registered before data gathering and which conform to widely accepted meta-analytic practices.

Unlike the strong pattern we observed in the forest plots, our other analyses, both registered and post hoc, produced either inconsistent patterns, weak patterns, or the absence of patterns. Our registered analyses found that deviations from the meta-analytic mean by individual effect sizes (\bar{Z}_r) or the predicted values of the dependent variable (\bar{y}) were poorly explained by our hypothesized predictors: peer rating of each analysis team's method section, a measurement of the distinctiveness of the set of predictor variables included in each analysis, or whether the model included random effects. However, in our post hoc analyses, we found that dropping analyses identified as unpublishable or in need of major revision by at least one reviewer modestly reduced the observed heterogeneity among the Z_r outcomes, but only for *Eucalyptus* analyses, apparently because this led to the dropping of the major outlier. This limited role for peer review in explaining the variability in our results should be interpreted cautiously because the inter-rater reliability among peer reviewers was extremely low, and at least some analyses that appeared flawed to us were not marked as flawed by reviewers. Thus, it seems that the peer reviews we received were of mixed quality, possibly due to lack of expertise or lack of care on the part of some reviewers. However, the hypothesis that poor quality analyses drove a substantial portion of the heterogeneity we observed was also contradicted by our observation that analysts' self-declared statistical expertise appeared unrelated to heterogeneity. When we retained only analyses from teams including at least one member with high self-declared levels of expertise, heterogeneity among effect sizes remained high. Thus, our results suggest lack of statistical expertise is not the primary factor responsible for the heterogeneity we observed, although further work is merited before rejecting a role for statistical expertise. Besides variability in expertise, it is also possible that the volunteer analysts varied in the effort they invested, and low effort presumably drove at least some heterogeneity in results. However, analysts often submitted thoughtful and extensive code, tables, figures, and textual explanation and interpretations, which is evidence of substantial investment. Further, we are

confident that low effort alone is an insufficient explanation for the heterogeneity we observed because we have worked with these datasets ourselves, and we know from experience that there are countless plausible modelling alternatives that can produce a diversity of effects. Additionally, heterogeneity in analytical outcomes differed notably between datasets, and there is no reason to expect that one set of analysts took this project less seriously than the other. Returning to our exploratory analyses, not surprisingly, simply dropping outlier values of Z_r for *Eucalyptus* analyses, which had more extreme outliers, led to less observable heterogeneity in the forest plots, and also reductions in our quantitative measures of heterogeneity. We did not observe a similar effect in the blue tit dataset because that dataset had outliers that were much less extreme and instead had more variability across the core of the distribution.

Our major observations raise two broad questions; why was the variability among results so high, and why did the pattern of variability differ between our two datasets. One important and plausible answer to the first question is that much of the heterogeneity derives from the lack of a precise relationship between the two biological research questions we posed and the data we provided. This lack of a precise relationship between data and question creates many opportunities for different model specifications, and so may inevitably lead to varied analytical outcomes [8]. However, we believe that the research questions we posed are consistent with the kinds of research question that ecologists and evolutionary biologists typically work from. When designing the two biological research questions, we deliberately sought to represent the level of specificity we typically see in these disciplines. This level of specificity is evident when we look at the research questions posed by some recent meta-analyses in these fields:

- “how [does] urbanization impact mean phenotypic values and phenotypic variation ... [in] paired urban and non-urban comparisons of avian life-history traits” [22]
- “[what are] the effects of ocean acidification on the crustacean exoskeleton, assessing both exoskeletal ion content (calcium and magnesium) and functional properties (biomechanical resistance and cuticle thickness)” [95]
- “[what is] the extent to which restoration affects both the mean and variability of biodiversity outcomes ... [in] terrestrial restoration” [7]
- “[does] drought stress [have] a negative, positive, or null effect on aphid fitness” [58]

- “[what is] the influence of nitrogen-fixing trees on soil nitrous oxide emissions” [53]

There is not a single precise answer to any of these questions, nor to the questions we posed to analysts in our study. And this lack of single clear answers will obviously continue to cause uncertainty since ecologists and evolutionary biologists conceive of the different answers from the different statistical models as all being answers to the same general question. A possible response would be a call to avoid these general questions in favor of much more precise alternatives [8]. However, the research community rewards researchers who pose broad questions [98], and so researchers are unlikely to narrow their scope without a change in incentives. Further, we suspect that even if individual studies specified narrow research questions, other scientists would group these more narrow questions into broader categories, for instance in meta-analyses, because it is these broader and more general questions that often interest the research community.

Although variability in statistical outcomes among analysts may be inevitable, our results raise questions about why this variability differed between our two datasets. We are particularly interested in the differences in the distribution of Z_r , since the distributions of out-of-sample predictions were on different scales for the two datasets, thus limiting the value of comparisons. The forest plots of Z_r from our two datasets showed distinct patterns, and these differences are consistent with several alternative hypotheses. The results submitted by analysts of the *Eucalyptus* dataset showed a small average (close to zero) with most estimates also close to zero (± 0.2), though about a third far enough above or below zero to cross the traditional threshold of statistical significance. There were a small number of striking outliers that were very far from zero. In contrast, the results submitted by analysts of the blue tit dataset showed an average much further from zero (-0.35) and a much greater spread in the core distribution of estimates across the range of Z_r , values (± 0.5 from the mean), with few modest outliers. So, why was there more spread in effect sizes (across the estimates that are not outliers) in the blue tit analyses relative to the *Eucalyptus* analyses?

One possible explanation for the lower heterogeneity among most *Eucalyptus* Z_r effects is that weak relationships may limit the opportunities for heterogeneity in analytical outcome. Some evidence for this idea comes from two sets of “many labs” studies in psychology [49, 50]. In these studies, many independent

lab groups each replicated a large set of studies, including, for each study, the experiment, data collection, and statistical analyses. These studies showed that, when the meta-analytic mean across the replications from different labs was small, there was much less heterogeneity among the outcomes than when the mean effect sizes were large [49, 50]. Of course, a weak average effect size would not prevent divergent effects in all circumstances. As we saw with the *Eucalyptus* analyses, taking a radically smaller subset of the data can lead to dramatically divergent effect sizes even when the mean with the full dataset is close to zero.

Our observation that dramatic sub-setting in the *Eucalyptus* dataset was associated with correspondingly dramatic divergence in effect sizes leads us towards another hypothesis to explain the differences in heterogeneity between the *Eucalyptus* and blue tit analysis sets. It may be that when analysts often divide a dataset into subsets, the result will be greater heterogeneity in analytical outcome for that dataset. Although we saw sub-setting associated with dramatic outliers in the *Eucalyptus* dataset, nearly all other analyses of *Eucalyptus* data used close to the same set of 351 samples, and as we saw, these effects did not vary substantially. However, analysts often analyzed only a subset of the blue tit data, and as we observed, sample sizes were much more variable among blue tit effects, and the effects themselves were also much more variable. Important to note here is that subsets of data may differ from each other for biological reasons, but they may also differ due to sampling error. Sampling error is a function of sample size, and sub-samples are, by definition, smaller samples, and so more subject to variability in effects due to sampling error [46].

Other features of datasets are also plausible candidates for driving heterogeneity in analytical outcomes, including features of covariates. In particular, relationships between covariates and the response variable as well as relationships between covariates and the primary independent variable (collinearity) can strongly influence the modeled relationship between the independent variable of interest and the dependent variable [27, 70]. Therefore, inclusion or exclusion of these covariates can drive heterogeneity in effect sizes (Z_r). Also, as we saw with the two most extreme Z_r values from the blue tit analyses, in multivariate models with collinear predictors, extreme effects can emerge when estimating partial correlation coefficients due to high collinearity, and conclusions can differ dramatically depending on which relationship receives the researcher’s attention. Therefore, differences between datasets in the presence of strong and/

or collinear covariates could influence the differences in heterogeneity in results among those datasets.

Although it is too early in the many-analyst research program to conclude which analytical decisions or which features of datasets are the most important drivers of heterogeneity in analytical outcomes, we must still grapple with the possibility that analytical outcomes may vary substantially based on the choices we make as analysts. If we assume that, at least sometimes, different analysts will produce dramatically different statistical outcomes, what should we do as ecologists and evolutionary biologists? We review some ideas below.

The easiest path forward after learning about this analytical heterogeneity would be simply to continue with “business as usual”, where researchers report results from a small number of statistical models. A case could be made for this path based on our results. For instance, among the blue tit analyses, the precise values of the estimated Z_r effects varied substantially, but the average effect was convincingly different from zero, and a majority of individual effects (84%) were in the same direction. Arguably, many ecologists and evolutionary biologists appear primarily interested in the direction of a given effect and the corresponding p -value [30], and so the variability we observed when analyzing the blue tit dataset may not worry these researchers. Similarly, most effects from the *Eucalyptus* analyses were relatively close to zero, and about two-thirds of these effects did not cross the traditional threshold of statistical significance. Therefore, a large proportion of people analyzing these data would conclude that there was no effect, and this is consistent with what we might conclude from the meta-analysis.

However, we find the counter arguments to “business as usual” to be compelling. For blue tits, there were a substantial minority of calculated effects that would be interpreted by many biologists as indicating the absence of an effect (28%), and there were three traditionally “significant” effects in the opposite direction to the average. The qualitative conclusions of analysts also reflected substantial variability, with fully half of teams drawing a conclusion distinct from the one we draw from the distribution as a whole. These teams with different conclusions were either uncertain about the negative relationship between competition and nestling growth, or they concluded that effects were mixed or absent. For the *Eucalyptus* analyses, this issue is more concerning. Around two-thirds of effects had confidence intervals overlapping zero, and of the third of analyses with confidence intervals excluding zero, almost half were positive, and the rest

were negative. Accordingly, the qualitative conclusions of the *Eucalyptus* teams were spread across the full range of possibilities. But, as we describe in the next paragraph, even this striking lack of consensus may be much less of a problem than what could emerge as scientists select which results to publish.

A potentially larger argument against “business as usual” is that it provides the raw material for biasing the literature. When different model specifications readily lead to different results, analysts may be tempted to report the result that appears most interesting, or that is most consistent with expectation [36, 33]. There is growing evidence that researchers in ecology and evolutionary biology often report a biased subset of the results they produce [26, 48] and that this bias exaggerates the average size of effects in the published literature between 30 and 150% [83, 121]. The bias then accumulates in meta-analyses, apparently more than doubling the rate of conclusions of “statistical significance” in published meta-analyses above what would have been found in the absence of bias [121]. Thus, “business as usual” does not just create noisy results, it helps create systematically misleading results.

If we move away from “business as usual”, where do we go? Many obvious options involve multiple analyses per dataset. For instance, there is the traditional robustness or sensitivity check [17, 85], in which the researcher presents several alternative versions of an analysis to demonstrate that the result is “robust” [59]. Unfortunately, robustness checks are at risk of the same potential biases of reporting found in other studies [96], especially given the relatively few models typically presented. However, these risks could be minimized by running more models and doing so with a pre-registration or registered report. Another option is model averaging. Averages across models often perform well [105], and in some forms this may be a relatively simple solution. Model averaging, as most often practiced in ecology and evolutionary biology, involves first identifying a small suite of candidate models [20], then using Akaike weights, based on Akaike’s Information Criterion (AIC), to calculate weighted averages for parameter estimates from those models. As with typical robustness checks, the small number of models limits the exploration of specification space, but examining a larger number of models could become the norm. However, there are more concerning limitations. The largest of these limitations is that averaging regression coefficients is problematic when models differ in interaction terms or collinear variables [21]. Additionally, weighting by AIC may often be inconsistent with our

modelling goals. AIC balances the trade-off between model complexity and predictive ability, but penalizing models for complexity may not be suited for testing hypotheses about causation [4]. So, AIC may often not offer the weight we want to use, and we may also not wish to just generate an average at all. Instead, if we hope to understand an extensive universe of possible modelling outcomes, we could conduct a multiverse analysis, possibly with a specification curve [99, 100]. This could mean running hundreds or thousands of models (or more!) to examine the distribution of possible effects, and to see how different model specification choices map onto these effects. However, exploring large areas of specification space may come at the cost of including biologically implausible specifications. Thus, we expect a trade-off, and attempts to limit models to the most biologically plausible may become increasingly difficult in proportion to the number of variables and modelling choices. To make selecting plausible models easier, one could recruit multiple analysts to design one or a few plausible specifications each as with our “many analyst” study [96]. An alternative that may be more labor intensive for the primary analyst, but which may lead to a more plausible set of models, could involve hypothesizing about causal pathways with DAGs [directed acyclic graphs, Arif and MacNeil ([5])] to constrain the model set. As with other options outlined above, generating model specifications with DAGs could be partnered with pre-registration to hinder bias from undisclosed data dredging.

Responses to heterogeneity in analysis outcomes need not be limited to simply conducting more analyses, especially if it turns out that analysis quality drives some of the observed heterogeneity. As we noted above, we cannot yet rule out the possibility that insufficient statistical expertise or poor-quality analyses might drive some portion of the heterogeneity we observed. Improving the quality of analyses might be accomplished with a deliberate increase in investment in statistical education. Many ecology and evolutionary biology students learn their statistical practice informally, with many ecology doctoral programs in the USA not requiring a statistics course [107]), and no formal courses of any kind included in doctoral degrees in most other countries. In cases where formal investment in statistical education is lacking, informal resources, such as guidelines and checklists, may help researchers avoid common mistakes. However, unless following guidelines or checklists is enforced for publication, the adherence to guidelines is patchy. For example, despite the publication of guidelines for conducting

meta-analyses in ecology, the quality of meta-analyses did not improve substantially over time [52]. Even in medical research where adherence to guidelines such as the PRISMA standards for systematic reviews and meta-analyses is more highly valued, adherence is often poor [81].

Although we have reviewed a variety of potential responses to the existence of variability in analytical outcomes, we certainly do not wish to imply that this is a comprehensive set of possible responses. Nor do we wish to imply that the opinions we have expressed about these options are correct. Determining how the disciplines of ecology and evolutionary biology should respond to knowledge of the variability in analytical outcome will benefit from the contribution and discussion of ideas from across these disciplines. We look forward to learning from these discussions and to seeing how these disciplines ultimately respond.

Conclusions

Overall, our results suggest to us that, where there is a diverse set of plausible analysis options, no single analysis should be considered a complete or reliable answer to a research question. Further, because of the evidence that ecologists and evolutionary biologists often present a biased subset of the analyses they conduct [26, 48, 121], we do not expect that even a collection of different effect sizes from different studies will accurately represent the true distribution of effects [121]. Therefore, we believe that an increased level of skepticism of the outcomes of single analyses, or even single meta-analyses, is warranted going forward. We recognize that some researchers have long maintained a healthy level of skepticism of individual studies as part of sound and practical scientific practice, and it is possible that those researchers will be neither surprised nor concerned by our results. However, we doubt that many researchers are sufficiently aware of the potential problems of analytical flexibility to be appropriately skeptical. We hope that our work leads to conversations in ecology, evolutionary biology, and other disciplines about how best to contend with heterogeneity in results that is attributable to analytical decisions.

Appendix 1

R Package References and Session Information

Table 7 shows all R packages and their versions used in the production of the manuscript.

Table 7 R packages used to generate this manuscript. Please see the ManyEcoEvo: package for a full list of packages used in the analysis pipeline

Package	Version	Citation
base	4.4.0	R Core Team (R Core Team [87])
betapart	1.6	Baselga et al. (2023)
broom.mixed	0.2.9.5	Bolker et al. [13]
colorspace	2.1.0	Zeileis et al. [122]
cowplot	1.1.3	Wilke [118]
devtools	2.4.5	Wickham et al. (Wickham et al. [116])
EnvStats	2.8.1	Millard (Millard [68])
GGally	2.2.1	Schloerke et al. (Schloerke et al. [91])
ggforestplot	0.1.0	Scheinin et al. (Scheinin et al. [90])
ggh4x	0.2.8	van den Brand [109]
ggpubr	0.6.0	Kassambara (Kassambara [47])
ggrepel	0.9.5	Slowikowski (Slowikowski [102])
ggthemes	5.1.0	Arnold [6]
glmmTMB	1.1.8	Brooks et al. ([18])
gt	0.10.1	Iannone et al. (Iannone et al. [45])
gtsummary	1.7.2	Sjoberg et al. [101]
here	1.0.1	Müller Müller [71]
Hmisc	5.1.2	Harrell Jr [41]
irr	0.84.1	Gamer, Lemon, and Singh ([35])
janitor	2.2.0	Firke (Firke [32])
knitr	1.46	Xie (Xie [119])
latex2exp	0.9.6	Meschiari (Meschiari [66])
lme4	1.1.35.3	Bates et al. [11]
ManyEcoEvo	2.7.6	Gould et al. [38]
metafor	4.6.0	Viechtbauer [114]
modelbased	0.8.7	Makowski et al. (Makowski et al. [64])
multilevelmod	1.0.0	Kuhn and Frick (Kuhn and Frick [54])
MuMIn	1.47.5	Bartoń [9]
naniar	1.1.0	Tierney and Cook [106]
NatParksPalettes	0.2.0	Blake (Blake [12])
orchaRd	2	Nakagawa, Lagisz, et al. (2023)
parameters	0.21.7	Lüdecke et al. [60]
patchwork	1.2.0	Pedersen [84]
performance	0.11.0	Lüdecke, Ben-Shachar, et al. [61]
renv	1.0.2	Ushey and Wickham (Ushey and Wickham [108])
rmarkdown	2.27	Allaire et al. [3]
sae	1.3	Molina and Marhuenda (Molina and Marhuenda [69])
scales	1.3.0	Wickham, Pedersen, and Seidel [117]
see	0.8.4	Lüdecke, Patil, et al. [62]
showtext	0.9.7	Qiu Qiu [86]
specr	1.0.0	Masur and Scharlow (Masur and Scharlow [65])
targets	1.7.0	Landau [57]
tidymodels	1.1.1	Kuhn and Wickham [55]
tidytext	0.4.2	Silge and Robinson [97]
tidyverse	2.0.0	Wickham et al. [115]
withr	3.0.0	Hester et al. (Hester et al. [42])
xfun	0.44	Xie (Xie [120])

Appendix 2

Table 8 Session info

setting	value
version	R version 4.4.0 (2024-04-24)
os	macOS Ventura 13.6.9
system	aarch64, darwin20
ui	X11
language	(EN)
collate	en_US.UTF-8
ctype	en_US.UTF-8
tz	Australia/Melbourne
date	2024-09-17
pandoc	3.1.12.2 @ /opt/homebrew/bin/ (via rmarkdown)

Supplementary Information

The online version contains supplementary material available at <https://doi.org/10.1186/s12915-024-02101-x>.

Supplementary Material 1.

Authors' contributions

HF, THP and FF conceptualized the project. PV provided raw data for *Eucalyptus* analyses and SG and THP provided raw data for blue tit analyses. DGH, HF and THP prepared surveys for collecting participating analysts and reviewer's data. EG, HF, THP, PV, SN and FF planned the analyses of the data provided by our analysts and reviewers, EG, HF, DGH and THP curated the data, EG and HF wrote the software code to implement the analyses and prepare data visualizations. EG ensured that analyses were documented and reproducible. THP and HF administered the project, including coordinating with analysts and reviewers. FF provided funding for the project. THP, HF, and EG wrote the manuscript. Authors listed alphabetically contributed analyses of the primary datasets or reviews of analyses. All authors read and approved the final manuscript.

Funding

EG's contributions were supported by an Australian Government Research Training Program Scholarship, AIMOS top-up scholarship (2022) and Melbourne Centre of Data Science Doctoral Academy Fellowship (2021). FF's contributions were supported by ARC Future Fellowship FT150100297.

Data availability

All materials and data are archived and hosted on the OSF at <https://osf.io/mn5aj/>, including survey instruments and analyst / reviewer consent forms. The Evolutionary Ecology Data and Ecology and Conservation Data provided to analysts are available at <https://osf.io/34fzc/> and <https://osf.io/t76uy/> respectively. Data has been anonymized, and the non-anonymized data is stored on the project OSF within private components accessible to the lead authors.

We built an R package, ManyEcoEvo to conduct the analyses described in this study [38], which can be downloaded from <https://github.com/egouldo/ManyEcoEvo/> to reproduce our analyses or replicate the analyses described here using alternate datasets. Data cleaning and preparation of analysis-data, as well as the analysis, is conducted in [87] reproducibly using the targets package [57]. This data and analysis pipeline is stored in the ManyEcoEvo package repository and its outputs are made available to users of the package when the library is loaded.

The full manuscript, including further analysis and presentation of results is written in Quarto [2]. The source code to reproduce the manuscript is hosted

at <https://github.com/egouldo/ManyAnalysts/>, and the rendered version of the source code may be viewed at <https://egouldo.github.io/ManyAnalysts/>. All R packages and their versions used in the production of this manuscript are listed in Appendix 1.

Declarations

Ethics approval and consent to participate

We obtained permission to conduct this research from the Whitman College Institutional Review Board (IRB). As part of this permission, the IRB approved the consent form (<https://osf.io/xyp68/>) that all participants completed prior to joining the study. The authors declare that they have no competing interests.

Competing interests

The authors declare that they have no competing interests.

Author details

¹School of Agriculture Food and Ecosystem Sciences, University of Melbourne, Grattan Street, Parkville, Victoria 3010, Australia. ²School of Historical and Philosophical Studies, University of Melbourne, Grattan Street, Parkville, Victoria 3010, Australia. ³Department of Biology, Whitman College, 345 Boyer Ave, Walla Walla, WA 99362, USA. ⁴School of Biological, Earth & Environmental Sciences, University of New South Wales, Sydney, NSW 2052, Australia. ⁵School of Natural Sciences, Macquarie University, Balacra Rd, Macquarie Park, Sydney, NSW 2109, Australia. ⁶School of Public Health and Preventive Medicine, Monash University, 750 Collins Street, Docklands, VIC 3008, Australia. ⁷Länsstyrelsen Östergötland, Östgötagatan 3, 58186 Linköping, Sweden. ⁸Biology Department, Lund University, Sölvegatan 37, 22362 Lund, Sweden. ⁹Department of Biology, University of Massachusetts, 1 Campus Center Way, Amherst, MA 01003, USA. ¹⁰Marine and Continental Waters, IRTA, Carretera Poble Nou Km 5.5, 43540 La Ràpita, Catalonia, Spain. ¹¹Department of Life Sciences, Ben Gurion University of the Negev, P.O.Box 653, 84105 Beer Sheva, Israel. ¹²Department of Psychology, The University of Edinburgh, 7 George Square, Edinburgh EH9 1HB, UK. ¹³Centre for Ecological Sciences, Indian Institute of Science, Indian Institute of Science, Bengaluru, Karnataka 560012, India. ¹⁴Southern Research Station, USDA Forest Service, PO Box 700, New Ellenton, SC 29809, USA. ¹⁵Center for Ecological Dynamics in a Novel Biosphere (ECONOVO), Department of Biology, Aarhus University, Ny Munkegade 114-116, 8000 Aarhus C, Denmark. ¹⁶School of Mathematics and Statistics, University of Melbourne, Parkville, VIC 3052, Australia. ¹⁷Biology, Indiana University Purdue University Indianapolis, 420 University Blvd, Indianapolis, IN 46202, USA. ¹⁸School of Life and Environmental Sciences, Deakin University, 221 Burwood Highway, Burwood, VIC 3125, Australia. ¹⁹Department of Arid Land Agriculture, King Abdulaziz University, Jeddah 80200, Kingdom of Saudi Arabia. ²⁰Department of Biological Sciences, Macquarie University, 205ACR Culloden Road, Macquarie Park, New South Wales 2113, Australia. ²¹Department of Plant Ecology, University of Hohenheim, Institute of Landscape and Plant Ecology, Ottilie-Zeller-Weg, 70599 Stuttgart, Germany. ²²Department of Wildlife, Fish, and Environmental Studies, Swedish University of Agricultural Sciences, Skogsmarksgränd 17, SE-907 36 Umeå, Sweden. ²³Department of Ecology and Evolutionary Biology, University of Connecticut, 75 N. Eagleville Rd, Storrs, CT 06226, USA. ²⁴STEM Center, Southern Illinois University Edwardsville, 1 Hairpin Dr, Edwardsville, IL 62026, USA. ²⁵University of Guelph, 50 Stone Road East, Guelph, Ontario N1G 2W1, Canada. ²⁶Department of Evolutionary Biology and Environmental Studies, University of Zurich, Winterthurerstrasse 190, 8057 Zürich, Switzerland. ²⁷Department of Biological Sciences, California State Polytechnic University, Pomona, USA. ²⁸Centre d'Études Biologiques de Chizé, UMR 7372, Université de la Rochelle - Centre National de la Recherche Scientifique, 405 route de Prissé la Charrière, 79360 Villiers en Bois, France. ²⁹Faculty of Life Sciences, Bar Ilan University, Ramat Gan 529000, Israel. ³⁰School of Natural Sciences, University of Tasmania, TAS, Private Bag 55, Hobart 7001, Australia. ³¹Whitebark Institute, 3399 Main Street, Suite W5, Mammoth Lakes, CA 93546, USA. ³²Department of Biology, Rhodes College, 2000 N. Parkway, Memphis, TN 38112, USA. ³³Centre for Conservation Science, RSPB, 2 Lochside View, Edinburgh EH12 9DH, UK. ³⁴Environmental Studies, Wofford College, 429 N. Church St, Spartanburg, SC 29303, USA. ³⁵IFM Biology, Linköping University, 581 83 Linköping, Sweden. ³⁶Faculty of Environmental Sciences, Czech University of Life Sciences Prague, Czech Republic, Kamýčká 129, Praha

- Suchdol 165 00, Czech Republic. ³⁷Biological and Environmental Sciences & Gothenburg Global Biodiversity Centre, University of Gothenburg, Medicinargatan 7B, SE-413 90 Gothenburg, Sweden. ³⁸School of Biological Sciences, Monash University, Rainforest Walk 25, Clayton, Victoria, Australia. ³⁹Ecology and Evolutionary Biology, University of Tennessee Knoxville, 569 Dabney Hall, Knoxville, TN 37996, USA. ⁴⁰Departamento de Ecología e Zoología, Centro de Ciências Biológicas, Universidade Federal de Santa Catarina, UFSC, Campus Universitário - Córrego Grande Florianópolis - SC; CEP, Florianópolis 88040-900, Brazil. ⁴¹School of Biological and Forensic Sciences, University of South Wales, The Alfred Russel Wallace Building, 9 Graig Fach, Glyntaff, Pontypridd CF37 4BB, UK. ⁴²Centre for Ecology, Evolution and Environmental Changes (CE3c) & CHANGE - Global Change and Sustainability Institute, Faculdade de Ciências, Universidade de Lisboa, 1749-016 Lisbon, Portugal. ⁴³Forest and Rangeland Stewardship, Colorado State University, 1472 Campus Delivery, Fort Collins, CO 80523-1472, USA. ⁴⁴Department of Ornithology, Academy of Natural Sciences of Drexel University, 1900 Benjamin Franklin Parkway, Philadelphia, PA 19096, USA. ⁴⁵Salisbury University, 1101 Camden Ave, Biology, Salisbury, MD 21801, USA. ⁴⁶Groningen Institute for Evolutionary Life Sciences, University of Groningen, Nijenborgh 7, 9747 AG Groningen, Netherlands. ⁴⁷Department of Biology, McMaster University, 1280 Main Street West, Hamilton, ON L8S 4K1, Canada. ⁴⁸Department of Zoology, University of Cambridge, Downing St, Cambridge CB2 3EJ, UK. ⁴⁹Entomology and Nematology, University of Florida, 700 Experiment Station Rd, Lake Alfred, FL 33850, USA. ⁵⁰Environmental Studies, Elon University, McMichael Science Building, 2625 Campus Box, Elon, NC 27244, USA. ⁵¹BirdLife International, David Attenborough Building, Pembroke Street, Cambridge CB2 3QZ, UK. ⁵²School of Integrative Biological and Chemical Sciences, The University of Texas Rio Grande Valley, One West University Boulevard, Brownsville, TX 78520, USA. ⁵³Department of Ecology and Evolutionary Biology, University of Toronto, 25 Willcocks St, Toronto, ON M5S 3B2, Canada. ⁵⁴Department of Statistics, University of Auckland, Auckland, New Zealand. ⁵⁵LLC, Catbird Stats, PO Box 2018, Gautier, MS 39553, USA. ⁵⁶School of Biodiversity, One Health & Veterinary Medicine, University of Glasgow, University Avenue, Glasgow G12 8QQ, UK. ⁵⁷School of Forestry, Northern Arizona University, 200 E Pine Knoll Dr, Flagstaff, AZ 86001, USA. ⁵⁸Department of Behavioural Neurobiology, Max Planck Institute for Biological Intelligence, Eberhard-Gwinner-Strasse, 82319 Seewiesen, Oberbayern, Germany. ⁵⁹School of Sciences: Center for Health and Cognition, Bath Spa University, Newton Park, Bath BA2 9BN, UK. ⁶⁰Department of Biological Sciences, University of Alberta, Edmonton, AB T6G 2R3, Canada. ⁶¹Applied Zoology, Zellescher Weg 20b, 01217 Dresden, TU Dresden, Germany. ⁶²School of Molecular Biosciences, College of Medical Veterinary & Life Sciences, University of Glasgow, Glasgow G12 8QQ, UK. ⁶³School of Life and Environmental Sciences, The University of Sydney, Camperdown, NSW 2006, Australia. ⁶⁴Institute of Environmental Sciences, Jagiellonian University, Gronostajowa 7, 30-387 Krakow, Poland. ⁶⁵Biology Department, Brigham Young University, 4102 Life Science Building, Provo, UT, USA. ⁶⁶Institute of Evolutionary Sciences Montpellier, University of Montpellier, CNRS, IRD, Montpellier, France. ⁶⁷Baruch Marine Field Laboratory, University of South Carolina, 2306 Crabhaul Rd, Georgetown, SC 29440, USA. ⁶⁸Department of Life Sciences, Imperial College London, Buckhurst road, Berkshire SL5 7PY, UK. ⁶⁹European Forest Institute, Platz d. Vereinten Nationen 7, 53113 Bonn, Germany. ⁷⁰Department of Ornithology, Max Planck Institute for Biological Intelligence, Eberhard-Gwinner Str. 7, 82319 Seewiesen, Germany. ⁷¹Forestry and Environmental Conservation, National Bobwhite and Grassland Initiative, Clemson University, 243 Lehotsky Hall, Clemson, SC 29634, USA. ⁷²Department of Psychology and Vision Science, University of Birmingham, 52 Pritchatts Road, Edgbaston, Baily Thomas Grant Birmingham B15 2TT, UK. ⁷³School of Biology and Environmental Science, University College Dublin, Dublin 4, Belfield D04 V1W8, Ireland. ⁷⁴Department of Biological Sciences, San José State University, 129 S 10th Street, San Jose, CA 95112, USA. ⁷⁵Apicultural State Institute, University of Hohenheim, Erna-Hruschka-Weg 6, 70599 Stuttgart, Germany. ⁷⁶Department of Biology, St. Norbert College, 100 Grant St, De Pere, WI 54115, USA. ⁷⁷Department of Biological Sciences, Purdue University, 915 W. State Street, West Lafayette, IN 47907, USA. ⁷⁸Biodiversity, Faculty of Forest Sciences and Forest Ecology, University of Göttingen, Macroecology & Biogeography Büsgenweg 1, 37077 Göttingen, Germany. ⁷⁹Department of Fisheries, Wildlife and Conservation Biology, University of Minnesota-Twin Cities, 135 Skok Hall, 2003 Upper Buford Circle, St. Paul, MN 55108, USA. ⁸⁰CABI, Bakeham Lane, Egham, Surrey, UK. ⁸¹Department of Ecology and Evolutionary Biology, School of Biological Sciences, University of California, 321 Steinhaus Hall, Irvinerlrvine, CA 92697, USA. ⁸²School of Biological Sciences, University of Aberdeen, King Street, Aberdeen AB244FX, UK. ⁸³Department of Natural Resource Sciences, McGill University, 2111 Lakeshore Rd, Ste Anne-de-Bellevue, Montreal, QC H9X 3V9, Canada. ⁸⁴Institute of Ecology and Environmental Sciences (IEES), Univ. Paris-Est Creteil, 61 avenue du Général de Gaulle, 94010 Créteil, France. ⁸⁵Department of Forest Ecosystems and Society, Oregon State University, 321 Richardson Hall, Corvallis, OR 97331, USA. ⁸⁶Wake Forest University, 1834 Wake Forest Road, Winston Salem, NC 27109, USA. ⁸⁷Laboratorio de Invasiones Biológicas (LIB), Instituto de Ecología y Biodiversidad, Victoria 631, Concepción, Chile. ⁸⁸Institute for Biodiversity, Animal Health and Comparative Medicine, University of Glasgow, Glasgow G12 8QQ, UK. ⁸⁹Department of Forest Ecosystems and Society, College of Forestry, Oregon State University, Corvallis, OR 97333, USA. ⁹⁰Biodiversity and Conservation Area, Rey Juan Carlos University, C/ Tulipán s/n, 28933 Móstoles, Madrid, Spain. ⁹¹Graduate School of Biomedical Sciences, Tufts University, 136 Harrison Ave #813, Boston, MA 02111, USA. ⁹²Department of Integrative Biology, University of Guelph, 50 Stone Rd E, Guelph, ON N1G 2W1, Canada. ⁹³School of Life and Environmental Sciences (Burwood Campus), Deakin University, Geelong, Victoria, Australia. ⁹⁴CNRS, University of Rennes, 263 Avenue du Général Leclerc, 35042 Rennes, France. ⁹⁵Department of Behavioural Ecology, Bielefeld University, Konsequenz 45, 33615 Bielefeld, Germany. ⁹⁶Fakultät für Biologie, Arbeitsgruppe Evolutionsbiologie, Universität Bielefeld, Morgenbreede 45, 33615 Bielefeld, Germany. ⁹⁷Chair of Meteorology, Institute for Hydrology and Meteorology, Faculty of Environmental Sciences, Technische Universität Dresden, Piennner Str. 23, 01737 Tharandt, Germany. ⁹⁸Department of Wildlife, Fisheries, and Conservation Biology, University of Maine, 5755 Nutting Hall, Room 210, Orono, ME 04469-5755, USA. ⁹⁹Department of Biological Sciences, Boise State University, 1910 W University Dr, Boise, ID 83725, USA. ¹⁰⁰School of Agriculture, Food and Ecosystem Sciences, The University of Melbourne, Grattan Street, Parkville, Victoria 3010, Australia. ¹⁰¹Pastures Systems and Watershed Management Research Unit, USDA Agricultural Research Service, USDA-ARS PSWMRU, Bldg. 3702 Curtin Road, University Park, PA 16802, USA. ¹⁰²Department of Natural Sciences, Baruch College, City University of New York, 17 Lexington Avenue, New York, NY 10010, USA. ¹⁰³Department of Biological Sciences, Binghamton University, 4400 Vestal Parkway East, Binghamton, NY 13902, USA. ¹⁰⁴Dipartimento di Biologia, Università di Roma "Tor Vergata", Via Cracovia, 1, 00133 Rome, Italy. ¹⁰⁵Department of Anthropology, University of Michigan, 1085 S. University Ave, Ann Arbor, MI 48109, USA. ¹⁰⁶College of Forestry, Oregon State University, 3100 SW Jefferson Way, Corvallis, OR 97333, USA. ¹⁰⁷University of Antwerp, Universiteitsplein 1, 2610 Wilrijk, België, Belgium. ¹⁰⁸Earth & Environmental Sciences, Wesleyan University, 45 Wyllys Ave, Middletown, CT 06459, USA. ¹⁰⁹Department of Psychiatry, Yale School of Medicine, Yale University, 389 Whitney Ave, New Haven, CT 06511, USA. ¹¹⁰Biology Department and Environmental Studies Department, St. Olaf College, 1520 St Olaf Ave, Northfield, MN 55057, USA. ¹¹¹Department of Plant Protection, Faculty of Agriculture, Department of Plant Protection, Faculty of Agriculture, Ordu University, Ordu University, 52200 Altinordu/Ordu, Turkey. ¹¹²Department of Earth and Environmental Sciences, KU Leuven, Celestijnenlaan 200E, 3001 Leuven, Belgium. ¹¹³Department of Marine Sciences, University of Gothenburg, Box 461, SE-40530 Gothenburg, Sweden. ¹¹⁴Department of Biology, Wilfrid Laurier University, 75 University Ave West, Waterloo, Ontario N2L 3C5, Canada. ¹¹⁵Natural Resource Ecology and Management, Iowa State University, 2310 Pammel Dr, Ames, IA 50011, USA. ¹¹⁶School of Biological Sciences, University of Western Australia, 35 Stirling Highway, Crawley, Western Australia 6009, Australia. ¹¹⁷Department of Biological Sciences, Middle East Technical University, Üniversiteler Mahallesi, Dumlupınar Bulvarı No: 1, 06800 Çankaya/Ankara, Turkey. ¹¹⁸Dept. of Integrative Biology, University of Texas at Austin, 2415 Speedway #C0930, Austin, TX, USA. ¹¹⁹Grand Bay National Estuarine Research Reserve, 6005 Bayou Heron Rd, Moss Point, MS 39562, USA. ¹²⁰Universidad de los Andes, Carrera 1 # 18A-12, Bogotá, Colombia. ¹²¹Department of Biology, University of Turku, Turun Yliopisto, FI-20014 Turku, Finland. ¹²²Instituto de Ciencia Animal, Facultad de Ciencias Veterinarias, Universidad Austral de Chile, Campus Isla Teja s/n, Valdivia, Chile. ¹²³Department of Integrative Biology, The University of Texas at Austin, 2415 Speedway #C0930, Austin, Texas 78712, USA. ¹²⁴Department of Botany, University of Wyoming, Laramie, WY 82071, USA. ¹²⁵Department of Animal Breeding and Genetics, Swedish University of Agricultural Sciences, Box 7023, 750 07 Uppsala, Sweden. ¹²⁶Department of Biology, Brigham Young University, Brigham Young University, Brigham Young University, Provo, UT 84602, USA. ¹²⁷School of Biological Sciences, University of Bristol, 24 Tyndall Avenue, Bristol BS8 1TQ, UK. ¹²⁸International Institute for Applied Systems Analysis (IIASA), Schlossplatz 1, A-2361 Laxenburg, Austria. ¹²⁹Department of Ecology,

Swedish University of Agricultural Sciences, Ulls Väg 16, 750 07 Uppsala, Sweden. ¹³⁰Université de Montpellier, ISEM, University of Montpellier, CNRS, EPHE, 34000 Montpellier, IRD, France. ¹³¹Department of Wildlife, Fish, and Conservation Biology, University of California, 1 Shields Ave, Davis/Davis, CA 95616, USA. ¹³²Département des Sciences biologiques, Université du Québec à Montréal, 141 Avenue du Président-Kennedy, Montréal, Québec H2X 1Y4, Canada. ¹³³Institute of Evolutionary Biology, University of Edinburgh, King's Buildings, Edinburgh EH9 3JW, UK. ¹³⁴Center for Limnology, University of Wisconsin - Madison, 680 N Park St, Madison, WI 53706, USA. ¹³⁵Center for Biodiversity and Global Change, Yale University, 165 Prospect St, New Haven, CT 06511, USA. ¹³⁶Department of Ecology, Evolution, and Behavior, University of Minnesota, Ecology Building, 1987 Upper Buford Cir, St. Paul/St Paul, MN 55108, USA. ¹³⁷Institute of Environment and Department of Biological Sciences, Florida International University, 3000 NE 151st St, North Miami, FL 33181, USA. ¹³⁸Department of Life Sciences, Aberystwyth University, Penglais, Aberystwyth SY23 3DA, UK. ¹³⁹Institut de recherche en biologie végétale, Université de Montréal, 4101, Sherbrooke St E, Montréal, Québec H1X 2B2, Canada. ¹⁴⁰Institute of Evolutionary Ecology and Conservation Genomics, Ulm University, Albert-Einstein-Allee 11, 89081 Ulm, Germany. ¹⁴¹Department of Biology, Memorial University of Newfoundland, 45 Arctic Ave, St John's NL A1C5S7, Canada. ¹⁴²Evolution & Ecology Research Centre, School of Biological, Earth & Environmental Sciences, University of New South Wales, UNSW Sydney, High Street 2052, Kensington, NSW, Australia. ¹⁴³The Nature Conservancy, 258 Main Street, Lander, WY 82520, USA. ¹⁴⁴Ecology and Evolutionary Biology, University of Arizona, 1041 E Lowell St, Tucson, AZ 85721, USA. ¹⁴⁵Biological and Environmental Sciences, University of Stirling, Cottrell Building, Stirling FK9 4LA, UK. ¹⁴⁶Department of Biology, University of Antwerp, Universiteitsplein 1, 2610 Wilrijk, Belgium. ¹⁴⁷Aquatic Ecology and Evolution Group, Limnological Institute, University of Konstanz, Mainaustraße 252, 78464 Konstanz, Germany. ¹⁴⁸Campus Cerro Largo, Universidade Federal da Fronteira Sul, Rua Jacob Haupenthal, Cerro Largo, RS, CEP 158097900-000, Brazil. ¹⁴⁹School of Psychology and Social Work, University of Hull, Cottingham Rd, Hull HU6 7RX, UK. ¹⁵⁰UMR 1224, ECOBIOP, Université de Pau et des Pays de l'Adour, 173 Route de Saint-Jean-de-Luz, 64310 Saint-Pée-sur-Nivelle, France. ¹⁵¹School of Biological & Environmental Sciences, Liverpool John Moores University, James Parsons Building, Byrom Street, Liverpool L3 3AF, UK. ¹⁵²Institute of Ecology and Evolution, University of Edinburgh, The University of Edinburgh, King's Buildings, Charlotte Auerbach Road, Edinburgh EH9 3FL, UK. ¹⁵³Phnom Penh, Cambodia. ¹⁵⁴Statistical Ecotoxicology, Bayreuth Center of Ecology and Environmental Research (BayCEER), University of Bayreuth, Universitätsstraße 30, 95440 Bayreuth, Germany. ¹⁵⁵Ecology and Environmental Science, Umeå University, Linnaeus väg 6, 907 36 Umeå, Sweden. ¹⁵⁶Molecular Ecology Group (MEG), Water Research Institute (IRSA), National Research Council of Italy (CNR), 28922 Corso Tonolli 50, Verbania, Italy. ¹⁵⁷Department of Natural History, Norwegian University of Science and Technology, Høgskoleringen 1, 7034 Trondheim, Norway. ¹⁵⁸Center for Advanced Biotechnology and Medicine, Rutgers University Robert Wood Johnson Medical School, 679 Hoes Lane West, Piscataway, NJ 08854, USA. ¹⁵⁹Departamento de Ciencias Biológicas, Universidad de los Andes, Carrera 1 No 18A - 12, 111711 Bogotá, Bogotá D. C., Colombia. ¹⁶⁰Institut de recherche sur les forêts, Université du Québec en Abitibi-Témiscamingue, 445 Boulevard de l'Université, Rouyn-Noranda, QC J9X 5E4, Canada. ¹⁶¹Université du Québec à Trois-Rivières, 3351, boulevard des Forges, Trois-Rivières (Québec) G8Z 4M3, Canada. ¹⁶²Institute of Plant Sciences, University of Bern, Altenbergrain 21, 3013 Bern, Switzerland. ¹⁶³School of Biological, Earth and Environmental Sciences, University of New South Wales, Randwick, Sydney, NSW 2052, Australia. ¹⁶⁴Whitney Laboratory for Marine Bioscience, University of Florida, 9505 N Ocean Shore Blvd, St. Augustine, Gainesville, FL 32080, USA. ¹⁶⁵Biological Sciences, Eastern Illinois University, 600 Lincoln Avenue, Charleston, IL 61920, USA. ¹⁶⁶Centre d'Investigations Clinique Plurithématique - Institut Lorrain du Coeur et des Vaisseaux, Université de Lorraine, Inserm1433 CIC-P CHRU de Nancy, bâtiment Louis Mathieu - 5, rue du Morvan - 54500, Vandoeuvre-les-nancy, France. ¹⁶⁷Evolutionary biology department, Bielefeld University, Bielefeld University, Konsequenz 45, 33615 Bielefeld, Germany. ¹⁶⁸Department of Ecology and global change, Centro de Investigaciones sobre Desertificación, Consejo Superior de Investigaciones Científicas (CIDE-CSIC/UV/GV), Carretera CV-315 km 10,7, 46113 Moncada (Valencia), Spain. ¹⁶⁹Department of Biological Sciences, University of Arkansas, 850 W. Dickson Street SCEN601, Fayetteville, AR 72701, USA. ¹⁷⁰School of the Environment, Faculty of Science, The University of Queensland, The University of Queensland, Brisbane, QLD 4072, Australia. ¹⁷¹Department of Animal Health and Production, Oyo State College of Agriculture and Technology, Igbo-Ora 201103, Oyo State, Nigeria. ¹⁷²Department of Forest & Wildlife Ecology, University of Wisconsin-Madison, 1630 Linden Drive, Madison, WI 53706, USA. ¹⁷³Organismal and Evolutionary Biology Research Programme, Faculty of Biological and Environmental Sciences, University of Helsinki, 00014 Helsinki, Finland. ¹⁷⁴South Iceland Research Centre, University of Iceland, Lindarbraut 4, 840 Laugarvatn, Iceland. ¹⁷⁵Department of Clinical Neuroscience, Karolinska Institutet, Nobels väg 9, 171 77 Stockholm, Sweden. ¹⁷⁶Animal Ecology and Tropical Biology, University of Würzburg, Biocenter-Am Hubland, 97074 Würzburg, Germany. ¹⁷⁷Hiram College, 11700 Dean St, Biology, Hiram, OH 44234, USA. ¹⁷⁸Department of Aquatic Resources, Swedish University of Agricultural Sciences, Almas allé 5, 756 51 Uppsala, Sweden. ¹⁷⁹Department of Earth Sciences, Montana State University, Bozeman, MT 59717, USA. ¹⁸⁰Department of Evolution, Ecology, and Organismal Biology, University of California, 900 University Ave, Riverside/Riverside, CA 92521, USA. ¹⁸¹Sección Ornitológia, Universidad Nacional de La Plata, Paseo del Bosque s/n, La Plata, B1900FWA Buenos Aires, Argentina. ¹⁸²Bangor University, Bangor University, Deiniol Road, Bangor LL57 2UW, UK. ¹⁸³MRC Biostatistics Unit, University of Cambridge, East Forvie Building, Forvie Site, Robinson Way, Cambridge CB2 0SR, UK. ¹⁸⁴Harvard Forest, Harvard University, 324 N Main St, Petersham, MA 01366, USA. ¹⁸⁵Departamento de Biodiversidad, Ecología y Evolución, Universidad Complutense de Madrid, C. de José Antonio Novais, 12, 28040 Madrid, Spain. ¹⁸⁶Biology Department, Technische Universität Darmstadt, Schnittspahnstraße 3, 64287 Darmstadt, Germany. ¹⁸⁷UMR 1309, ASTRE, CIRAD, Campus international de Baillarguet, 34398 Montpellier, France. ¹⁸⁸Department of Entomology, The Ohio State University, 1680 Madison Ave, Wooster, OH 44691, USA. ¹⁸⁹Department of Psychology, University of Bath, 10 West, Bath BA2 7AY, UK. ¹⁹⁰Chaire de recherche en intégrité écologique, Université du Québec à Trois-Rivières, 3351 Boul. Des Forges, Trois-Rivières, QC G8Z 4M3, Canada. ¹⁹¹Mississippi Based RESTORE Act Center of Excellence, University of Southern Mississippi, 703 E. Beach Drive, Ocean Springs, MS 39564, USA. ¹⁹²Department of Integrative Biology, University of California, Valley Life Science Building 5075, Berkeley/Berkeley, CA 94720, USA. ¹⁹³Mention Zoologie et Biodiversité Animale, Faculté des Sciences, Université d'Antananarivo, Mention Zoologie et Biodiversité Animale, Université d'Antananarivo, BP 906, 101 Antananarivo, Madagascar. ¹⁹⁴Department of Integrative Biology, Valley Life Sciences Building 3140, University of California, University of California Berkeley, Berkeley/Berkeley, CA 94720, USA. ¹⁹⁵Department of Ecology, Behavior and Evolution, University of California, San Diego, 9500 Gilman Dr, La Jolla, CA 92093, USA. ¹⁹⁶Institute for Environmental Sciences, VU Amsterdam, De Boelelaan 1111, 1081 HV Amsterdam, The Netherlands. ¹⁹⁷Department of Evolutionary Anthropology, University of Vienna, Djerassiplatz 1 (JBB), 1030 Wien, Austria. ¹⁹⁸Konrad Lorenz Institute for Ethology, University of Veterinary Medicine, Savoyenstrasse 1A, 1160 Vienna, Austria. ¹⁹⁹School of Biological Sciences, The University of Hong Kong, Pokfulam Road, Hong Kong SAR, China. ²⁰⁰Institut de biologie, Université de Neuchâtel, Emile-Argand 11, 2000 Neuchâtel, Switzerland. ²⁰¹School of Information, University of Arizona, 1103 E. 2nd St, Tucson, AZ 85721, USA. ²⁰²Center for Biological Data Science, Virginia Commonwealth University, 1000 W. Cary St, Box 842030, Richmond, VA 23284-2030, USA. ²⁰³University of Wisconsin, 1525 Observatory Dr. Madison, Madison, WI 53706, USA. ²⁰⁴School of the Environment, Yale University, 195 Prospect Street, New Haven, CT 06511, USA. ²⁰⁵Institute of the Environment, Florida International University, 3000 NE 151st St, North Miami, FL 33181, USA. ²⁰⁶Department of Biology, Missouri State University, 910 S John Q Hammons Pkwy, Springfield, MO 65897, USA. ²⁰⁷Department of Biological and Environmental Science, University of Jyväskylä, Survantie 9C, 40500 Jyväskylä, Finland. ²⁰⁸Institute for Interdisciplinary Mountain Research, OeAW (Austrian Academy of Sciences), GLORIA, Silbergasse 30/3, A-1190 Wien, Austria. ²⁰⁹Department of Natural Resources, Newe Ya'ar Research Center, Agricultural Research Organization (Volcani Institute), POB 1021, 3009500 Ramat Yishay, Israel. ²¹⁰Department of Animal Behaviour, Bielefeld University, Konsequenz 45, 33615 Bielefeld, Germany. ²¹¹Office for National Statistics, Segensworth Rd, Titchfield, Fareham PO15 5RR, UK. ²¹²Institute of Avian Research "Vogelwarte Helgoland", An der Vogelwarte 21, 26386 Wilhelmshaven, Germany. ²¹³Department of Evolutionary Biology, Bielefeld University, North Rhine-Westphalia, Konsequenz 45, 33615 Bielefeld, Germany. ²¹⁴Ecology Department, Universidade Federal de Goiás, Av. Esperança, Campus Samambaia, Goiânia, Goiás 74690-900, Brazil. ²¹⁵Centre for Ecology and Conservation, University of Exeter, Penryn Campus, Penryn, Cornwall TR10 9FE, UK. ²¹⁶New South Wales, Department of Primary Industries Fisheries, Locked Bag 1, Nelson

Bay, NSW 2315, Australia. ²¹⁷Research Data Management, Leibniz Centre for Agricultural Landscape Research (ZALF), Eberswalder Straße 84, 15374 Müncheberg, Germany. ²¹⁸Biology Department, University of Wisconsin-La Crosse, 1725 State St, La Crosse, WI 54601, USA. ²¹⁹Department of Evolutionary Anthropology, Duke University, 130 Science Dr, Durham, NC 27708, USA. ²²⁰Earth and Life Institute, Ecology and Biodiversity, UCLouvain, Croix du Sud 4, L7.07.04, 1348 Louvain-la-Neuve, Belgium. ²²¹Future Regions Research Centre, Federation University Australia, Mt Helen, VIC 3350, Australia. ²²²United States, Department of Agriculture- Agricultural Research Service, 1701 10th Ave SW, Mandan, ND 58554, USA. ²²³Arthur Rylah Institute for Environmental Research, 123 Brown Street, Heidelberg, Victoria 3084, Australia. ²²⁴Epidemiology and Surveillance Support Unit, University of Lyon - French Agency for Food, Environmental and Occupational Health and Safety (ANSES), 31 Avenue Tony Garnier, 69007 Lyon, France. ²²⁵Center for Impact, UCLA Anderson, University of California, 110 Westwood Plaza, Gold Hall, Suite B.201L, Los Angeles/Los Angeles, CA 90095-1481, USA. ²²⁶Facultad de Ciencias, Universidad de la República, Iguá 4225, 11400 Montevideo, Montevideo, Uruguay. ²²⁷Washington, USA. ²²⁸Programa de Pós-Graduação em Ecologia, Instituto de Biologia, Centro de Ciências da Saúde, Universidade Federal do Rio de Janeiro, Cidade Universitária, Av. Carlos Chagas Filho 373, Rio de Janeiro, RJ, CEP 21941-902, Brazil. ²²⁹Vive Crop Protection, 6275 Northam Drive, Suite 1, Mississauga, ON L4V 1Y8, Canada. ²³⁰University of Cambridge, Trinity Ln, The Old Schools, Cambridge CB2 1TN, UK. ²³¹British Trust for Ornithology, BTO, The Nunnery, Thetford, Norfolk IP24 2PU, UK. ²³²Technology & Society Department, Rochester Institute of Technology, 7 Lomb Memorial Drive, Rochester, NY 14623, USA. ²³³Nomad Ecology, 822 Main Street, Martinez, CA 94553, USA. ²³⁴Wildland Resources Department, Utah State University, 5200 Old Main Hill, Logan, UT 84322, USA. ²³⁵Center for Biological Control, Department of Zoology and Entomology, Rhodes University, 1 Lower University Road, Barratt Complex, Biological Sciences Building Eastern Cape, Makhanda, South Africa. ²³⁶School of Mathematics and Statistics and Centre for Research in Ecological and Environmental Modelling, University of St Andrews, Buchanan Gardens, St Andrews, Scotland KY16 9LZ, UK. ²³⁷Department of Ecology and Evolutionary Biology, Cornell University, 215 Tower Road, Ithaca, NY 14853, USA. ²³⁸Department of Computational Landscape Ecology, Helmholtz Centre for Environmental Research – UFZ, Permoserstraße 15, 04318 Leipzig, Germany. ²³⁹Ecosystems and Global Change Group, School of the Environment, Trent University, 1600 West Bank Road, Peterborough, Ontario K0L 2V0, Canada. ²⁴⁰Instituto Universitario de Investigación en Gestión Forestal Sostenible (iuFOR), Universidad de Valladolid, Av. Madrid 44, 34071 Palencia, Spain. ²⁴¹Department of Biological Sciences, University of Bergen, Postboks, 7803, N-5020 Bergen, Norway. ²⁴²Department of Biological Science, University of Rhode Island, 9 East Alumni Ave, Kingston, RI 02881, USA. ²⁴³Department of Geography, McGill University, 805 Sherbrooke Street West, Montreal, Quebec H3A 0B9, Canada. ²⁴⁴School of Biological and Marine Sciences, University of Plymouth, Drake Circus, Plymouth, Devon PL4 8AA, UK. ²⁴⁵Biology Department, Wake Forest University, 1834 Wake Forest Rd., Winston Salem, NC 27109, USA. ²⁴⁶Plant Sciences, University of California, 1 Shields Ave, Davis/Davis, CA 95616, USA. ²⁴⁷College of Natural Resources, North Carolina State University, Jordan Hall, 2800 Faucette Dr, Raleigh, NC 27607, USA. ²⁴⁸Institute of Zoology, Technische Universität Dresden, Zellescher Weg 20b, 01217 Dresden, Germany. ²⁴⁹Helmholtz AI, Helmholtz Zentrum Muenchen, Ingolstaedter Landstr. 1, 85764 Neuherberg, Germany. ²⁵⁰FitzPatrick Institute of African Ornithology, University of Cape Town, University of Cape Town, Private Bag X3, Rondebosch, Cape Town 7701, South Africa. ²⁵¹Department of Cell & Developmental Biology, Division of Biosciences, University College London, London, UK. ²⁵²Biology, University of Saskatchewan, University of Saskatchewan, 112 Science Place, Saskatoon, SK S7N 5E2, Canada. ²⁵³Department of Biology, University of Regina, 3737 Wascana Pkwy, Regina, Saskatchewan S4S 0A2, Canada. ²⁵⁴Biology, University of Waterloo, 200 University Ave W, Waterloo, Ontario N2L 3G1, Canada. ²⁵⁵Department of Ecology & Evolutionary Biology, Biological Science Building, University of Michigan, 1105 North University Avenue, Ann Arbor, MI 48109-1085, USA. ²⁵⁶Dept. Ecologia, Instituto de Biologia, Universidade Federal do Rio de Janeiro, Av. Carlos Chagas Filho 373, Rio de Janeiro/RJ, CP 6820021942-902, Brazil. ²⁵⁷Lothian Analytical Services, Public Health Scotland, 1 South Gyle Crescent, Edinburgh EH12 9EB, UK. ²⁵⁸Institute for Evolution and Biodiversity, University of Muenster, Huefferstr. 1, DE-48149 Muenster, Germany. ²⁵⁹Department of Environmental Sciences, Western Norway University of Applied Sciences, P.O. box 133, 6851 Sogndal, Norway. ²⁶⁰Department of Biological Sciences, University of Southern Maine, 70 Falmouth St, Portland, ME 04103, USA.

²⁶¹Center for Ecosystem Science and Society, Northern Arizona University, PO Box 5620, Flagstaff, AZ 86011, USA. ²⁶²Center for Watershed Sciences, University of California, Davis, 1 Shields Ave, Davis, CA 95616, USA. ²⁶³School of Agriculture and Environmental Science, University of Southern Queensland, 487-535 West Street, Toowoomba, Qld 4350, Australia. ²⁶⁴Department of Ecology, Evolution, and Organismal Biology, Iowa State University, 2200 Osborn Dr, Ames, IA 50011, USA. ²⁶⁵Fram Project AS, Ymers vei 2, 0588 Oslo, Norway. ²⁶⁶Department of Food Science & Technology, Virginia Polytechnic Institute and State University, 22 Food Science Building (0418) 360 Duck Pond Drive Virginia Tech, Blacksburg, VA 24061, USA. ²⁶⁷School of Applied Sciences, School of Applied Sciences, University of Brighton, University of Brighton, Lewes Road, Brighton BN2 4GJ, UK. ²⁶⁸Department of Biology, Biology Research and Administration Building, University of Oxford, 11a Mansfield Rd, Oxford OX1 3SZ, UK. ²⁶⁹Department of Integrative Biology, University of Colorado, P.O. Box 173364, Denver/Denver, CO 80217-3364, USA. ²⁷⁰Ecology, Evolution, and Behavior Program, Michigan State University, East Lansing, MI 48824, USA. ²⁷¹ISEM, University of Montpellier, CNRS, Place Eugène Bataillon-Cedex 05, 34095 Montpellier, France. ²⁷²Laboratoire d'Ethologie Expérimentale et Comparée, LEEC, Université Sorbonne Paris Nord, 99 avenue Jean-Baptiste Clément, UR444393430 Villetaneuse, France. ²⁷³Department of Science and Environment, Lake Superior State University, 650 W Easterday Ave, Sault Sainte Marie, MI 49783, USA.

Received: 17 August 2019 Accepted: 19 December 2024
Published online: 06 February 2025

References

1. Aczel, Balazs, Barnabas Szasz, Gustav Nilsson, Olmo R van den Akker, Casper J Albers, Marcel ALM van Assen, Jojanneke A Bastiaansen, et al. 2021. "Consensus-Based Guidance for Conducting and Reporting Multi-Analyst Studies. *eLife*. 10 (November). <https://doi.org/10.7554/elife.72185>.
2. Allaire, J. J., Charles Teague, Carlos Scheidegger, Yihui Xie, and Christophe Dervieux. 2024. Quarto. <https://doi.org/10.5281/zenodo.5960048>.
3. Allaire, JJ, Yihui Xie, Christophe Dervieux, Jonathan McPherson, Javier Luraschi, Kevin Ushey, Aron Atkins, et al. 2024. markdown: Dynamic Documents for r. <https://github.com/rstudio/markdown>.
4. Arif S, Aaron MacNeil M. Predictive models aren't for causal inference. *Ecology Letters*. 2022;25(8):1741–5. <https://doi.org/10.1111/ele.14033>.
5. Arif, Suchinta, and M. Aaron MacNeil. 2023. "Applying the Structural Causal Model Framework for Observational Causal Inference in Ecology." *Ecological Monographs*. 93(1):e1554. <https://doi.org/10.1002/ecm.1554>.
6. Arnold, Jeffrey B. 2024. ggthemes: Extra Themes, Scales and Geoms for "ggplot2". <https://jrnold.github.io/ggthemes/>.
7. Atkinson, Joe, Lars A. Brudvig, Max Mallen-Cooper, Shinichi Nakagawa, Angela T. Moles, and Stephen P. Bonser. 2022. "Terrestrial Ecosystem Restoration Increases Biodiversity and Reduces Its Variability, but Not to Reference Levels: A Global Meta-Analysis. *Ecology Letters*. 25 (7): 1725–37. <https://doi.org/10.1111/ele.14025>.
8. Auspurg Katrin, Brüderl Josef. Has the Credibility of the Social Sciences Been Credibly Destroyed? Reanalyzing the 'Many Analysts, One Data Set' Project. *Socius*. 2021;7:23780231211024420. <https://doi.org/10.1177/23780231211024421>.
9. Bartoń, Kamil. MuMIn: Multi-Model Inference. R package version 1.47.5. 2023. <https://cran.r-project.org/web/packages/MuMIn>.
10. Baselga, Andres, David Orme, Sebastien Villegier, Julien De Bortoli, Fabien Leprieur, Maxime Logez, Sara Martinez-Santalla, et al. 2023. beta-part: Partitioning Beta Diversity into Turnover and Nestedness Components. <https://CRAN.R-project.org/package=beta-part>.
11. Bates, Douglas, Martin Mächler, Ben Bolker, and Steve Walker. 2015. "Fitting Linear Mixed-Effects Models Using lme4. 2015;67(1):48. <https://doi.org/10.18637/jss.v067.i01>.
12. Blake, Kevin. 2022. NatParksPalettes: Color Palettes Inspired by National Parks. <https://github.com/kevinsblake/NatParksPalettes>.
13. Bolker, Ben, David Robinson, Dieter Menne, Jonah Gabry, Paul Buerkner, Christopher Hau, William Petry, et al. 2024. broom.mixed: Tidying Methods for Mixed Models. <https://github.com/bbolker/broom.mixed>.

14. Borenstein Michael, Higgins Julian P T, Hedges Larry, Rothstein Hannah. Basics of Meta-Analysis: I^2 Is Not an Absolute Measure of Heterogeneity. *Research Synthesis Methods*. 2017;8(5):5–18. <https://doi.org/10.1002/rsm.1230>.
15. Botvinik-Nezer Rotem, Holzmeister Felix, Camerer Colin F, Dreber Anna, Huber Juergen, Johannesson Magnus, Kirchler Michael, et al. Variability in the Analysis of a Single Neuroimaging Dataset by Many Teams. *Nature*. 2020;582(7810):84–8.
16. Breznau Nate, Rinke Eike Mark, Wuttke Alexander, Nguyen Hung H V, Adem Muna, Adriaans Jule, Alvarez-Benjumea Amalia, et al. Observing Many Researchers Using the Same Data and Hypothesis Reveals a Hidden Universe of Uncertainty. *Proceedings of the National Academy of Sciences*. 2022;119(44): e2203150119. <https://doi.org/10.1073/pnas.2203150119>.
17. Briga, Michael, and Simon Verhulst. 2021. "Mosaic Metabolic Ageing: Basal and Standard Metabolic Rates Age in Opposite Directions and Independent of Environmental Quality, Sex and Life Span in a Passerine. *Functional Ecology*. 35 (5): 1055–68. <https://doi.org/10.1111/1365-2435.13785>.
18. Brooks, Mollie E., Kasper Kristensen, Koen J. van Benthem, Arni Magnusson, Casper W. Berg, Anders Nielsen, Hans J. Skaug, Martin Maechler, and Benjamin M. Bolker. 2017. "glmmTMB Balances Speed and Flexibility Among Packages for Zero-Inflated Generalized Linear Mixed Modeling. *The R Journal*. 9(2):378–400. <https://doi.org/10.32614/RJ-2017-066>.
19. Buck Robert J, Fieberg John, Larkin Daniel J. The use of weighted averages of Hedges' d in meta-analysis: Is it worth it? *Methods in Ecology and Evolution*. 2022;13(5):1093–105. <https://doi.org/10.1111/2041-210X.13818>.
20. Burnham, K. P., and D. R. Anderson. 2002. *Model Selection and Multimodel Inference: A Practical Information-Theoretical Approach*. Book. 2nd ed. New York: Springer-Verlag. <https://doi.org/10.1007/b97636>.
21. Cade, Brian S. 2015. "Model Averaging and Muddled Multimodel Inferences. *Ecology*. 96(9):2370–82. <http://www.jstor.org.ezproxy.whitman.edu/stable/24702343>.
22. Capilla-Lasheras Pablo, Thompson Megan J, Sánchez-Tójar Alfredo, Hadou Yacob, Branston Claire J, Réale Denis, Charmantier Anne, Dominoni Davide M. A Global Meta-Analysis Reveals Higher Variation in Breeding Phenology in Urban Birds Than in Their Non-Urban Neighbours. *Ecology Letters*. 2022;25(11):2552–70. <https://doi.org/10.1111/ele.14099>.
23. Coretta Stefano, Casillas Joseph V, Roessig Simon, Franke Michael, Ahn Byron, Al-Hoorie Ali H, Al-Tamimi Jalal, et al. Multidimensional Signals and Analytic Flexibility: Estimating Degrees of Freedom in Human-Speech Analyses. *Advances in Methods and Practices in Psychological Science*. 2023;6(3):25152459231162570. <https://doi.org/10.1177/25152459231162567>.
24. Dancho, Matt, and Davis Vaughan. 2023. *Timetk: A Tool Kit for Working with Time Series*. <https://CRAN.R-project.org/package=timetk>.
25. DeKogel, C. H. 1997. "Long-Term Effects of Brood Size Manipulation on Morphological Development and Sex-Specific Mortality of Offspring. *Journal of Animal Ecology*. 66(2):167–78. <Go to ISI>://WOS:A1997WQ19600003.
26. Deressa, Teshome, David Stern, Jaco Vangronsveld, Jan Minx, Sebastien Lizin, Robert Malina, and Stephan Bruns. 2023. "More Than Half of Statistically Significant Research Findings in the Environmental Sciences Are Actually Not. *EcoEvoRxiv*. <https://doi.org/10.32942/X24G6Z>.
27. Dormann, Carsten F., Jane Elith, Sven Bacher, Carsten Buchmann, Gudrun Carl, Gabriel Carré, Jaime R. García Marquéz, et al. 2013. "Collinearity: A Review of Methods to Deal with It and a Simulation Study Evaluating Their Performance. *Ecography*. 36(1):27–46. <https://doi.org/10.1111/j.1600-0587.2012.07348.x>.
28. Fanelli Daniele, Costas Rodrigo, Ioannidis John P. A. Meta-Assessment of Bias in Science. *Proceedings of the National Academy of Sciences*. 2017;114:3714–9. <https://doi.org/10.1073/pnas.1618569114>.
29. Fanelli Daniele, Ioannidis John P. A. US Studies May Overestimate Effect Sizes in Softer Research. *Proceedings of the National Academy of Sciences*. 2013;110(37):15031–6. <https://doi.org/10.1073/pnas.1302997110>.
30. Fidler Fiona, Burgman Mark A, Cumming Geoff, Buttrose Robert, Thomson Neil. Impact of Criticism of Null-Hypothesis Significance Testing on Statistical Reporting Practices in Conservation Biology. *Conservation Biology*. 2006;20(5):1539–44. <https://doi.org/10.1111/j.1523-1739.2006.00525.x>.
31. Fidler Fiona, Chee Yung En, Wintle Bonnie C, Burgman Mark A, McCarthy Michael A, Gordon Ascelin. Metaresearch for Evaluating Reproducibility in Ecology and Evolution. *BioScience*. 2017;67(3):282–9. <https://doi.org/10.1093/biosci/biw159>.
32. Firke, Sam. 2023. *janitor: Simple Tools for Examining and Cleaning Dirty Data*. <https://github.com/sfirke/janitor>.
33. Forstmeier Wolfgang, Wagenmakers Eric-Jan, Parker TH. Detecting and Avoiding Likely False-Positive Findings – a Practical Guide. *Biological Reviews*. 2017;92:1941–68. <https://doi.org/10.1111/brv.12315>.
34. Fraser Hannah, Parker Tim, Nakagawa Shinichi, Barnett Ashley, Fidler Fiona. Questionable Research Practices in Ecology and Evolution. *PLOS ONE*. 2018;13(7): e0200303. <https://doi.org/10.1371/journal.pone.0200303>.
35. Gamer, Matthias, Jim Lemon, and Ian Fellows Puspendra Singh. 2019. *irr: Various Coefficients of Interrater Reliability and Agreement*. <https://www.r-project.org>.
36. Gelman Andrew, Loken Eric. The Garden of Forking Paths: Why Multiple Comparisons Can Be a Problem, Even When There Is No 'Fishing Expedition' or 'p-Hacking' and the Research Hypothesis Was Posited Ahead of Time. Department of Statistics: Columbia University; 2013.
37. Gelman Andrew, Weakliem David. Of Beauty, Sex, and Power. *American Scientist*. 2009;97:310–6.
38. Gould Elliot, Fraser Hannah S, Nakagawa Shinichi, Parker Timothy H. *ManyEcoEvo: Meta-Analyse Data from ManyAnalyst Style Studies*. 2023. Zenodo. <https://doi.org/10.5281/zenodo.10046153>.
39. Gould Elliot, Fraser Hannah S, Nakagawa Shinichi, Parker Timothy H. *egoliot/ManyAnalysts: Manuscript Source Code for "Same data, different analysts: variation in effect sizes due to analytical decisions in ecology and evolutionary biology."* 2024. Zenodo. <https://doi.org/10.5281/zenodo.13850927.Version2.0.2>.
40. Grueber, C. E., S. Nakagawa, R. J. Laws, and I. G. Jamieson. 2011. "Multimodel Inference in Ecology and Evolution: Challenges and Solutions. *Journal of Evolutionary Biology*. 24(4):699–711. <https://doi.org/10.1111/j.1420-9101.2010.02210.x>.
41. Harrell Jr, Frank E. 2024. *Hmisc: Harrell Miscellaneous*. <https://hbiostat.org/R/Hmisc/>.
42. Hester, Jim, Lionel Henry, Kirill Müller, Kevin Ushey, Hadley Wickham, and Winston Chang. 2024. *withr: Run Code "With" Temporarily Modified Global State*. <https://withr.r-lib.org>.
43. Higgins Julian P T, Thompson Simon G, Deeks Jonathan J, Altman Douglas G. Measuring Inconsistency in Meta-Analyses. *BMJ*. 2003;327(7414):557–60. <https://doi.org/10.1136/bmj.327.7414.557>.
44. Huntington-Klein, Nick, Andreu Arenas, Emily Beam, Marco Bertoni, Jeffrey R. Bloem, Pralhad Burli, Naibin Chen, et al. 2021. "The Influence of Hidden Researcher Decisions in Applied Microeconomics. *Economic Inquiry*. 59(3):944–60. <https://doi.org/10.1111/ecin.12992>.
45. Iannone, Richard, Joe Cheng, Barret Schloerke, Ellis Hughes, Alexandra Lauer, and JooYoung Seo. 2024. *gt: Easily Create Presentation-Ready Display Tables*. <https://gt.rstudio.com>.
46. Jennions, M. D., C. J. Lortie, M. S. Rosenberg, and H. R. Rothstein. 2013. "Publication and Related Biases." Book Section. In *Handbook of Meta-Analysis in Ecology and Evolution*, edited by J. Koricheva, J. Gurevitch, and K. Mengersen, 207–36. Princeton, USA: Princeton University Press.
47. Kassambara, Alboukadel. 2023. *ggpubr: "ggplot2" Based Publication Ready Plots*. <https://rpkgs.datanovia.com/ggpubr/>.
48. Kimmel Kaitlin, Avolio Meghan L, Ferraro Paul J. Empirical Evidence of Widespread Exaggeration Bias and Selective Reporting in Ecology. *Nature Ecology & Evolution*. 2023. <https://doi.org/10.1038/s41559-023-02144-3>.
49. Klein Richard A, Ratliff Kate A, Vianello Michelangelo, Adams Jr Reginald B, Bahník Štěpán, Bernstein Michael J, Bocian Konrad, et al. Investigating Variation in Replicability: A "Many Labs" Replication Project. *Social Psychology*. 2014;45(3):142–52. <https://doi.org/10.1027/1864-9335/a000178>.
50. Klein Richard A, Vianello Michelangelo, Hasselman Fred, Adams Byron G, Adams Reginald B, Alper Sinan, Aveyard Mark, et al. *Many Labs 2: Investigating Variation in Replicability Across Samples and*

- Settings. *Advances in Methods and Practices in Psychological Science*. 2018;1(4):443–90. <https://doi.org/10.1177/2515245918810225>.
51. Knight K. *Mathematical Statistics*. Book. New York: Chapman; Hall; 2000.
 52. Koricheva, Julia, and Jessica Gurevitch. 2014. "Uses and Misuses of Meta-Analysis in Plant Ecology. *Journal of Ecology*. 102(4):828–44. <https://doi.org/10.1111/1365-2745.12224>.
 53. Kou-Giesbrecht, Sian, and Duncan N. L. Menge. 2021. "Nitrogen-Fixing Trees Increase Soil Nitrous Oxide Emissions: A Meta-Analysis. *Ecology*. 102(8):e03415. <https://doi.org/10.1002/ecy.3415>.
 54. Kuhn, Max, and Hannah Frick. 2022. *multilevelmod: Model Wrappers for Multi-Level Models*. <https://github.com/tidymodels/multilevelmod>.
 55. Kuhn, Max, and Hadley Wickham. 2020. *Tidymodels: A Collection of Packages for Modeling and Machine Learning Using Tidyverse Principles*. <https://www.tidymodels.org>.
 56. Kuznetsova, Alexandra, Per B. Brockhoff, and Rune H. B. Christensen. 2017. "lmerTest Package: Tests in Linear Mixed Effects Models. *Journal of Statistical Software*. 82(13):1–26. <https://doi.org/10.18637/jss.v082.i13>.
 57. Landau, William Michael. 2021. "The Targets r Package: A Dynamic Make-Like Function-Oriented Pipeline Toolkit for Reproducibility and High-Performance Computing. *Journal of Open Source Software*. 6(57):2959. <https://doi.org/10.21105/joss.02959>.
 58. Leybourne, Daniel J., Katharine F. Preedy, Tracy A. Valentine, Jorunn I. B. Bos, and Alison J. Karley. 2021. "Drought Has Negative Consequences on Aphid Fitness and Plant Vigor: Insights from a Meta-Analysis. *Ecology and Evolution*. 11(17):11915–29. <https://doi.org/10.1002/ece3.7957>.
 59. Lu, Xun, and Halbert White. 2014. "Robustness Checks and Robustness Tests in Applied Economics. *Journal of Econometrics*. 178:194–206. <https://doi.org/10.1016/j.jeconom.2013.08.016>.
 60. Lüdecke, Daniel, Mattan S. Ben-Shachar, Indrajeet Patil, and Dominique Makowski. 2020. "Extracting, Computing and Exploring the Parameters of Statistical Models Using R. *Journal of Open Source Software*. 5(53):2445. <https://doi.org/10.21105/joss.02445>.
 61. Lüdecke, Daniel, Mattan S. Ben-Shachar, Indrajeet Patil, Philip Waggoner, and Dominique Makowski. 2021. "performance: An R Package for Assessment, Comparison and Testing of Statistical Models. *Journal of Open Source Software*. 6(60):3139. <https://doi.org/10.21105/joss.03139>.
 62. Lüdecke, Daniel, Indrajeet Patil, Mattan S. Ben-Shachar, Brenton M. Wiernik, Philip Waggoner, and Dominique Makowski. 2021. "see: An R Package for Visualizing Statistical Models. *Journal of Open Source Software*. 6(64):3393. <https://doi.org/10.21105/joss.03393>.
 63. Luke SG. Evaluating Significance in Linear Mixed-Effects Models in r. *Behavior Research Methods*. 2017;49(4):1494–502.
 64. Makowski, Dominique, Mattan S. Ben-Shachar, Indrajeet Patil, and Daniel Lüdecke. 2020. "Estimation of Model-Based Predictions, Contrasts and Means. CRAN. <https://github.com/easystats/modelbased>.
 65. Masur, Philipp K., and Michael Scharrow. 2020. "specr: Conducting and Visualizing Specification Curve Analyses (Version 1.0.0)." <https://CRAN.R-project.org/package=specr>.
 66. Meschiari, Stefano. 2022. *Latex2exp: Use LaTeX Expressions in Plots*. <https://www.stefanom.io/latex2exp/>.
 67. Miles, C. 2008. "Testing Market-Based Instruments for Conservation in Northern Victoria." Book Section. In *Biodiversity: Integrating Conservation and Production: Case Studies from Australian Farms, Forests and Fisheries*, edited by T. Norton, T. Lefroy, K. Bailey, and G. Unwin, 133–46. Melbourne, Australia: CSIRO Publishing.
 68. Millard, Steven P. 2013. *EnvStats: An r Package for Environmental Statistics*. New York: Springer. <https://www.springer.com>.
 69. Molina, Isabel, and Yolanda Marhuenda. 2015. "sae: An R Package for Small Area Estimation. *The R Journal*. 7(1):81–98. <https://journal.r-project.org/archive/2015/RJ-2015-007/RJ-2015-007.pdf>.
 70. Morrissey, Michael B., and Graeme D. Ruxton. 2018. "Multiple Regression Is Not Multiple Regressions: The Meaning of Multiple Regression and the Non-Problem of Collinearity. *Philosophy, Theory, and Practice in Biology*. 10(3). <https://doi.org/10.3998/ptpbio.16039257.0010.003>.
 71. Müller, Kirill. 2020. *here: A Simpler Way to Find Your Files*. <https://here.r-lib.org/>.
 72. Nakagawa Shinichi, Cuthill Innes C. Effect Size, Confidence Interval and Statistical Significance: A Practical Guide for Biologists. *Biological Reviews*. 2007;82(4):591–605. <https://doi.org/10.1111/j.1469-185X.2007.00027.x>.
 73. Nakagawa, Shinichi, Malgorzata Lagisz, Michael D. Jennions, Julia Koricheva, Daniel W. A. Noble, Timothy H. Parker, Alfredo Sánchez-Tójar, Yefeng Yang, and Rose E. O'Dea. 2022. "Methods for Testing Publication Bias in Ecological and Evolutionary Meta-Analyses. *Methods in Ecology and Evolution*. 13(1):4–21. <https://doi.org/10.1111/2041-210X.13724>.
 74. Nakagawa, Shinichi, Malgorzata Lagisz, Rose E. O'Dea, Patrice Pottier, Joanna Rutkowska, Alistair M. Senior, Yefeng Yang, and Daniel W. A. Noble. 2023. "orchaRd 2.0: An R Package for Visualizing Meta-Analyses with Orchard Plots. *EcoEvoRxiv*. 12:4–12. <https://doi.org/10.32942/X2QC7K>.
 75. Nakagawa Shinichi, Yang Yefeng, Macartney Erin L, Spake Rebecca, Lagisz Malgorzata. Quantitative Evidence Synthesis: A Practical Guide on Meta-Analysis, Meta-Regression, and Publication Bias Tests for Environmental Sciences. *Environmental Evidence*. 2023;12(1):8. <https://doi.org/10.1186/s13750-023-00301-6>.
 76. Nakagawa S, Noble DW, Senior AM, Lagisz M. Meta-Evaluation of Meta-Analysis: Ten Appraisal Questions for Biologists. *BMC Biology*. 2017;15(1):18. <https://doi.org/10.1186/s12915-017-0357-7>.
 77. Nicolaus M, Michler SPM, Ubels R, van der Velde M, Komdeur J, Both C, Tinbergen JM. Sex-Specific Effects of Altered Competition on Nestling Growth and Survival: An Experimental Manipulation of Brood Size and Sex Ratio. *Journal of Animal Ecology*. 2009;78(2):414–26. <https://doi.org/10.1111/j.1365-2656.2008.01505.x>.
 78. Noble Daniel W. A, Lagisz Malgorzata, O'Dea Rose E, Nakagawa Shinichi. Nonindependence and Sensitivity Analyses in Ecological and Evolutionary Meta-Analyses. *Molecular Ecology*. 2017;26(9):2410–25. <https://doi.org/10.1111/mec.14031>.
 79. O'Hara Robert B, Johan Kotze D. Do Not Log-Transform Count Data. *Methods in Ecology and Evolution*. 2010;1(2):118–22. <https://doi.org/10.1111/j.2041-210X.2010.00021.x>.
 80. Open Science Collaboration. 2015. "Estimating the Reproducibility of Psychological Science." *Science*. 349(6251):aac4716. <https://doi.org/10.1126/science.aac4716>.
 81. Page Matthew J, Moher David. Evaluations of the Uptake and Impact of the Preferred Reporting Items for Systematic Reviews and Meta-Analyses (PRISMA) Statement and Extensions: A Scoping Review. *Systematic Reviews*. 2017;6(1):263. <https://doi.org/10.1186/s13643-017-0663-8>.
 82. Parker Timothy H, Forstmeier Wolfgang, Koricheva Julia, Fidler Fiona, Hadfield Jarrod D, Chee Yung En, Kelly Clint D, Gurevitch Jessica, Nakagawa Shinichi. Transparency in Ecology and Evolution: Real Problems, Real Solutions. *Trends in Ecology & Evolution*. 2016;31(9):711–9. <https://doi.org/10.1016/j.tree.2016.07.002>.
 83. Parker Timothy H, Yang Yefeng. Exaggerated Effects in Ecology. *Nature Ecology & Evolution*. 2023. <https://doi.org/10.1038/s41559-023-02156-z>.
 84. Pedersen, Thomas Lin. 2024. *patchwork: The Composer of Plots*. <https://patchwork.data-imaginist.com>.
 85. Pei Yifan, Forstmeier Wolfgang, Wang Daiping, Martin Katrin, Rutkowska Joanna, Kempenaers Bart. Proximate Causes of Infertility and Embryo Mortality in Captive Zebra Finches. *The American Naturalist*. 2020;196(5):577–96. <https://doi.org/10.1086/710956>.
 86. Qiu, Yixuan. 2024. *showtext: Using Fonts More Easily in r Graphs*. <https://github.com/yixuan/showtext>.
 87. R Core Team. 2024. *R: A Language and Environment for Statistical Computing*. Vienna, Austria: R Foundation for Statistical Computing. <https://www.R-project.org/>.
 88. Rosenberg, M. S. 2013. "Moment and Least-Squares Based Approaches to Metaanalytic Inference." Book Section. In *Handbook of Meta-Analysis in Ecology and Evolution*, edited by J. Koricheva, J. Gurevitch, and K. Mengersen, 108–24. Princeton, USA: Princeton University Press.
 89. Royle NJ, Hartley IR, Owens IPF, Parker GA. Sibling Competition and the Evolution of Growth Rates in Birds. *Proceedings of the Royal Society B-Biological Sciences*. 1999;266(1422):923–32. <https://doi.org/10.1098/rspb.1999.0725>.
 90. Scheinin, Ilari, Maria Kalimeri, Vilma Jagerroos, Juuso Parkkinen, Emmi Tikkanen, Peter Würtz, and Antti Kangas. 2020. *ggforestplot: Forestplots of Measures of Effects and Their Confidence Intervals*. <https://github.com/NightingaleHealth/ggforestplot>.
 91. Schloerke, Barret, Di Cook, Joseph Larmarange, Francois Briatte, Moritz Marbach, Edwin Thoen, Amos Elberg, and Jason Crowley. 2024. *GGally: Extension to "ggplot2"*. <https://ggobi.github.io/ggally/>.

92. Schweinsberg M, Feldman M, Staub N, van den Akker OR, van Aert RCM, Malm van Assen Y, Liu, et al. Same Data, Different Conclusions: Radical Dispersion in Empirical Results When Independent Analysts Operationalize and Test the Same Hypothesis. *Organizational Behavior and Human Decision Processes*. 2021;165:228–49. <https://doi.org/10.1016/j.jobhdp.2021.02.003>.
93. Senior Alistair M, Grueber Catherine E, Kamiya Tsukushi, Lagisz Malgorzata, O'Dwyer Katie, Santos Eduardo S. A, Nakagawa Shinichi. Heterogeneity in Ecological and Evolutionary Meta-Analyses: Its Magnitude and Implications. *Ecology*. 2016;97(12):3293–9. <https://doi.org/10.1002/ecy.1591>.
94. Shavit, A., and Aaron M. Ellison. 2017. *Stepping in the Same River Twice: Replication in Biological Research*. Edited Book. New Haven, Connecticut, USA: Yale University Press.
95. Siegel, Kyle R., Muskanjot Kaur, A. Calvin Grigal, Rebecca A. Metzler, and Gary H. Dickinson. 2022. "Meta-Analysis Suggests Negative, but pCO₂-Specific, Effects of Ocean Acidification on the Structural and Functional Properties of Crustacean Biomaterials." *Ecology and Evolution*. 12(6):e8922. <https://doi.org/10.1002/ece3.8922>.
96. Silberzahn R, Uhlmann EL, Martin DP, Anselmi P, Aust F, Awtry E, Bahnik Š, et al. Many Analysts, One Data Set: Making Transparent How Variations in Analytic Choices Affect Results. *Advances in Methods and Practices in Psychological Science*. 2018;1(3):337–56. <https://doi.org/10.1177/2515245917747646>.
97. Silge, Julia, and David Robinson. 2016. "tidytext: Text Mining and Analysis Using Tidy Data Principles in r". *JOSS*. 1(3). <https://doi.org/10.21105/joss.00037>.
98. Simons Daniel J, Shoda Yuichi, Stephen Lindsay D. Constraints on Generality (COG): A Proposed Addition to All Empirical Papers. *Perspectives on Psychological Science*. 2017. <https://doi.org/10.1177/174569161770863>.
99. Simonsohn, Uri, Joseph P. Simmons, and Leif D. Nelson. 2015. "Specification Curve: Descriptive and Inferential Statistics on All Reasonable Specifications." Manuscript. SSRN Electronic Journal. <https://doi.org/10.2139/ssrn.2694998>.
100. Simonsohn Uri, Simmons Joseph P, Nelson Leif D. Specification Curve Analysis. *Nature Human Behaviour*. 2020;4(11):1208–14. <https://doi.org/10.1038/s41562-020-0912-z>.
101. Sjöberg, Daniel D., Karissa Whiting, Michael Curry, Jessica A. Lavery, and Joseph Larmarange. 2021. "Reproducible Summary Tables with the Gtsummary Package." *The R Journal*. 13:570–80. <https://doi.org/10.32614/RJ-2021-053>.
102. Slowikowski, Kamil. 2024. *ggrepel: Automatically Position Non-Overlapping Text Labels with "ggplot2"*. <https://ggrepel.slowkow.com/>.
103. Stanton-Geddes, John, Cintia Gomes de Freitas, and Cristian de Sales Dambros. 2014. "In Defense of p Values: Comment on the Statistical Methods Actually Used by Ecologists." *Ecology*. 95(3):637–42. <https://doi.org/10.1890/13-1156.1>.
104. Steegen Sara, Tuerlinckx Francis, Gelman Andrew, Vanpaemel Wolf. Increasing Transparency Through a Multiverse Analysis. *Perspectives on Psychological Science*. 2016;11(5):702–12. <https://doi.org/10.1177/1745691616658637>.
105. Taylor, James W., and Kathryn S. Taylor. 2023. "Combining Probabilistic Forecasts of COVID-19 Mortality in the United States." *European Journal of Operational Research*. 304(1):25–41. <https://doi.org/10.1016/j.ejor.2021.06.044>.
106. Tierney, Nicholas, and Dianne Cook. 2023. "Expanding Tidy Data Principles to Facilitate Missing Data Exploration, Visualization and Assessment of Imputations." *Journal of Statistical Software*. 105(7):1–31. <https://doi.org/10.18637/jss.v105.i07>.
107. Touchon, Justin C., and Michael W. McCoy. 2016. "The Mismatch Between Current Statistical Practice and Doctoral Training in Ecology." *Ecosphere*. 7(8):e01394. <https://doi.org/10.1002/ecs2.1394>.
108. Ushey, Kevin, and Hadley Wickham. 2023. *renv: Project Environments*. <https://rstudio.github.io/renv/>.
109. van den Brand, Teun. 2024. *Ggh4x: Hacks for "ggplot2"*. <https://github.com/teunbrand/ggh4x>.
110. Werf Vander, Eric. Lack's Clutch Size Hypothesis: An Examination of the Evidence Using Meta-Analysis. *Ecology*. 1992;73(5):1699–705. <https://doi.org/10.2307/1940021>.
111. Hoef Ver, Jay M. Who Invented the Delta Method? *The American Statistician*. 2012;66(2):124–7. <https://doi.org/10.1080/00031305.2012.687494>.
112. Verhulst S, Holveck MJ, Riebel K. Long-Term Effects of Manipulated Natal Brood Size on Metabolic Rate in Zebra Finches. *Biology Letters*. 2006;2(3):478–80. <https://doi.org/10.1098/rsbl.2006.0496>.
113. Vesik PA, Morris WK, McCallum W, Apter R, Miles C. Processes of Woodland Eucalypt Regeneration: Lessons from the Bush Returns Trial. *Proceedings of the Royal Society of Victoria*. 2016;128:54–63.
114. Viechtbauer, Wolfgang. 2010. "Conducting Meta-Analyses in R with the metafor Package." *Journal of Statistical Software*. 36(3):1–48. <https://doi.org/10.18637/jss.v036.i03>.
115. Wickham, Hadley, Mara Averick, Jennifer Bryan, Winston Chang, Lucy D'Agostino McGowan, Romain François, Garrett Grolemund, et al. 2019. "Welcome to the tidyverse." *Journal of Open Source Software*. 4(43):1686. <https://doi.org/10.21105/joss.01686>.
116. Wickham, Hadley, Jim Hester, Winston Chang, and Jennifer Bryan. 2022. *devtools: Tools to Make Developing r Packages Easier*. <https://devtools.r-lib.org/>.
117. Wickham, Hadley, Thomas Lin Pedersen, and Dana Seidel. 2023. *scales: Scale Functions for Visualization*. <https://scales.r-lib.org>.
118. Wilke, Claus O. 2024. *cowplot: Streamlined Plot Theme and Plot Annotations for "ggplot2"*. <https://wilkelab.org/cowplot/>.
119. Xie, Yihui. 2024a. *knitr: A General-Purpose Package for Dynamic Report Generation in r*. <https://yihui.org/knitr/>.
120. ———. 2024b. *xfun: Supporting Functions for Packages Maintained by "Yihui Xie"*. <https://github.com/yihui/xfun>.
121. Yang Yefeng, Sánchez-Tójar Alfredo, O'Dea Rose E, Noble Daniel W. A, Koricheva Julia, Jennions Michael D, Parker Timothy H, Lagisz Malgorzata, Nakagawa Shinichi. Publication Bias Impacts on Effect Size, Statistical Power, and Magnitude (Type m) and Sign (Type s) Errors in Ecology and Evolutionary Biology. *BMC Biology*. 2023;21(1):71. <https://doi.org/10.1186/s12915-022-01485-y>.
122. Zeileis, Achim, Jason C. Fisher, Kurt Hornik, Ross Ihaka, Claire D. McWhite, Paul Murrell, Reto Stauffer, and Claus O. Wilke. 2020. "colspace: A Toolbox for Manipulating and Assessing Colors and Palettes." *Journal of Statistical Software*. 96(1):1–49. <https://doi.org/10.18637/jss.v096.i01>.

Publisher's Note

Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.