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Article

Integrated Approach to Ship Electrification Using Fuel Cells and an Ammonia Decomposition System

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Abstract: This study investigates the environmental and economic performance of integrating a proton exchange membrane fuel cell, battery systems, and an organic Rankine cycle-based waste heat recovery system for ship electrification. The analysis examines an onboard ammonia decomposition system for hydrogen production and ammonia production pathways. Additionally, the study benchmarks the effectiveness of onboard ammonia decomposition against green hydrogen bunkering scenarios (H₂-BS). The analysis is based on data collected over two years from a bulk carrier provided by Laskaridis Shipping Co., Ltd. The environmental analysis includes well-to-wake emissions calculations. At the same time, economic performance is assessed through levelised cost of energy (LCOE) computations for 2025 and 2040, factoring in different fuel and carbon price scenarios. Consequently, the analysis utilises the Complex Proportional Assessment method to compare configurations featuring various ammonia production pathways across economic cases. The results indicate that green and pink ammonia feedstocks achieve maximum equivalent carbon dioxide reductions in the electrification plant of up to 47.28% and 48.47%, respectively, compared to H₂-BS and 95.56% and 95.66% compared to the base scenario. Ammonia decomposition systems prove more economically viable than H₂-BS due to lower storage and fuel costs, leading to competitive LCOE values that improve under higher carbon pricing scenarios.

Keywords: maritime decarbonisation; ship electrification; ammonia (NH₃) decomposition; hydrogen (H₂); proton exchange membrane fuel cell (PEMFC); levelised cost of energy (LCOE); multi-criteria decision-making (MCDM)



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1. Introduction

Maritime transportation has been crucial to carrying cargo and passengers for decades, and its worldwide capacity has increased drastically recently [1]. The increased fossil fuel usage for marine vessels' propulsion and electricity generation significantly contributes to global warming [2]. The International Maritime Organization (IMO) has instituted regulations designed to advance sustainable shipping practices while establishing ambitious

targets for the reduction of waterborne emissions [3]. The short-term objectives for 2030 include a 40% reduction in carbon dioxide (CO₂) emissions and a 20% decrease in overall greenhouse gases (GHG), with the intention of achieving 30%. Additionally, there is a target for 5% utilisation of zero-emission fuels, striving for 10%. In the midterm, the goals are to achieve a 70% reduction in total GHG emissions, with aspirations for an 80% decrease, ultimately leading to net-zero emissions by 2050 [4].

In recent years, ammonia (NH₃) and NH₃-involved compounds have emerged as a popular zero-carbon alternative fuel for marine vessels [5,6]. Its use in marine diesel engines (MDEs) and as a hydrogen (H₂) carrier has been a timely topic for the maritime sector [6]. NH₃ contains 17.6 wt.% H₂, and it is a gas at room temperature and pressure, becoming liquid at 263 K or 1000 kPa [7]. Its well-established industrial production, primarily for fertiliser, makes it a cost-effective option [8].

NH₃ has been prominent as an H₂ carrier in maritime applications, mainly due to the technical and safety challenges associated with storing H₂ onboard ships. H₂ storage is complicated by its low volumetric energy density and high flammability, which demand either cryogenic conditions for liquefaction or high-pressure containment systems. These requirements introduce significant engineering and safety concerns on marine vessels [9].

The most significant disadvantage of NH₃ is its toxicity and dangerous nature for human health. Considerable health consequences, including blindness, lung damage, brain damage, and maybe even death, can arise from exposure to elevated levels of NH₃ in the air [10,11]. Consequently, NH₃ leaks can quickly escalate into a significant risk to the accommodations and crew members on board [12].

Despite its high toxicity and corrosiveness [13], NH₃ is an effective H₂ carrier [6]. However, its properties require an improved design meeting specific codes and rules, in addition to increased requirements [12]. The production methodology of NH₃ is also a significant subject since green (renewables-sourced) or pink (nuclear-sourced) NH₃ has been required to meet the net-zero targets. The clean NH₃ capacity and prices to meet the demand of the maritime sector have been a remarkable challenge for the utilisation in shipping operations [14]. Compared to the bunkering, storage, and transfer operations for H₂, producing it from NH₃ would be much more cost-effective and applicable [15].

1.1. Literature Review

Several research papers about the onboard utilisation of NH₃-based systems to generate H₂ on marine vessels have been published. Over the past few years, the utilisation of NH₃ as an H₂ carrier or as a direct fuel for fuel cells (FCs) in marine applications has attained prominence, driven by the IMO's targets for maritime decarbonisation.

Boggs and Botte [16] applied NH₃ electrolysis with an NH₃ electrolytic cell to generate H₂ onboard vessels in a Proton Exchange Membrane Fuel Cell (PEMFC). H₂ can be extracted from NH₃ via electrolysis, which theoretically requires 95% less energy than water electrolysis. Specifically, NH₃ electrolysis consumed 1.55 Whg⁻¹ of H₂, whereas water electrolysis requires 33 Whg⁻¹ of H₂.

Wang et al. [17] demonstrated the combined usage of NH₃ auto-thermal reforming and selective NH₃ oxidation to produce H₂ by utilising the waste heat in the exhaust gas of diesel engines. H₂ production at 2.5–3.2 L/min was achieved when the NH₃ was supplied steadily at a 3 L/min flow rate.

McKinlay et al. [18] presented a dynamic simulation of NH₃ decomposition to produce H₂ for PEMFCs on marine vessels. The analysis included a detailed assessment of NH₃ demand, revealing that the reference cargo ship would require 150 t of NH₃ per voyage, necessitating a 586 m³ storage tank.

Ye et al. [19] compared conventional diesel engines with FC propulsion systems using NH_3 and H_2 as fuels for a small ship working as a sea taxi. The storage options of H_2 were evaluated while the presentation of an NH_3 cracking system was ensured in the study. The FC options met the demand of the vessel overall with significant equivalent CO_2 (CO_{2e}) reduction ratios.

Zhu et al. [20] compared methanol and NH_3 as H_2 sources using the life cycle approach, using commercial software. Methanol and NH_3 were evaluated as H_2 carriers compared to coal, natural gas, and renewable energy sources. The solar energy-based NH_3 plant produced the lowest CO_{2e} per MJ at 43.9 g.

Spatolisano et al. [21] assessed the potential of NH_3 as a zero-carbon H_2 carrier, analysed the maritime transportation of feedstocks and products, and evaluated the maturity of decomposing technologies for the industry.

Duong et al. [22] presented an innovative multigeneration system that utilises NH_3 as the primary fuel for marine applications. The integration of various components, including PEMFC, solid-oxide FC (SOFC), gas turbines, and waste heat recovery systems (WHRSs) based on diverse thermodynamic cycles, was ensured. The analysis revealed an energy efficiency of 60.69% and an exergy efficiency of 57.50%, with waste heat recovery contributing 1634.46 kW, which accounts for 30.07% of the total power output.

Restelli et al. [23] performed a comprehensive thermo-economic assessment of NH_3 as a green H_2 source onboard ships. The transport cost for the industrial application was between 5.49 and 6.34 EUR/kg, whereas for the mobility end use, it varied between 6.80 and 12.22 EUR/kg.

Di Micco et al. [24] investigated the design and viability of NH_3 -based propulsion systems for maritime applications, focusing on two FC technologies: PEMFC and SOFC. The findings revealed that implementing these NH_3 systems increases weight and volume, decreasing cargo capacity by 3.3% to 4.8%.

Duong et al. [25] examined the techno-economic feasibility of a direct NH_3 SOFC system, enhanced by a gas turbine and integrated with a multi-generation framework that includes various energy recovery systems. Key findings indicated that the levelised cost of energy (LCOE) ranges from USD 0.482 to USD 0.554 per kWh, with a variation of about 6.2%. At the same time, the discounted payback period for unsubsidised and subsidised scenarios falls between 6.7 and 9.5 years.

Research on NH_3 utilisation for H_2 production in marine vessels can be broadly categorised into electrolysis-based, thermo-chemical, and system-integrated approaches. Electrolytic systems highlighted the energy efficiency potential of NH_3 electrolysis compared to water electrolysis [16]. Thermo-chemical methods, including auto-thermal reforming [17,18], demonstrated effective H_2 generation using onboard waste heat, though storage and logistics remain challenging.

Integrated system designs have taken a broader view of NH_3 as a fuel or carrier, balancing efficiency, emissions, and cost. Studies comparing NH_3 -based systems to conventional marine fuels and H_2 alternative carriers, such as methanol, demonstrated that NH_3 frequently emerges as a lower-emission option when renewable sources are used for NH_3 production [19,20]. Techno-economic [23,25] and system-level assessments [21,22] have shown promising efficiencies and payback periods but have also highlighted concerns regarding volume, weight, and infrastructure demands [24].

In summary, NH_3 is a technically viable H_2 source and direct fuel, offering substantial emission reduction potential, especially when integrated into hybrid or WHRSs. Economic feasibility varies, with performance and system complexity depending heavily on technology choice and operating conditions.

Various research papers have focused on the PEMFC applications on ships apart from NH_3 usage, exploring their potential to enhance energy efficiency, reduce emissions, and improve overall performance in maritime operations.

Sarı et al. [26] established a “Reference Energy System” for a chemical tanker to analyse energy flow from various sources. It evaluated a baseline scenario and introduced two alternatives, including one focused on H_2 PEMFCs. The findings indicated that H_2 PEMFCs can achieve a 60% reduction in carbon emissions compared to diesel generators as a primary propulsion system.

Vieira et al. [27] determined the best configuration for a ship’s power system, integrating FCs and batteries, considering battery cycles. The focus was on a retrofitted platform supply vessel with its original generators and additional FCs and batteries. Simulations using commercial software showed that the optimal setup, main and auxiliary generators, a 3119-kW lithium nickel manganese cobalt battery, a 250 kW PEMFC, and 581 kg of H_2 achieved a 10.69% reduction in CO_2 emissions.

Bang et al. [28] examined methane (CH_4) with PEMFCs for their performance and cost advantages over hydrogen. CH_4 was reformed using steam reforming and evaluated through five gas treatment systems. The results showed that combining water–gas shift and partial oxidation reforming keeps carbon monoxide below 10 ppm and prevents CH_4 in the exhaust.

Bagherabadi et al. [29] developed a model for a marine power system using PEMFCs and batteries. It validated the model against a 500 kW PEMFC and demonstrated its effectiveness for performance analysis and control design. The model supported flexibility in configuration and could assess various manoeuvring scenarios.

Lee et al. [30] introduced a PEMFC system combined with an organic Rankine cycle (ORC) that utilises cold exergy from liquid H_2 and waste heat from the PEMFC for marine applications. Findings indicated that the system can produce an additional 221 kW of power, with energy and exergy efficiencies of 40.45% and 43.52%, respectively. Economic analysis suggests a payback period of 11.2 years and a net present value of USD 295,268, demonstrating the system’s potential viability.

Wang et al. [31] assessed FC technologies for ship power, focusing on energy efficiency and environmental impact. Low-temperature PEMFCs emerged as the most viable option for sustainable propulsion, with the lowest Energy Efficiency Design Index value at 10.05 g $\text{CO}_2/\text{t}\cdot\text{km}$.

Penga et al. [32] explored the potential of hybrid systems combining PEMFCs and battery packs as a cleaner alternative to diesel propulsion for marine vessels. A numerical model was developed to optimise a hybrid power system for a specific route, featuring a 300 kWe PEMFC stack and a 424-kWh battery. The results demonstrated that this new system significantly reduced emissions, requiring only 284.7 kg of H_2 compared to 1524 kg of diesel previously consumed.

Yuksel et al. [33] evaluated the integration of various FC technologies with battery storage systems and WHRS for retrofitting marine electricity generation plants. The study employed mathematical modelling and real-time operational data from a Kamsarmax bulk carrier. The analysis showed that hybrid configurations combining PEMFC and SOFC, powered by liquefied green H_2 , can reduce well-to-wake equivalent CO_2 emissions from the ship’s electrification system by up to 91.79%.

Aziz et al. [34] presented a coordinated control strategy for a hybrid shipboard power system using PEMFCs and batteries. The strategy improved operational efficiency and stabilised power during load changes, achieving a 14.16% increase in system efficiency over traditional methods while maintaining stable voltage tracking.

PEMFCs have received considerable attention for marine decarbonisation, with studies focusing on propulsion replacement, hybridisation, and system optimisation. Initial work [18,26,27] showed that replacing diesel generators with PEMFCs could significantly reduce emissions. Hybrid configurations combining PEMFCs with batteries or other FCs enhanced operational flexibility and emission reduction outcomes [32,33].

Performance modelling and control strategies [29,34] supported system adaptability and efficiency under variable marine conditions. Integration with WHRSs [30] and comparative assessments of PEMFCs against other FC types [31] further confirmed their strong potential in meeting environmental and regulatory targets. Overall, PEMFCs are a well-supported and adaptable low-emission option for marine applications, especially when deployed in hybrid systems or with energy recovery enhancements.

The literature review on PEMFC and NH_3 utilisation in marine vessels reveals a significant surge in research publications in recent years, driven by IMO decarbonisation objectives and growing environmental awareness. NH_3 evaluations as a fuel in power systems have focused on economic and environmental performance.

The studies investigated onboard H_2 production from NH_3 decomposition and evaluated the financial performance, required NH_3 capacity, and ecological benefits. The presented designs have been proposed as a concept for the propulsion system. A gap in the literature has been identified regarding the integrated design evaluation of onboard NH_3 cracking systems and FCs for marine electricity generation plants. The existing literature has been notably deficient in comprehensive analyses that integrate well-to-wake environmental evaluations alongside assessments of economic factors and design requirements for the combined configuration of an integrated NH_3 decomposer and H_2 PEMFC, battery, and WHRS.

1.2. Objectives, Motivation and Novelty

This study presents a comparative analysis of conventional marine diesel generators (D/Gs) utilising heavy fuel oil (HFO), with FCs that employ green H_2 bunkering and NH_3 decomposition systems for onboard H_2 production derived from diverse NH_3 production pathways. Previous research conducted by the authors of [33] evaluated the potential of liquefied natural gas (LNG) reform for H_2 production in FCs, alongside the performance of FCs utilising green H_2 bunkering. The top-ranked configurations utilising green H_2 bunkering from the prior study [33] were employed in this analysis.

This research aims to present an integrated design involving the onboard H_2 production and PEMFC/battery/WHRS hybrid configuration to meet the electricity demand of a Kamsarmax bulk carrier. The analysis has been performed using a robust system for real-time data collection. The feasibility of the proposed hybrid electrification system's economic and environmental performance has been assessed. The financial viability and bunkering challenges of H_2 systems have rendered them unfeasible for current vessels. The study's objective is to reduce fuel costs, increase fuel availability, and thereby enhance the economic performance of H_2 configurations by implementing NH_3 -cracking systems for the ships.

The upstream (well-to-tank) emissions arising from H_2 generation concerning the NH_3 production methodology and operational emission reductions have been examined to ascertain whether this configuration can serve as a viable option for meeting forthcoming decarbonisation targets. The tank capacities have been identified to illustrate the implementation's specific design and operational challenges. This article distinguishes itself from existing literature by presenting a comprehensive design structure that encompasses both H_2 production and consumption to address the electricity demands of a commercial marine vessel. The system's operational performance was evaluated, and the various NH_3

production methodologies were analysed across different economic projection scenarios using a multi-criteria decision-making (MCDM) algorithm. This paper’s innovative aspect lies in integrating a streamlined well-to-wake environmental approach with a multi-criteria decision-making (MCDM) framework for the onboard NH₃ decomposition system. This framework is designed for PEMFC/battery/WHRS configurations within the ship electrification system, accompanied by a discussion of its merits and disadvantages. The thorough investigation of various facets of the system contributes to the existing literature while acting as a significant resource for ship owners, managers, and designers to identify the decarbonisation pathway for their fleets.

2. Materials and Methods

This section provides an overview of the analytical framework and context underlying the research findings, detailing the description of the system utilised in the study. The case study focuses on the Kamsarmax bulk carrier M/V KASTOR, operated by Laskaridis Shipping Co. Ltd. (Athens, Greece) built in 2020, the vessel has a deadweight of 80,996 tonnes and an overall length of 229 metres. Its propulsion system features a HYUNDAI 6S60ME-C8.5 (Ulsan, South Korea) engine with a maximum power output of 9930 kW. The ship also includes an electrification plant comprising three YANMAR 6EY22LW D/Gs (Almere, The Netherlands) operating HFOs, each with a terminal power capacity of 720 kW. Ship particulars and sample data are provided in Yuksel et al. [33].

A bulk carrier without cargo handling gear (gearless bulk carrier) was selected for this study due to its structural potential and further suitability for integrating alternative energy systems, such as photovoltaic (PV) panels or wind turbines, given the available deck space and minimal obstructions. Additionally, these vessels typically exhibit a more stable electrical load profile during port operations, which supports consistent performance evaluation of onboard energy systems. The selection was also driven by the availability of high-quality, long-term sensor data, essential for robust analysis and validation of the proposed methods. Figure 1 demonstrates the application process of the techniques used in the study.

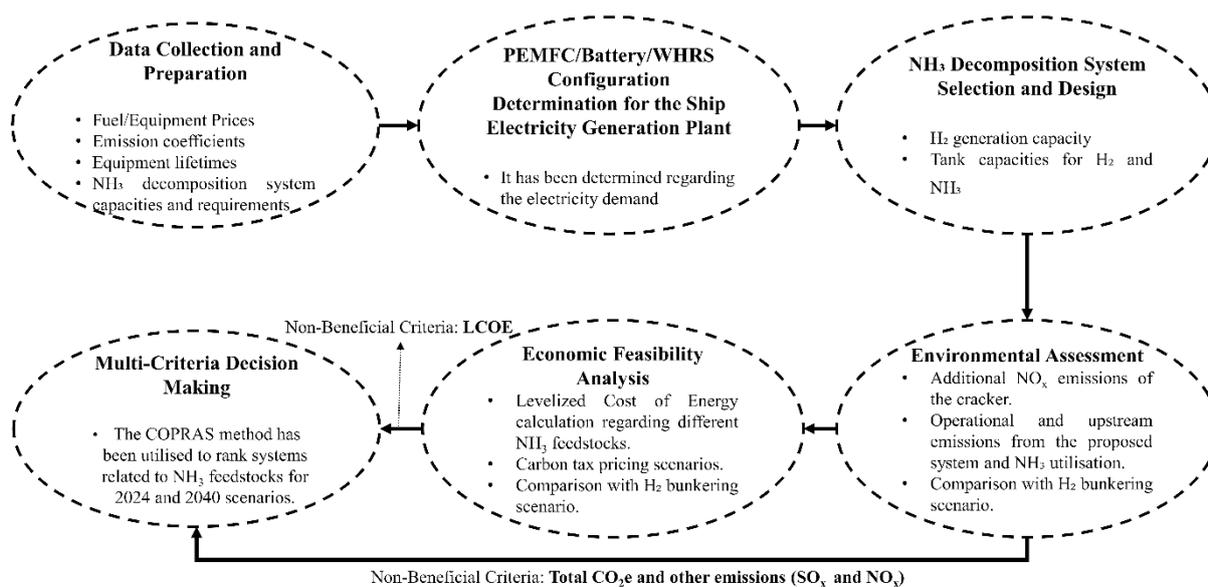


Figure 1. Methodology flowchart.

A data acquisition system was implemented to reliably gather information from sensors and control mechanisms on the case study ship. This system functions through a wireless network, facilitating efficient data collection and monitoring. It has been certified

by Bureau Veritas and complies with maritime safety and operational standards. Major components include the Quax 8S Node, which records voyage-specific metrics such as vessel speed and navigation information; the Quax G Node, which tracks the MDEs' performance; and the Quax S Node, which captures data on flow and revolutions to enhance overall monitoring abilities [35].

The data acquisition period spanned from 1 February 2021 to 10 February 2023. After preprocessing, the dataset was refined to comprise 1,003,490 entries, representing 1.96 years of operational metrics. The collected parameters included measures involving fuel flow, temperature, density, and engine powers, complemented by analyses of electrical load and exhaust data. Furthermore, details related to FC/battery curves, converter voltage, and conversion efficiencies were sourced from the literature and commercial system datasheets.

In the initial stage of the analysis, data on prices, emission coefficients, and properties of the NH_3 cracker were collected. The two PEMFC/battery/WHRS configurations for the marine electricity generation unit were proposed. Subsequently, the design of the NH_3 decomposition system, including the tank and production capacity for H_2 , was determined.

The environmental performance of the cracker-involved operations was assessed, considering nitrous oxide (NO_x) emissions from NH_3 combustion and the sulphur oxides (SO_x) and CO_{2e} resulting from the NH_3 production methodology. System economic viability was evaluated by calculating the LCOE for 2025 and 2040 using different colour-coded NH_3 as the feedstock. This financial performance was compared with the green H_2 bunkering scenarios (H_2 -BS). The outcomes of these analyses were combined in an MCDM analysis, and the configurations were ranked regarding the NH_3 colour codes.

The integration of the PEMFC/battery/WHRS system into the ship electrification plant was achieved using a simulation created in the Python 3.11 programming environment. It benchmarks the hybrid systems' emission reductions against conventional diesel engines.

The simulation workflow begins with importing required libraries and reading input data. Initial parameters for the hybrid system are defined, and writable lists for logging results are prepared. The algorithm first checks whether the WHRS can generate power based on the main engine (M/E) load. If so, it interpolates the exhaust flow and temperature according to the engine power, then calculates the power generation from the ORC-based WHRS. If not, the WHRS power supply is set to zero. The number of working FCs and D/Gs (if needed) and battery support, including charging and discharging, are determined depending on the power demand. As described above, the hierarchy in PEMFC/battery/WHRS hybrid configurations begins with WHRS and prioritises the PEMFCs. If the required power is within the capacity of the PEMFCs, the FC plant solely meets the load, and the number of active FCs is determined based on the demanded power. The PEMFC H_2 consumption is calculated based on the power– H_2 curves provided by the manufacturer [36]. If the required power exceeds the capacity of the PEMFC plant, the batteries are activated. However, the necessary charging power is added to the required grid power if the batteries need to be charged. Battery charging and discharging operations are determined by assessing the state-of-charge (SoC). Additionally, batteries are used to compensate for instantaneous load demands. If the PEMFC/battery/WHRS plant cannot meet the required power, D/Gs are utilised. The required D/G power is determined by subtracting the required total power from the available PEMFC/battery/WHRS supply. Then, depending on the required D/G power, the specific fuel consumption (SFC) is interpolated, and fuel consumption is calculated. Subsequently, the model computes the utilisation times of each equipment, fuel consumption, and emissions for the PEMFC/battery/WHRS hybrid configurations. Finally, the results are logged, structured into data frames, and exported to spreadsheets for further analysis. The algorithmic framework, which visually elucidates the processes, is presented in Appendix A. The energy management strategy

deciding the hierarchy between power equipment is detailed in a simplified format in Appendix B. Table 1 indicates the analysed configurations.

Table 1. Configurations analysed in the case study.

Case	DGs	PEMFCs	Battery Capacity	WHRS
Base	3 × 720 kW	N/A	N/A	N/A
C1	1 × 720 kW	4 × 200 kW	443 kWh	ORC (Average: 197.01 kW)
C2	2 × 720 kW	2 × 200 kW	123 kWh	ORC (Average: 197.01 kW)

The selection of FC types and sizes in this study was based on a comprehensive prior analysis ranking configurations by performance and size [33]. The chosen commercially available FCs align with the vessel’s load profile, which ranges from 200 kW to 800 kW per generator (see Figure 2 given in Yuksel et al. [33]), with a critical threshold power of 400 kW. Thus, the selected FC sizes are positioned either in the middle of this range (C2) or fully satisfy the requirements (C1) within this range to meet the specified power output.

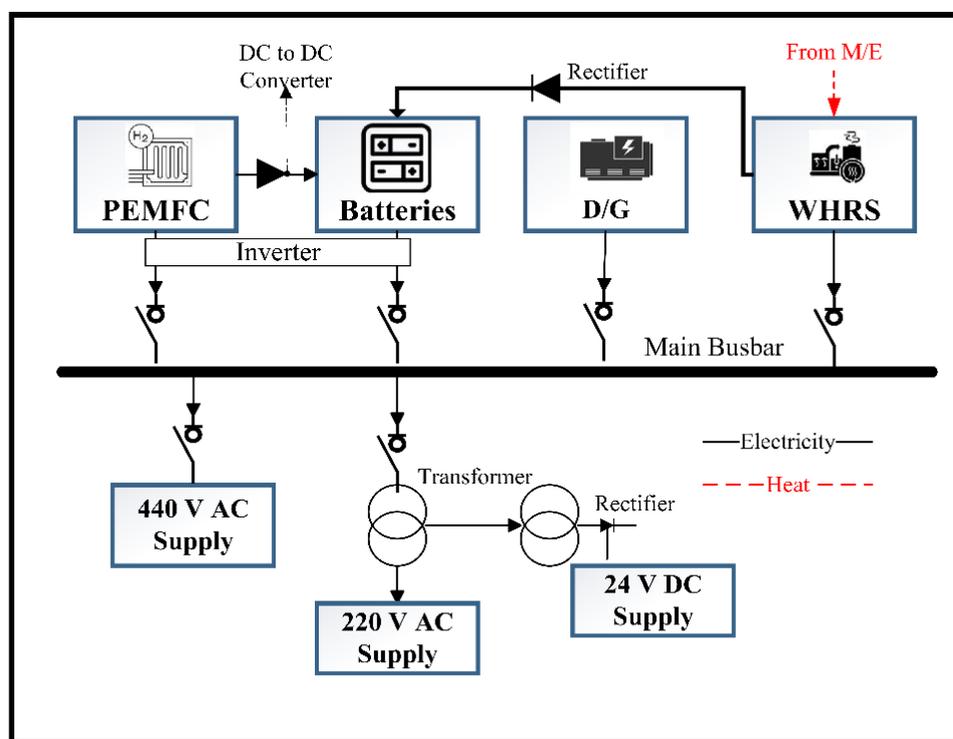


Figure 2. The general system scheme of the investigated hybrid configurations.

Battery sizes were adjusted according to FC sizes and the available average power from the ORC. In the context of the hybrid system, the battery sizes were increased based on the number of generators’ loads that needed to be met by the hybrid configuration. The dimensions of the batteries were determined to ensure that, after the FCs and WHRS fulfil the load profile requirements, any remaining demand can be adequately addressed by the batteries during emergency situations. These batteries can support the system for a minimum of one hour and, in certain cases, can provide extended support at a slower discharge rate, thereby enhancing overall system reliability.

This analysis examines two distinct hybrid designs. C1 features a larger FC system with an enhanced battery capacity that can support the load of two generators. This setup comprises four FC units, each rated at 800 kW, supplemented by a WHRS providing

average 197.01 kW and a battery storage capacity of 443 kWh. Together, this configuration effectively meets the combined load of the two generators, totalling 1440 kW.

Configuration C2 integrates an FC, battery, and WHRS to support a single generator's load of 720 kW. In this setup, two FCs share the load of 400 kW, while the WHRS consistently contributes an average additional 197.01 kW during navigation. The battery system supplies the remaining power requirements. It is important to note that the WHRS has a capacity of up to 449.47 kW at full ME load, equivalent to 354.41 kW at 85% ME load. These configurations were the most prominent in the benchmarking against other FC types and sizes within the electrification plant of the case study vessel, as indicated in the study of Yuksel et al. [33]. Figure 2 illustrates the PEMFC/battery/WHRs integrated system scheme in the marine power distribution system.

The batteries are charged by the WHRS utilising the exhaust waste heat from the M/E, in conjunction with PEMFCs, as illustrated in Figure 2. The direct current (DC) of PEMFCs and batteries is inverted to alternating current (AC) before being given to the grid. The ship electrification components' specifications and model details were provided in Yuksel and Koseoglu [37].

2.1. Ammonia Decomposition System

NH₃ cracking, or decomposition, generates H₂ from NH₃ over a catalyst at normal pressures and high temperatures. Without a catalyst, thermal processes that facilitate the NH₃ decomposition typically commence at temperatures exceeding 773 K [38]. Equation (1) indicates the endothermic reaction of NH₃ decomposition [38].



In contrast, the presence of a catalyst allows for catalytic cracking to predominantly occur at temperatures below 698 K, achieving an efficiency rate of approximately 98–99% [15,39]. The key factor is the high cracking purity of H₂ (99.9%) since the toxic and corrosive properties of NH₃ can harm the FC systems [38].

Nickel-based catalysts (e.g., Ni/Al₂O₃, Ni/MgO) are among the most widely studied and commercially used due to their cost-effectiveness and reasonable activity, particularly at temperatures above 650 K. These catalysts typically operate with a space velocity of 3000–10,000 h⁻¹ and exhibit a deactivation rate influenced by sintering and nitrogen poisoning. In contrast, ruthenium-based catalysts (e.g., Ru/Al₂O₃, Ru/CeO₂) have demonstrated superior activity at lower temperatures (as low as 550–600 K), often achieving near-complete NH₃ conversion at lower residence times and with enhanced thermal stability over prolonged operation [40].

The reactor model developed by Devkota et al. [41] in the Aspen Plus V1.2. and the cracker of Crystec [42] was used in the analysis. The feed NH₃ at 96.3 kg/h capacity, having 298 K and 10 bars, was utilised in the reactor [42]. The reactor model was validated against the experimental and simulated results reported by Devkota et al. [41], with key performance metrics such as NH₃ conversion and reactor outlet temperature aligning within a ±5% margin. The PSA unit performance was also cross-referenced with industrial data from Crystec [42], ensuring technical feasibility and scale relevance. Figure 3 indicates the simplified NH₃ cracker system scheme.

The required heat was met by the combustion of NH₃, which was 9% of the feedstock. The unused NH₃ was mixed with new fuel and air and introduced into the boiler via a pre-heater. A two-bed Temperature Swing Adsorption (TSA) unit separated the remaining NH₃ from the degraded H₂ and nitrogen gas mixture. The waste heat was extracted from the product and flue gas streams using the air–fuel mixture. Ultimately, more than 99.999% pure H₂ at 16.7 kg/h capacity can be produced with a NO_x emission generation

at 0.021 kg-NH₃/kg-H₂ by recirculating the flue gas via an eight-step, four-bed Pressure Swing Adsorption (PSA) machine with two pressure equalisations [15,41,42].

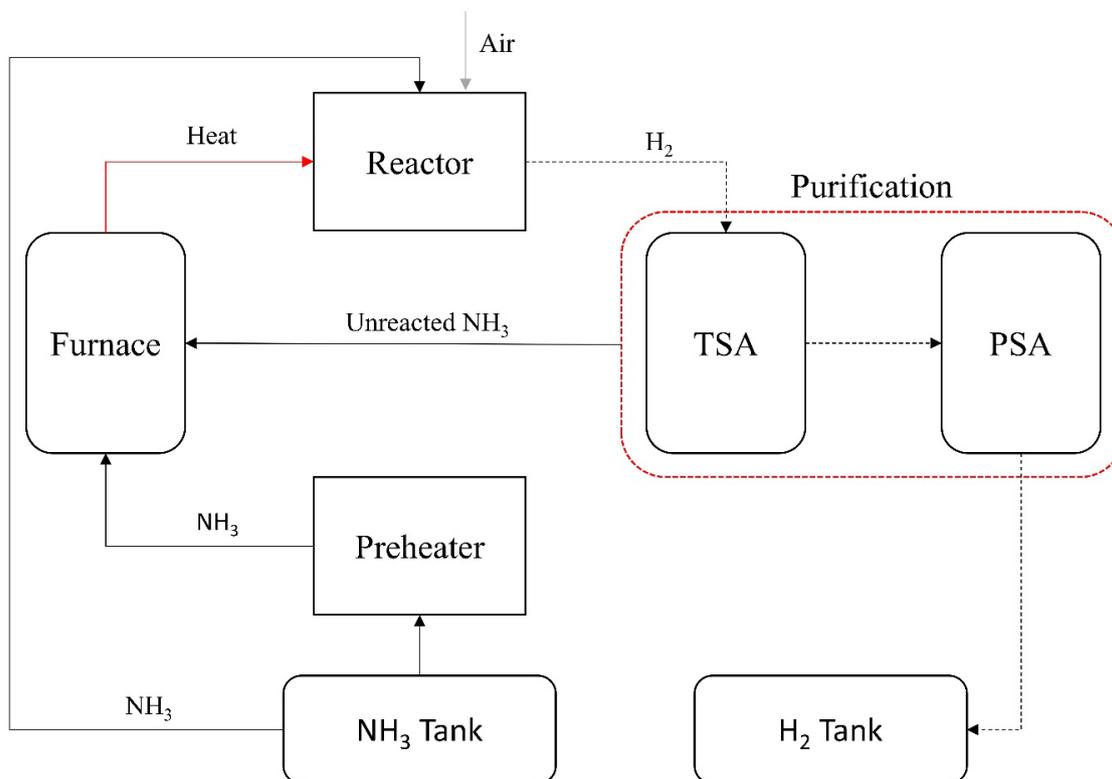


Figure 3. Basic schematic of NH₃ decomposition unit (adapted from [15,41,42]).

The reactor configuration typically involves a packed-bed tubular geometry with internal heating or external jacketed systems to compensate for the endothermic nature of the reaction. Maintaining a uniform temperature distribution in this configuration is critical, as local hot spots may lead to catalyst degradation or uneven NH₃ conversion profiles. Catalyst loading, bed porosity, and tube dimensions are optimised to ensure efficient mass and heat transfer throughout the reactor volume [40,41].

2.2. Ammonia Colour Coides

The Haber–Bosch process is one of the methods to produce NH₃ by combining nitrogen and H₂ through a thermocatalytic reaction [43]. When the manufacturing was performed using natural gas via steam-methane reforming (SMR), the NH₃ was labelled as “Grey”, and the process was assumed to emit 1.88 t-CO_{2e}/t-NH₃ in this study. “Blue NH₃” manufacturing includes a carbon capture system in the plant and reduces emissions to 1.23 t-CO_{2e}/t-NH₃ [44,45]. If the NH₃ has been produced utilising renewable/clean energy, it is named green NH₃, emitting 0.177 t-CO_{2e}/t-NH₃. Nuclear-sourced NH₃ is named pink in this study, having an emission level at 0.17 t-CO_{2e}/t-NH₃ [44,46].

The emission coefficients used in the analysis were derived from industry reports and peer-reviewed life cycle assessments reflecting current industrial NH₃ production processes. These sources were chosen to ensure realistic assumptions. While some variation may exist compared to site-specific data, the impact on overall results is expected to remain within an acceptable range of uncertainty.

2.3. Proton Exchange Membrane Fuel Cell

The PEMFC stack used in the study was the PowerCellGroup Marine System 200; its technical specifications are shown in Table 2 [36].

Table 2. Specifications of PEMFC.

Parameter	H ₂ Fuel Value	Unit
Power	200	KW
Output DC Voltage	580	V
Output Current	400	A
Electrical Efficiency	54% (peak)	-
Fuel Consumption	98.4	Nm ³ /h
Fuel Quality	Pure H ₂	-

The H₂ consumption of the PEMFC was determined using Equation (2) and performance curves, including the efficiency and SFC curves supplied by the manufacturer [36]. The design of the H₂ tank capacity was intended to provide support for two days, covering potential emergencies and maintenance needs of the decomposition plant. This calculation was predicated on the assumption that the hydrogen is stored in a compressed form at 75 kg/m³ [47,48].

$$H_2 \text{ Consumption (t)} = SFC \times P_{PEMFC} \times n_{PEMFC} \times t. \tag{2}$$

The operation time in hours is denoted as *t*, and SFC is interpolated from the curve. *P*_{PEMFC} and *n*_{PEMFC} represent the power and number of PEMFCs, respectively. Equation (3) indicates the emissions of PEMFC.

2.4. Battery Cell

A Panasonic NCR18650GA battery cell with a lithium-ion chemistry capacity of 3.45 Ah was utilised in the stack formation. Lithium-ion batteries, known for their higher specific energy and negligible memory effect, were employed in hybrid operation scenarios [48]. The constant-current constant-voltage charging strategy was employed, and the battery set was modelled accordingly. The cell’s nominal voltage is 3.6 V, and voltage drop was simulated using SoC-voltage curves from the manufacturer’s datasheet [49]. Battery health degradation, represented by state of health (SoH), was obtained from the datasheet.

An energy management strategy (EMS) determined the operational state of batteries and the available SoC. The time-dependent SoC is computed by employing the Coulomb counting method shown in Equation (3) [50].

$$SoC(t) = SoC(0) - \int_0^t \frac{I(t) \times \eta_C}{C_B \text{ (Ah)}}. \tag{3}$$

SoC (0) indicates the initial state of charge at the commencement of the simulation, whereas SoC(*t*) signifies the revised SoC at *t*. The EMS keeps the SoC within the range of 20% to 80% during active use to reduce internal resistance, which in turn supports the health and longevity of the battery [50]. The coulombic efficiency (η_C) is taken at 1, with the charging or discharging current denoted as *I*(*t*), and the available battery capacity in ampere-hours (Ah) represented by *C_B* in the equation [51].

The reduction in the battery’s capacity was computed throughout its operation, incorporating adjustments influenced by the C-rate. The SoH was determined using Equation (4),

which defines SoH as the ratio of the actual battery capacity (C_a) after degradation to the initial capacity (C_{in}) at the start of operation.

$$\text{SoH} (t) = \frac{C_a \text{ (Ah)}}{C_{in} \text{ (Ah)}}. \tag{4}$$

The SMA Sunny SCS2900 inverter model (SMA, Niestetal, Germany) transforms the DC from the battery to AC for integration into the ship’s grid. This inverter operates at an efficiency of 98.4% when converting from 800 V DC to 450 V AC [52].

2.5. Waste Heat Recovery System

An ORC-based WHRS was utilised to generate electricity from the heat of the M/E exhaust waste. The efficiency of the ORC model (η_{ORC}) was adapted at 13.2% from the studies of Konur et al. [53], Konur et al. [54]. The exhaust gas temperature after steam production (T_{in}) in °C and exhaust mass flow rate (\dot{m}_{ex}) in kg/s regarding the engine load (%), power (kW), and engine speed (rpm) are indicated in Table 3.

Table 3. The exhaust gas, power, and speed data of the M/E.

Load (%)	Power (kW)	Speed (rpm)	\dot{m}_{ex} (kg/s)	T_{in} (°C)
25	2483	56.9	8.7	190
35	3476	63.7	10	194
50	4965	71.8	13.9	217
71.6	7110	80.9	15.1	205
75	7448	82.1	19.4	208
100	9930	90.4	23.7	235

The simulation’s exhaust data, presented in Table 3, was utilised through interpolation based on the acquired real-time M/E data. Equation (5) calculates the generated power by the WHRS (\dot{W}_{WHRS}) in kW [55].

$$\dot{W}_{\text{WHRS}} = \dot{m}_{ex} \times (T_{in} - T_{out}) \times C_p \times \eta_{\text{ORC}}. \tag{5}$$

T_{out} represents the exit temperature from the ORC, set at 100 °C, while C_p denotes the specific heat capacity of the exhaust gas at constant pressure. This value, taken as 1.089 kJ/kgK, reflects the thermal energy exchanged by a unit mass of exhaust gas per unit temperature change [56].

2.6. Marine Diesel Generators

The simulation determined the required engine power and active generators at 1-min intervals. Generator load and power were simulated for hybrid scenarios, with SFC interpolated from D/G datasheets. Based on the approach of Yuksel and Koseoglu [24], load sharing activated additional generators when power exceeded 85% capacity, balancing load by frequency and power factors. Fuel consumption was calculated by multiplying each generator’s SFC, required power, and operation time.

2.7. Environmental Model

The emissions from the hybrid configurations and base scenario were computed using the emission coefficients. Table 4 demonstrates the upstream emission coefficients (UEC) and operational emission coefficients (OEC) to calculate operational emissions (OEs) and upstream emissions (UEs).

Table 4. UEC and OEC.

UEC (g UE/g OE)						
Fuel	CO ₂	N ₂ O	CH ₄	NO _x	SO _x	Reference
HFO	0.147	0.004	0.879	0.010	0.102	[57,58]
LNG	0.131	0.004	0.879	0.007	* 0.158	[57–59]
H ₂	* 110.902	0.000	0.000	* 0.191	* 0.141	[57,58]
OEC (g emission/g fuel)						
Fuel	CO ₂	N ₂ O	CH ₄	NO _x	SO _x	Reference
HFO	3.114	0.00015	0.00006	0.903	0.025	[60]

* Unit is g/kWh.

This study adopted a streamlined well-to-wake approach, focusing on the operational and production phase of the fuel pathways and excluding the energy consumption and carbon emissions associated with the manufacturing of FCs and battery systems. While this allows for a consistent and comparable analysis of fuel-related emissions, it introduces limitations, particularly for battery systems, which can involve significant embodied energy and associated emissions during production. The manufacturing phase for FCs and batteries can lead to a deviation of around 20–30% higher emissions compared to the operational phase alone [61]. Upstream and operational emissions were weighted equally, utilising emission coefficients sourced from the literature.

The 100-year global warming potential (GWP100) was demonstrated in the calculation of CO_{2e}, as shown in Equation (6) [62].

$$CO_{2e} = CO_2 + 265 \times N_2O + 28 \times CH_4. \tag{6}$$

The coefficients for CO_{2e} from the 2024 IMO life cycle assessment guidelines [62] were utilised to compute the GWP of nitrous oxide (N₂O) and CH₄. These gases had GWPs approximately 260–273 times for N₂O and 27–30 times for CH₄ greater than CO₂, respectively [58]. The CO_{2e} coefficients for NH₃ production methods are provided in Section 2.2.

2.8. Economic Model

To reduce uncertainty, a scenario-based approach was employed to calculate the LCOE for 2025 and 2040, incorporating projected energy and carbon prices from reliable sources, including currently limited applications, like the Emission Trading System, and academic literature. The primary aim was to illustrate how variations in price levels influence system performance, particularly identifying the carbon price threshold at which renewable energy options become competitive.

Among various methodologies available for techno-economic evaluation, LCOE was selected due to its ability to provide a single, consistent indicator of economic performance across different fuel and technology configurations. It allows for directly comparing energy generation options by accounting for all relevant costs, capital, operational, maintenance, and fuel over the system’s lifetime. Moreover, LCOE is one of the most widely used and recognised metrics in energy system analysis, which enhances the comparability and relevance of the findings within the broader academic and industry discourse [63].

The LCOEs was used to gauge each configuration’s economic performance. The main instrument of choice for evaluating the unit costs of various baseload technologies at the

plant level throughout their operating lives is LCOE [64]. Equation (7) shows how the LCOE is calculated [65].

$$LCOE \left(\frac{USD}{kWh} \right) = \frac{\sum_{t=1}^{LT} \frac{(C_{plant} + C_{fuel} + C_{o\&m}) (USD)}{(1+r)^n}}{\sum_{t=1}^{LT} \frac{P_{gen} (kW) \times t(h)}{(1+r)^n}} \tag{7}$$

The installation cost is C_{plant} , the fuel cost is C_{fuel} , and operation and maintenance costs are represented by $C_{o\&m}$ in Equation (7). The plant’s generated total power (P_{gen}) and operation time (t) were calculated from the operational data. Plant lifetime (LT) was assumed to be 20 years, and the discount rate (r) was taken at 10% [64,66]. Table 5 demonstrates the C_{plant} and $C_{O\&M}$ of the equipment.

Table 5. Lifetime, installation, and operational costs of systems.

Equipment	C_{plant} —2025	C_{plant} —2040	Unit	$C_{o\&m}$	LT (Years)	Reference
PEMFC	2540.46	1304.04	USD/kW	1.50%	10 years or 40,000 h	[67,68]
MDE	605,997	605,997	USD	1.50%	20	[69]
HFO Storage	1497	1497	USD/m ³	1%	20	[69]
NH ₃ Storage	3,145,236	2,028,677	USD	1%	20	[70]
NH ₃ Cracker	2,648,824	1,708,491	USD	6%	20	[71,72]
H ₂ Storage	460	243.5	EUR/kg	1%	20	[73]
H ₂ P/T	2440	2440	EUR/kW	4%	10	[73]
NH ₃ Catalyst/Adsorber	31,083	31,083	USD	N/A	N/A	[72]
ORC WHRS	1,348,579	1,348,579	USD	1.50%	20	[53]
Battery Cell	11.86	11.86	USD	1%	Usage dependent	[74]

The calculation of $C_{o\&m}$ was done by multiplying the percentage by C_{plant} , and the analysis was carried out using the United States dollar (USD). The pound (GBP) and Euro (EUR) to USD exchange rate were set at 1.27 and 1.1, respectively. The rates from the relevant data year and the most recent Chemical Engineering Plant Cost Index (CEPCI) were used to adjust the prices from prior years. According to Maxwell [75], the most recent CEPCI was 798.8 for June 2024 (the most recently announced value). Table 6 indicates the C_{fuel} and carbon price assumptions for 2025 and 2040.

Table 6. Carbon price and C_{fuel} (USD/kg—Fuel) for 2025 and 2040 regarding low/high cases.

Fuel	2025	2040 Low	2040 High	Reference
Grey NH ₃	0.229	0.229	0.229	[46]
Blue NH ₃	0.372	0.100	0.250	
Green NH ₃ by Nuclear	0.918	0.075	0.225	[46,76]
Green NH ₃ by Renewable	1.055	0.222	0.480	
Green H ₂	3.750	2.334	3.144	[77]
HFO	0.663	0.854	0.971	
Carbon Price	0.077	0.155	1.285	[78–80]

The current C_{fuel} of HFO was obtained from ShipandBunker [79] for global average. Fuel price data used in this study were obtained from reputable literature sources to ensure consistency with established projections and reflect the most credible and widely accepted estimates. Given the inherent uncertainties of future fuel pricing, both low and high projection cases were adopted. These scenarios were defined based on boundary conditions derived from fuel price trends, offering a more comprehensive understanding of potential

economic outcomes. A similar approach was applied to the carbon pricing collected from the literature by fuel price projections.

2.9. Multi-Criteria Decision-Making

The Complex Proportional Assessment (COPRAS) method has effectively ranked the alternatives. This method facilitates a clear and comprehensive comparison by evaluating one alternative’s relative merits and drawbacks against others [81,82]. It ensures the optimisation of criteria influenced by multiple factors by considering the utility level and relative importance [83].

The method relies on linear normalisation, standardising diverse criteria by converting them to a uniform scale, thus facilitating direct comparisons. Additionally, it integrates the weighted significance of each criterion, allowing decision-makers to express and incorporate their preferences and priorities into the evaluation framework [84]. This method is frequently utilised in decision-making contexts characterised by uncertainty or ambiguous outcomes [81,85].

The assessment criteria were determined as total CO_{2e}, total other emissions (summation of SO_x and NO_x), and LCOE. Since the LCOE depends on the years and economic projection scenarios, the ranking regarding 2025, 2040 Low, and 2040 High cases were assured. The entropy method introduced by Shannon [86] determined criterion weights, which encompass three stages. The initial step is to normalise the decision matrix using Equation (8) [87].

$$r_{ij} = \frac{f_{ij}}{\sum_{t=1}^n f_{ij}} \tag{8}$$

where f_{ij} depicts the data point to be normalised, n is the number of criteria, and r_{ij} represents the normalised data. The next stage is the calculation of entropy (e_j) employing Equation (9) [87].

$$e_j = \frac{1}{\ln(n)} \times \sum_{i=1}^m \ln(r_{ij}), j = 1, 2, \dots, n. \tag{9}$$

where m represents the total row number. The final stage computes the weights (w_j) utilising Equation (10) [87].

$$w_j = \frac{1 - e_j}{\sum_{i,j=1}^n 1 - e_j} \tag{10}$$

After determining the weights, the normalised matrix was multiplied by the relative w_j , and the weighted normalisation matrix (N_{ij}) was obtained for the COPRAS. Using Equations (11) and (12), the beneficial (B_i) and cost indexes (C_i) were obtained [88].

$$B_i = \sum_{j=1}^n N_{ij}, i = 1, \dots, k \text{ beneficial criteria}, \tag{11}$$

$$C_i = \sum_{j=k}^n N_{ij}, i = k + 1, \dots, m \text{ cost criteria}. \tag{12}$$

The relative significance of alternatives (Q_i) was calculated employing Equation (13) [89].

$$Q_i = B_i + \frac{\min(C_i) \times \sum_{i=1}^n C_i}{C_i \times \sum_{i=1}^n \frac{\min(C_i)}{C_i}} \tag{13}$$

The final stage of the COPRAS involves the calculation of utility degrees (UD_i) utilising Equation (14) and ranking determination using UD_i values [89].

$$UD_i = \frac{Q_i}{\max(Q_i)} \times 100\%. \tag{14}$$

A higher UD_i means a higher ranking for the configuration [88]. The validity of the COPRAS methodology has been assessed using a sensitivity analysis technique as proposed by Triantaphyllou and Sánchez [90].

Let δ' _{k,i,j} (for 1 ≤ I < j ≤ m and 1 ≤ k ≤ n) represent the minimum adjustment required in the assigned weight W_k of criterion k, to achieve a reversal in the ranking positions of alternative A_i and A_j, as illustrated in Equation (15) [90].

$$\delta'_{k,i,j} < \frac{P_j - P_i}{a_{j,k} - a_{i,k}} \times \frac{100}{w_k}, \text{ if } a_{j,k} > a_{i,k}, \delta'_{k,i,j} > \frac{P_j - P_i}{a_{j,k} - a_{i,k}} \times \frac{100}{w_k}, \text{ if } a_{j,k} < a_{i,k}. \tag{15}$$

P_j and P_i represent the weighted normalised decision matrix elements for the respective rows in this context. The normalised matrix values are defined as a_{j,k} and a_{i,k}. The following condition in Equation (16) should be met for the value of δ' _{k,i,j} [90].

$$\frac{P_j - P_i}{a_{j,k} - a_{i,k}} \leq w_k. \tag{16}$$

No weight adjustment can make A_j rank higher when alternative A_i consistently outperforms A_j across all criteria (a_{ik} ≥ a_{jk} for every k). A criterion is classified as redundant if altering its weight has no impact on the rankings of any alternatives, allowing it to be eliminated from consideration [90].

2.10. Uncertainty Analysis

Uncertainty serves as a metric for the validity of results and is vital for assessing the appropriateness of data in informed decision-making across various domains [91]. Statistical techniques, particularly uncertainty analysis, are instrumental in pinpointing scenarios impacted by uncertainty while improving data accuracy [92]. Varying levels of uncertainty are integrated using Equation (17) [93].

$$U_R = \sqrt{\left[\left(\frac{\delta R}{\delta x_1} U_1 \right)^2 + \left(\frac{\delta R}{\delta x_2} U_2 \right)^2 + \dots + \left(\frac{\delta R}{\delta x_n} U_n \right)^2 \right]}. \tag{17}$$

In Equation (17), U values denote the partial uncertainties of individual parameters (x₁, x₂, . . . x_n), U_R demonstrates the uncertainty of the merged calculation, while R is the utilised parameter for each independent metric [94].

An uncertainty analysis of the fuel consumption model was conducted to assess its reliability. Two main sources of uncertainty were identified. Initially MDEs' fuel usage was calculated from the SFC data provided by the manufacturer for proposed systems. In the baseline scenario, the discrepancy between the sensor data and the model outputs was 5.63%. Moreover, the fuel consumption efficiencies exhibited a 2% error margin reflected in the accompanying data sheets. Utilising Equation (17), the overall uncertainty calculated by the model was determined to be 5.98%.

3. Results

The environmental analysis highlights variations in CO_{2e}, NO_x, and SO_x emissions for systems C1 and C2 using grey, blue, pink, and green NH₃ as cracker feedstock. Emissions

from the decomposition system influenced upstream and operational pollutant levels. NH_3 production emissions were excluded, focusing only on those required by maritime regulations for this study.

The C1 and C2 designs required 304 and 298.12 t of H_2 to run PEMFCs for 1.96 years. The current HFO tank capacity (467 m^3) powers the D/Gs for 180 days. To maintain the same bunkering capacity with H_2 , the required tank sizes would be 1199.96 m^3 and 1176.76 m^3 , respectively. These values can be reduced to 650.10 m^3 and 624.62 m^3 with NH_3 cracking, including 40.11 m^3 and 39.33 m^3 H_2 storage tanks for two days. The NH_3 requirements of the decomposition system to produce the same amount of H_2 on board were set at 1653.43 t for C1 and 1586.47 t for C2. Limiting the capacity to the base value of 467 m^3 , the NH_3 tanks would provide fuel for 125.98 and 131.53 days, respectively, including two days of H_2 storage requirements given earlier. The economic analysis was conducted based on the 180-day values. Figure 4 illustrates the hourly power output from the PEMFC plants in both C1 and C2 configurations, along with the corresponding hourly power availability of the WHRS system.

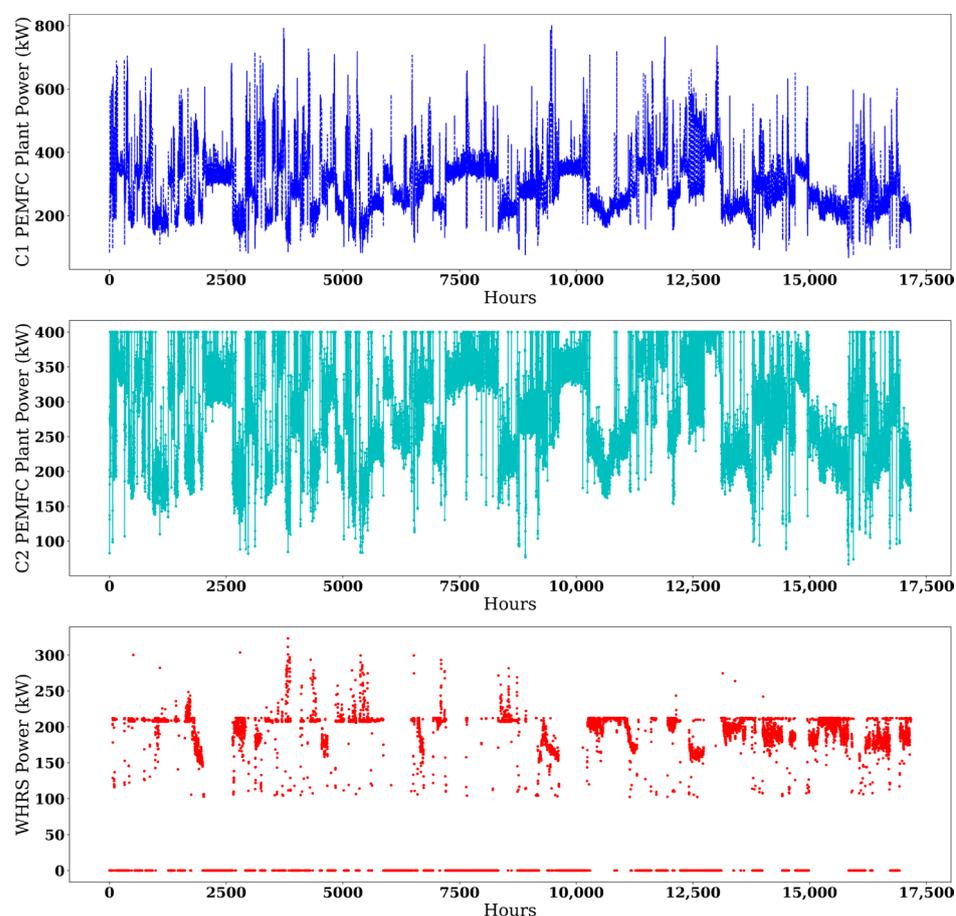


Figure 4. Power outputs from the PEMFC and WHRS in the examined configurations.

The total electrification operation time reached 17,169.6 h. C1 operates without utilising the MDE, resulting in zero HFO usage, and relies mainly on PEMFCs, with minimal battery usage limited to just 5.38 h. During its operation, C1 primarily employed two FCs, which accounted for 75.74% of the total operational time (13,003.72 h). A single PEMFC was used for 14.20% of the time (2438.63 h), while three and four PEMFCs were utilised for 9.11% (1563.70 h) and 0.94% (161.43 h), respectively.

C2 uses MDEs for 1262.88 h, consuming 37.84 tons of HFO, and relies on batteries for 1868.32 h. During its operation, 85.86% of the total runtime (14,741.73 h) was conducted

using two PEMFCs, with the remainder carried out by a single PEMFC. The life spans of PEMFCs were calculated as 9.28 years and 4.91 years for C1 and C2. Since the C1 configuration rarely involves the battery for operations and uses it only for emergency responses, Figure 5 illustrates the changes in SoC and SoH over operational hours.

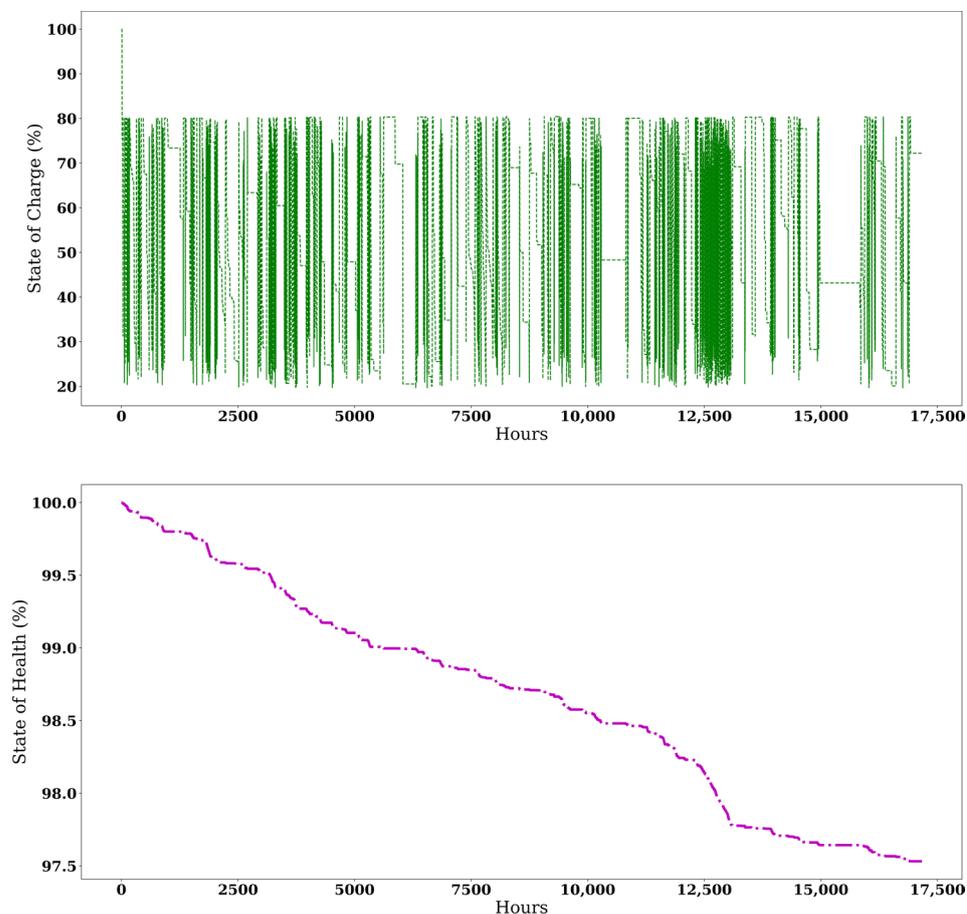


Figure 5. Variation of battery SoC and SoH during the operation of C2.

Based on the usage profile and the final SoH indicated in Figure 5, the estimated battery lifespan for both configurations was calculated over ten years, assuming that 80% SoH marks the end of battery life. In the economic analysis, battery replacements are assumed to occur every ten years. Figure 4 illustrates the CO_{2e} of each configuration regarding NH₃ colour codes for 1.96 years of operation time.

In Figure 6, the emissions represented by the striped bars correspond to operational emissions, while the solid bars indicate upstream emissions. The conventional configuration emitted 5752.82 tonnes of CO_{2e} during operation and 834.56 tonnes upstream, yielding a total of 6587.39 tonnes over 1.96 years. C1 and C2 configurations reduced it by 91.57% and 89.99% by using green H₂-BS. C2 exhibited operational emissions of 119.39 tonnes.

For grey NH₃ decomposition, operational emissions remained unchanged compared to H₂-BS, but upstream emissions significantly increased compared to both the base case and C1/C2 H₂-BS. Upstream CO_{2e} emissions for C1 were 3113.41 tonnes with zero operational CO_{2e}, while for C2, operational CO_{2e} emissions were 119.39 tonnes, and upstream emissions rose to 3127.59 tonnes. This represents CO_{2e} emissions 5.61 times higher for C1 and 4.74 times higher for C2 than their H₂-BS cases. However, grey NH₃ decomposition with C1 and C2 configurations reduced CO_{2e} by 52.74% and 52.52%, respectively, compared to the MDE configuration.

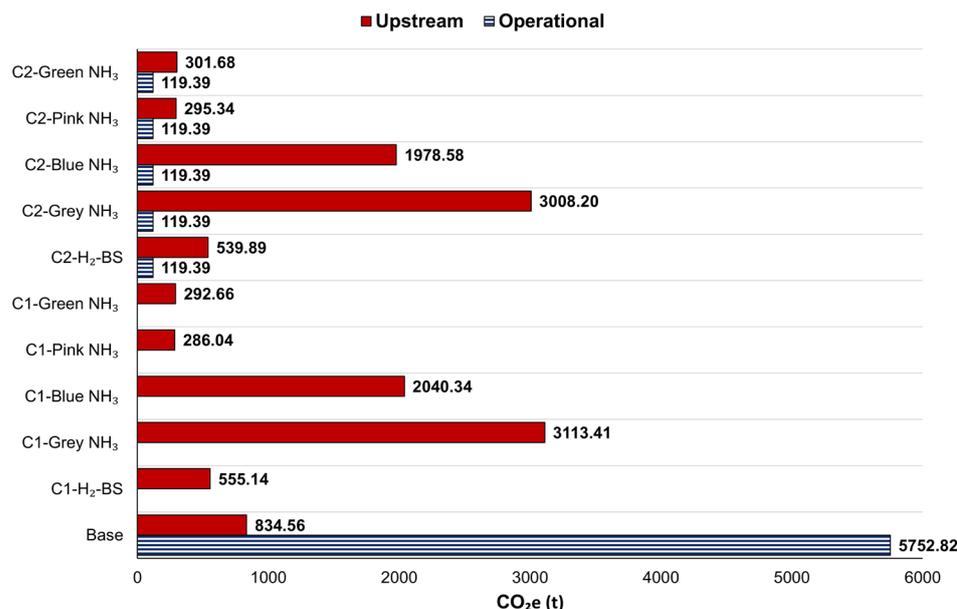


Figure 6. CO_{2e} of configurations regarding the NH₃ production method.

Blue NH₃ cracking reduced upstream CO_{2e} emissions, with C1 and C2 emitting 2040.34 and 1978.58 tonnes, respectively. Compared to the H₂-BS, total CO_{2e} emissions increased 3.68 times for C1 and 3.18 for C2. When benchmarked against the base case, the reduction rates were 69.03% for C1 and 68.15% for C2.

The use of green NH₃ for onboard H₂ production resulted in CO_{2e} emissions of 292.66 tonnes for C1 and 421.07 tonnes for C2. These values represented 47.28% and 36.13% reductions for C1 and C2 configurations, respectively, compared to their green H₂-BS counterparts. Additionally, these emissions corresponded to 95.56% and 93.61% decreases, respectively, compared to the base scenario.

Pink NH₃ decomposition achieved further reductions, with CO_{2e} emissions of 286.04 tonnes for C1 and 414.72 tonnes for C2. These emissions represented 48.47% and 36.13% reductions compared to the H₂-BS configurations and 95.66% and 93.70% compared to the base scenario. Figure 7 demonstrates the SO_x emissions from using different H₂ sources in PEMFCs for 1.96 years.

In the base case, SO_x emissions were at 45.98 tonnes from operational activities, 4.65 tonnes from upstream processes, for a total of 50.23 tonnes, as shown in Figure 7. The C1 and C2 designs, utilising outsourced H₂, emitted 5.73 and 6.52 tonnes of SO_x, representing significant reductions of 88.59% and 87.02%, respectively.

SO_x emissions for grey and blue NH₃ were identical. Using grey or blue NH₃ generally increased SO_x emissions compared to the H₂-BS case. C1 and C2 emitted 6.61 and 7.39 tonnes of SO_x, corresponding to increases of 15.37% and 28.88%, respectively. Despite this, PEMFCs powered by H₂ derived from blue or grey NH₃ under the C1 and C2 designs achieved SO_x reductions of 86.83% and 85.29% compared to the MDE utilisation scenario.

For green NH₃, SO_x emissions were identical regardless of whether nuclear or renewable energy sources were used. SO_x emissions were calculated at 3.31 tonnes for C1 and 4.22 tonnes for C2, equivalent to decreases of 93.42% and 91.61% compared to the base case. Additionally, reductions of 42.31% for C1 and 26.47% for C2 were observed relative to the H₂-BS case. Figure 8 depicts the NO_x emission comparison between NH₃ decomposition system usage and outsourcing H₂.

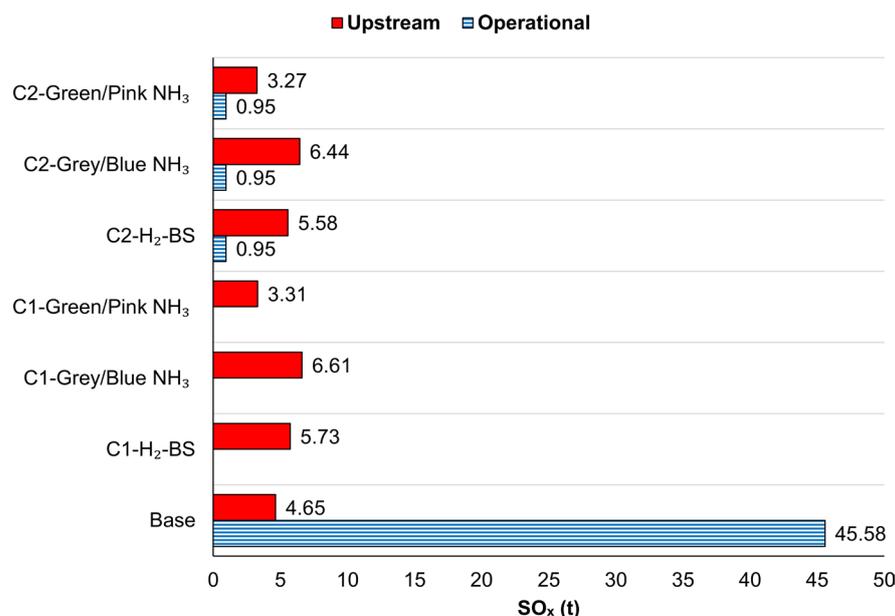


Figure 7. SO_x resulting from configurations regarding the NH₃ production method.

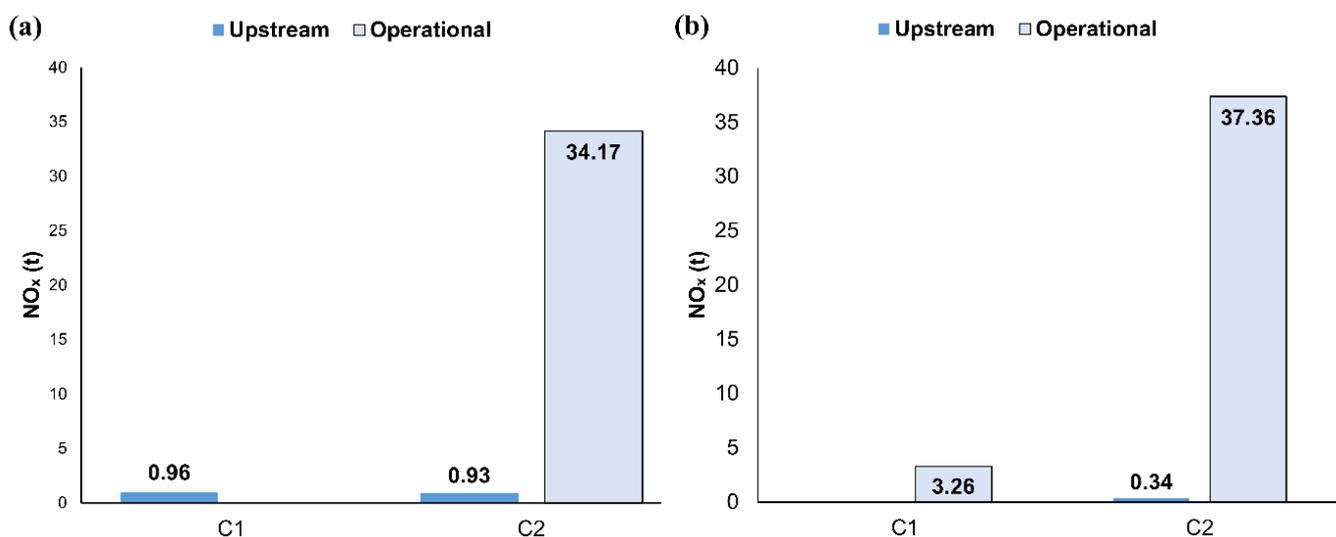


Figure 8. NO_x resulting from (a) H₂-BS, (b) H₂ derived from NH₃ decomposition onboard.

The upstream and operational NO_x emissions for the base case were calculated at 1646.31 tonnes and 16.46 tonnes, respectively, resulting in a total of 1662.77 tonnes. In the H₂-BS case, C1 emitted only 0.96 tonnes of upstream NO_x emissions, as shown in Figure 8a, corresponding to a 99.94% reduction in total NO_x emissions compared to the base scenario. For C2, total NO_x emissions amounted to 37.70 tonnes, representing a 97.89% reduction. This difference is attributed to the limited consumption of HFO during C2 operations.

The NH₃ cracking system uses heat generated by burning NH₃ for decomposition, leading to a slight increase in operational NO_x emissions. The additional NO_x emissions from this process were calculated at 3.26 tonnes for C1 and 3.19 tonnes for C2. However, since H₂ is no longer outsourced, upstream NO_x emissions were reduced to zero for C1 and 0.34 tonnes for C2, as illustrated in Figure 8b.

For C1, total NO_x emissions increased by 2.3 tonnes compared to the H₂-BS case, but a 99.8% reduction was still achieved relative to the base scenario. A similar trend was observed for C2, where total emissions increased by 7.43% compared to the H₂-BS case, achieving a 97.73% reduction compared to the base configuration. The economic

performance of various NH₃ feedstocks and the H₂-BS case was evaluated by calculating the LCOE as shown in Figure 9 for 2025 and 2040 under low and high projection scenarios.

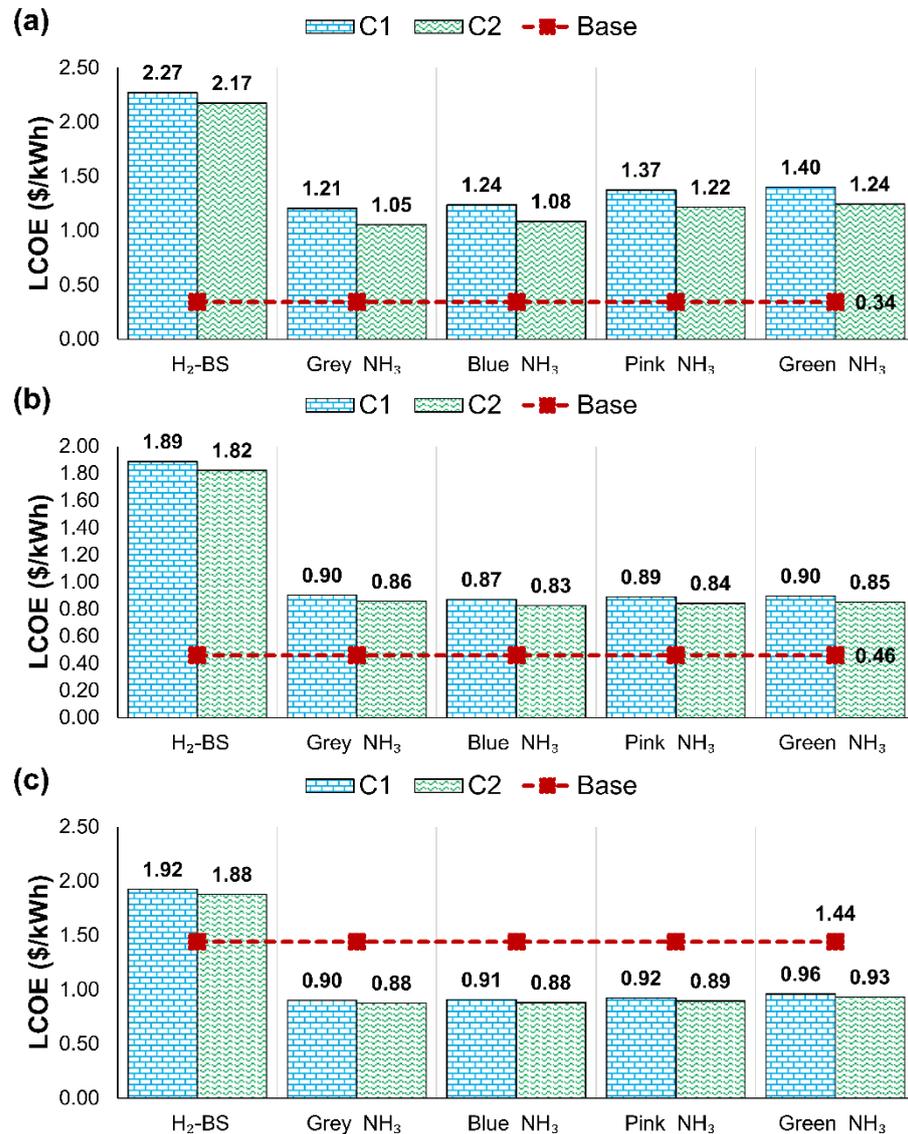


Figure 9. LCOE of different NH₃ feedstocks and H₂-BS for (a) 2025, (b) 2040-low, and (c) 2040-high.

The marine power distribution plant generated 3443.51 MWh of energy annually. The LCOE values presented in Figure 9 were derived by calculating the annual fuel and operation costs and the total investment costs in Appendix C. The base scenario was presented only with H₂-BS, as it remained unchanged in other calculations. Overall, the NH₃ decomposition system was a more economically viable option over the H₂-BS due to the decreased storage costs and lower fuel prices.

In 2025, the base scenario had a LCOE at USD 0.34 per kWh, and the PEMFC scenarios performed worse economically. The most feasible option was C2 with the grey NH₃ cracking system onboard at 1.05 USD/kWh, as shown in Figure 9a. Blue NH₃ with the C2 had a close performance at 1.08 USD/kWh, and considering its environmental superiority over grey NH₃ feedstock, it could be a more attractive option. A similar trend was observed for C1 with a higher level of LCOEs over 1.21 USD/kWh.

Due to the slightly increased carbon prices in the 2040 low-economic scenario, as illustrated in Figure 9b, the LCOEs of NH₃ decomposition scenarios became more viable options. The base-case LCOE rose to 0.46 USD/kWh, whereas the LCOE for H₂-BS de-

creased to USD 1.89 and USD 1.82 per kWh for scenarios C1 and C2, respectively. Among the decomposition pathways, blue NH₃ decomposition offered the lowest LCOE, at 0.83 and 0.87 USD/kWh, followed closely by pink NH₃ cracking.

The high economic projection for 2040, as illustrated in Figure 9c, identified the base case as the worst performer, with an LCOE of USD 1.44 per kWh, which was significantly higher than NH₃-cracking configurations due to the increased carbon prices. The H₂-BS system remained the costliest among all options. Grey and blue NH₃ decomposition under scenario C2 achieved an LCOE of USD 0.88 per kWh, while pink NH₃ reached USD 0.89 per kWh. For scenario C1, the same feedstocks resulted in LCOEs ranging between USD 0.90 and USD 0.92 per kWh. These findings underscore that only implementing higher carbon taxes makes NH₃ decomposition scenarios more cost-competitive than fossil fuel utilisation.

An MCDM analysis using the COPRAS methodology integrated with the entropy weighing method was conducted to combine the environmental and economic studies. The CO_{2e} and other emissions as the environmental indicator and LCOE as the financial performance index were taken in the analysis. Table 7 indicates the criteria weights based on the economic projection scenario and year.

Table 7. Criteria weights regarding the Entropy Method.

Scenario	CO _{2e}	Other Emissions	LCOE
2025	0.386	0.134	0.480
2040 Low	0.386	0.134	0.480
2040 High	0.384	0.134	0.482

The entropy method successfully assigned weights consistent with the criteria. LCOE received the highest weight, ranging from 0.478 to 0.480, followed by CO_{2e}, which varied between 0.384 and 0.386, and other emissions at 0.134. The LCOE accounted for 48% of the total weight, while the environmental metrics collectively represented 52%. The detailed results of the entropy method are demonstrated in Appendix D. Table 8 shows the outcomes of the COPRAS calculation stages.

Table 8. COPRAS outcomes.

Configuration	C _i 2025	C _i 2040 Low	C _i 2040 High	Q _i 2024	Q _i 2040 Low	Q _i 2040 High	UD _i 2025	UD _i 2040 Low	UD _i 2040 High
Base	0.26	0.48	0.30	0.03	0.02	0.03	19.78	7.85	13.07
C1 H ₂ -BS	0.09	0.08	0.08	0.08	0.10	0.10	59.27	48.59	48.92
C1 Grey NH ₃	0.10	0.09	0.15	0.07	0.08	0.06	50.19	39.99	26.40
C1 Blue NH ₃	0.08	0.07	0.11	0.09	0.10	0.08	62.58	52.27	36.66
C1 Pink NH ₃	0.05	0.04	0.04	0.14	0.20	0.21	99.50	100.00	100.00
C1 Green NH ₃	0.05	0.04	0.04	0.14	0.20	0.21	97.58	98.94	96.62
C2 H ₂ -BS	0.09	0.08	0.08	0.08	0.09	0.10	58.46	44.79	47.35
C2 Grey NH ₃	0.10	0.10	0.15	0.07	0.07	0.06	51.36	37.75	26.44
C2 Blue NH ₃	0.08	0.08	0.11	0.09	0.10	0.08	63.73	47.96	36.17
C2 Pink NH ₃	0.05	0.05	0.04	0.14	0.16	0.19	100	82.80	90.42
C2 Green NH ₃	0.05	0.05	0.05	0.14	0.16	0.19	98.12	81.98	87.66

The computation of B_i was excluded, and it was taken as zero since all the criteria were the non-beneficial or cost type in the analysis. For instance, the Q_i was calculated for C1 grey NH₃ by computing the ratio of minimum C_i of the criterion and C_i, which was found at 0.10 in 2025. Then, applying Equations (11) and (12), Q_i and UD_i were calculated

at 0.07 and 50.19 for 2025 in Table 8. Figure 10 illustrates the resulting rankings based on the UD_i values shown in the table.

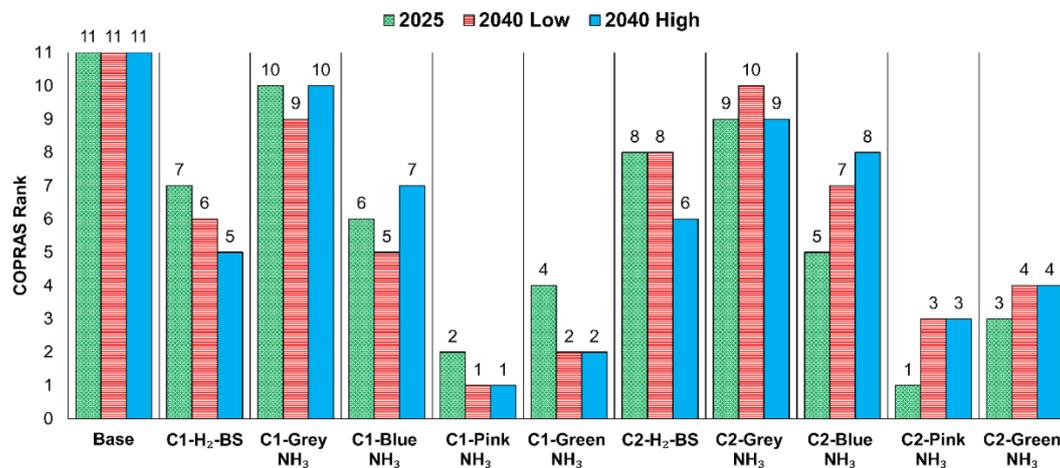


Figure 10. COPRAS ranks of configurations.

The base configuration consistently placed 11th across all economic cases. Similarly, in every economic scenario, the use of grey NH₃ for onboard H₂ production with the C1 and C2 configurations held the 10th and 9th positions, respectively. H₂-BS also performed poorly, with C1 and C2 designs falling between 7th and 8th. Overall, in 2025, the C2 pink NH₃ combination was identified as the most favourable option. However, by 2040, the top position shifted to the C1 pink NH₃ combination. Additionally, by 2040, C1 green NH₃ rose to second place, overtaking C2 pink NH₃, which had previously held that position in 2025.

The sensitivity analysis of COPRAS evaluated 495 instances, classifying them as feasible (F) or non-feasible (NF). Of these, 460 cases were deemed F, while 35 are NF, resulting in a feasibility rate of 92.93%. Among the 460 F instances, 93 were sensitive to changes (with a change rate below 5%), while 367 were resistant to weight changes, demonstrating the validity of the COPRAS method in this analysis. Further details can be found in Appendix E.

4. Discussion

The evaluated hybrid electrification systems reduced ship CO_{2e} emissions to 15.83% with pink NH₃ decomposition. It should be noted that the analysis can involve the FC/battery production phase emissions, elevating CO_{2e} emissions by 20–30%. However, this increase is offset by a substantial overall reduction in emissions compared to conventional systems [61]. The potential for emissions reduction increases to approximately 50% with the implementation of zero-carbon electrification alongside LNG-DF engines in the propulsion unit. Furthermore, employing LNG-fuelled DF engines within diesel-electric propulsion systems can lead to a CO_{2e} emission decrease of up to 20% while adhering to the EEXI and CII requirements [95]. The turbine technologies demonstrate greater carbon reduction potential than DF engines. Incorporating carbon-free fuels like NH₃ or H₂ in their operation is essential for sustainable energy transitions [96]. However, currently available alternative fuels such as LNG or methanol do not meet the 2050 decarbonisation targets. Bridging this gap requires advanced energy systems and a shift to carbon-neutral fuels [69]. FCs for ship electrification present a favourable avenue for enhancing overall ship energy efficiency when combined with DFs or turbines. Additionally, green H₂ can be regarded as a promising alternative fuel for the shipping industry, facilitating significant emissions reductions and promoting long-term sustainability [96].

The H₂-BS scenario has demonstrated inferior performance compared to the NH₃ decomposition process for H₂ generation on board, both economically and environmentally, particularly when utilising pink and green NH₃ as feedstocks. The H₂-BS scenarios for ship electrification are problematic for marine vessels due to shorter bunker intervals, increased volume requirements [33], and the absence of clearly defined safety protocols [97]. Additionally, the existing H₂ fuel production, storage, and distribution capacity is inadequate to support the comprehensive global deployment of marine vehicles utilising H₂ FC-based hybrid topologies [98]. Consequently, an industrial product such as using NH₃ as a H₂ carrier can address supply challenges by offering enhanced flexibility for these systems [15].

NH₃ is rich in H₂ and storable at ambient temperature and pressure and presents an effective medium for H₂ storage, even though it is toxic [99]. Nuclear energy has emerged as a significant source of NH₃ production in the analysis, characterised by low costs, and reduced emissions. Green NH₃, while following closely in emissions performance, is hindered by its higher production costs, resulting in its lower ranking than pink NH₃ options. Nuclear-sourced or pink NH₃ demonstrates significantly lower GHG emissions than the currently available and widely used industrial methods [100]. For instance, according to Bicer and Dincer [101], nuclear-based NH₃ production results in a significantly lower GWP compared to SMR. Specifically, SMR-based grey NH₃ production emits approximately 2.97 kg CO_{2e} per kilogram of NH₃, whereas nuclear-based production emits only about 0.23 kg CO_{2e} per kilogram, indicating a 92% reduction in GWP.

The maritime industry has started to see nuclear energy as a viable alternative, and the possible usage of onboard nuclear power is beginning to be considered [102]. Considering the public concerns regarding nuclear energy [103], challenges related to uranium sourcing and waste management issues [104], investments in green NH₃ are perceived as a more favourable option [105]. Currently, the Haber–Bosch method, commonly employed for NH₃ production, is highly energy-intensive and generates substantial CO₂ emissions. Although grey NH₃ may provide better economic performance, it does not meet emission targets, making such investments counterproductive. A strategic combination of blue, pink, and grey NH₃ will be essential to achieve forthcoming decarbonisation goals effectively. This integrated approach addresses economic viability and aligns with necessary environmental standards [106]. Enhanced carbon capture in blue NH₃ production and greater utilisation of green energy sources could further reduce costs and environmental impacts for both types of NH₃ [107].

The smaller plant (C2) investment is being prioritised in 2025 due to its immediate benefits. However, projections for 2040 indicate that the larger plant (C1) becomes more advantageous, providing greater long-term returns. The high capital costs of PEMFCs, which require multiple replacements during their lifespan, along with the expenses for WHRS and NH₃ cracking and storage, lead to a higher LCOE for electrification plants, making them less competitive than diesel engines in 2025 economic cases. Nonetheless, by 2040, substantially increased carbon taxes are expected to raise the LCOE of conventional engines, resulting in a more favourable LCOE for FC-based configurations compared to diesel engines [108].

Although there is growing interest in NH₃ as a fuel, the current technology is still theoretical and requires practical validation in real-world marine environments. For instance, utilising NH₃ in combustion engines presents significant challenges due to its corrosive properties, which can lead to the deterioration of critical engine components such as pistons, cylinders, and valves [14]. Furthermore, the storage and transportation of NH₃ necessitate stringent safety measures, as any leaks can pose severe health hazards. These factors underscore the need for careful consideration and robust engineering solutions when integrating NH₃ into existing fuel systems [109]. NH₃-powered smaller engines and

NH₃ as an H₂ carrier for FC-based ship electrification represent a more feasible initial stage for implementation [110,111].

5. Conclusions

The study evaluated two distinct PEMFC/battery/WHRS configurations within a marine power distribution system integrated with an NH₃ decomposition system, considering grey, blue, pink, and green NH₃ production methods. The main findings derived from this study can be listed as follows:

Environmental analysis demonstrated that the overall performance of the NH₃ decomposition system with pink and green NH₃ is superior to that of green H₂-BS.

- SO_x emissions mildly increased due to grey and blue NH₃, while green/pink NH₃ cracking performed better than H₂-BS.
- NO_x emissions slightly rose because of the onboard NH₃ decomposition system compared to H₂-BS.
- Economic analysis depicted that the NH₃ decomposition system is more economically viable than H₂-BS due to lower storage costs and fuel prices.
- Increased level of carbon penalties increased the economic viability of a larger PEMFC plant (C1) with pink/green NH₃-decomposition system.
- In the 2040-high cases, NH₃ cracking onboard was more cost-beneficial than the base case.

Limitations of the study are listed as follows:

- In the economic assessment, it has been postulated that the costs of equipment, for which future projections are lacking, will remain stable.
- The environmental analysis included only emissions from the NH₃ cracking system with available emission coefficients.
- The results are based on selected emission and efficiency coefficients sourced from current industrial data and literature; these represent the best available estimates at the time of the study. However, the outcomes may vary as these parameters are subject to change with future technological advancements in NH₃ production and related systems.
- This study uses a well-to-wake approach, excluding manufacturing emissions of FCs and batteries. As a result, the systems' environmental impact may be slightly underestimated due to their production emissions.
- The scrap values of FC and batteries have been excluded from the analysis.

Although based on a bulk carrier, this study's methodology offers insights applicable to other ship types, including container vessels and tankers. It is important to note that operational differences specific to ship types and equipment, such as tank heating or cargo handling, may affect load demand and influence emission reduction potential slightly.

This research adds to the existing literature by delivering an in-depth analysis of NH₃ decomposition systems for the onboard H₂ production utilised in PEMFCs within ship electrification systems. The impact of different NH₃ production methodologies on the environmental and economic performance was investigated in detail. The study's findings can benefit academicians working on maritime decarbonisation or NH₃ cracking areas, ship designers, and powertrain manufacturers willing to enhance systems for marine vessels.

Future research will investigate the integration of PV systems, power take-in/power take-off (PTI/PTO) technologies, and other suitable alternative energy sources alongside FCs and onboard hydrogen production systems. The optimisation of sizing for these systems will be a key focus of this investigation. A more comprehensive evaluation would benefit from incorporating full life cycle assessments that account for the embodied energy

and emissions from the production of battery and FC systems. Incorporating a detailed sensitivity analysis to evaluate the effects of price uncertainties on system performance and investment decisions would constitute a valuable direction for future research, offering deeper insights into the robustness and adaptability of the modelled scenarios.

Another research direction may prioritise a comprehensive risk assessment of the proposed system design, focusing on the bunkering, storage, and transfer of H₂ and NH₃ onboard. Identifying potential hazards and mitigating associated risks is vital for ensuring safety and regulatory compliance. Given that the analysis primarily considers theoretical aspects in an “ideal environment”, it is crucial to acknowledge that real marine conditions present challenges such as vibrations, impacts, and possible chemical interactions with NH₃.

Additionally, exploring various cracking systems and conducting thermodynamic analyses of NH₃ decomposition processes can enhance overall system efficiency. This application may aim to identify optimal methods for onboard H₂ production, facilitating integration into the broader energy system and promoting sustainability in NH₃ utilisation as a marine fuel. Furthermore, implementing FCs in conjunction with decomposition systems across different vessel types presents an opportunity for innovation. Assessing the integration of FCs with diverse vessel designs will provide insights into performance metrics, operational feasibility, and economic viability in various maritime contexts. Enhancing these areas is essential for the safe and efficient adoption of H₂ and NH₃ as alternative marine fuels, ultimately contributing to the decarbonisation of the shipping industry.

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Conflicts of Interest: Author Nikolaos Tsoulakos was employed by the company Laskaridis Shipping Co., Ltd. Milad Armin was the owner of the company Enki Marine Technology Consultancy. The remaining authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

Nomenclature

Abbreviation or Symbol	Explanation	Unit
A _i , A _j	Assessed alternatives in the sensitivity analysis	-
a _{j,k} , a _{i,k}	Normalised performance values of alternatives A _i and A _j for criterion k in the sensitivity analysis	-
AC	Alternating current	-
B _i	Cost index of beneficial criteria in COPRAS	-

C_a	Actual battery capacity	Ah
C_B	Available battery capacity	Ah
C_i	Cost index of non-beneficial criteria in COPRAS	-
C_{in}	Initial maximum capacity	Ah
C_{fuel}	Fuel prices	USD
$C_{o\&m}$	Operation and maintenance cost	USD
C_p	Specific heat rate of the exhaust	$\text{kJ}/\text{kg}\cdot\text{K}$
C_{plant}	Capital cost of systems	USD
CEPCI	Chemical Engineering Plant Cost Index	-
CH_4	Methane	-
CO_2	Carbon dioxide	-
CO_{2e}	Equivalent carbon dioxide	-
COPRAS	Complex Proportional Assessment	-
D/G	Diesel generator	-
DC	Direct current	-
e_j	Entropy	-
F	Feasible (in sensitivity analysis)	-
f_{ij}	The data point to be normalised in the entropy method	-
FC	Fuel cell	-
GHG	Greenhouse gas	-
GWP	Global warming potential for 100 years	-
H_2	Hydrogen	-
$\text{H}_2\text{-BS}$	Green hydrogen bunkering scenario	-
HFO	Heavy fuel oil	-
I	Current	A
IMO	International Maritime Organization	-
LCOE	Levelised cost of energy	-
LNG	Liquefied natural gas	-
LT	Plant lifetime	Years or hours
m	The number of rows in the entropy method	-
\dot{m}_{ex}	Exhaust mass flow rate	kg/s
M/E	Main engine	-
M/V	Motor vessel	-
MCDM	Multi-criteria decision-making	-
MDE	Marine diesel engines	-
N_2O	Nitrous oxide	-
n	Number of criteria in the entropy method	-
N_{ij}	Weighted normalisation matrix in COPRAS	-
n_{PEMFC}	Number of working PEMFCs	-
NF	Non-feasible (in sensitivity analysis)	-
NH_3	Ammonia	-
NO_x	Nitrogen oxides	-
OE	Operational (tank-to-wake) emissions	g or t g emission/g fuel)
OEC	Operational emission coefficient	-
ORC	Organic Rankine cycle	-
P_i, P_j	Aggregate scores for assessed alternatives in the sensitivity analysis in the weighted normalised decision matrix.	-
P_{gen}	The plant's generated total power	kW
P_{PEMFC}	Power output of one PEMFC	kW

PEMFC	Proton exchange membrane fuel cell	-
PV	Photovoltaic	-
r	Discount rate	-
R	The utilised parameter for each independent metric in the uncertainty analysis	-
r_{ij}	Normalised data in the entropy method	-
SFC	Specific fuel consumption	-
SMR	Steam methane reforming	-
SO _x	Sulphur oxides	-
SoC	Stage of charge	-
SOFC	Solid oxide fuel cell	-
SoH	State of health	-
Q _i	The relative significance of alternatives in COPRAS	-
t	Operation time	h
T _{in, ex}	The inlet temperature of the exhaust (after the exhaust boiler)	K
T _{out, ex}	Outlet temperature from ORC	K
U	Uncertainty	-
UD _i	Utility degrees in COPRAS	-
UE	Upstream (well-to-tank) emissions	g or t
UEC	Upstream emission coefficient	g UE/g OE or g/kWh
w _j	Criteria weights calculated in the entropy method	-
W _k	Criteria weights assessed in the sensitivity analysis	-
WHRS	Waste heat recovery system	-
Greek Symbols		
$\delta'_{k,i,j}$	Minimum required adjustment in weight of criterion k to reverse the ranking between alternatives in the sensitivity analysis	-
ΔH_0	Enthalpy change	kJ/mol
η_C	Columbic efficiency	-
η_{FC}	Efficiency of fuel cell	-
η_{ORC}	Organic Rankine cycle efficiency	-

Appendix A

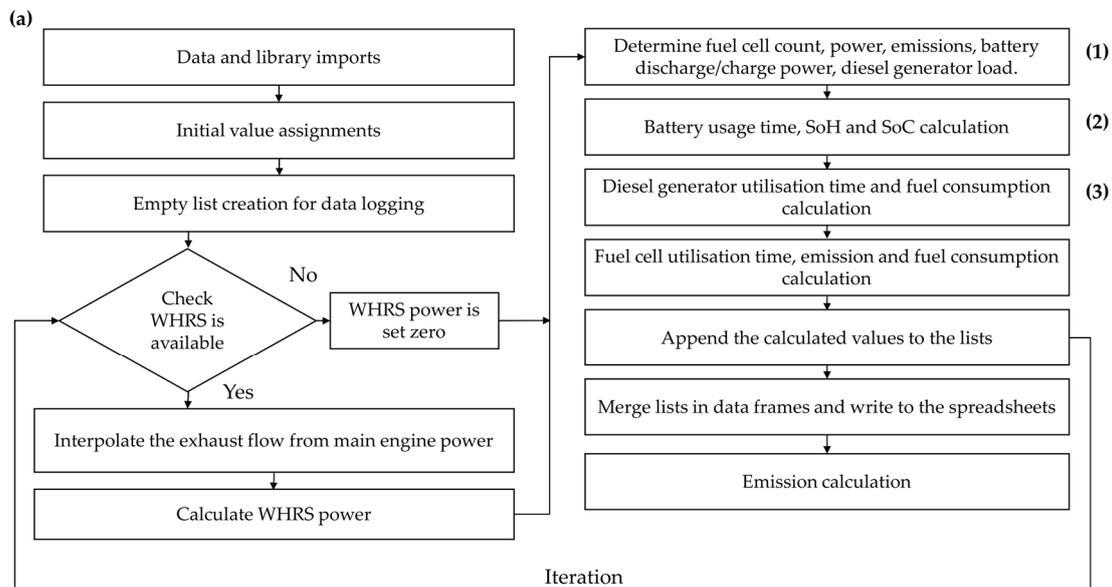


Figure A1. Cont.

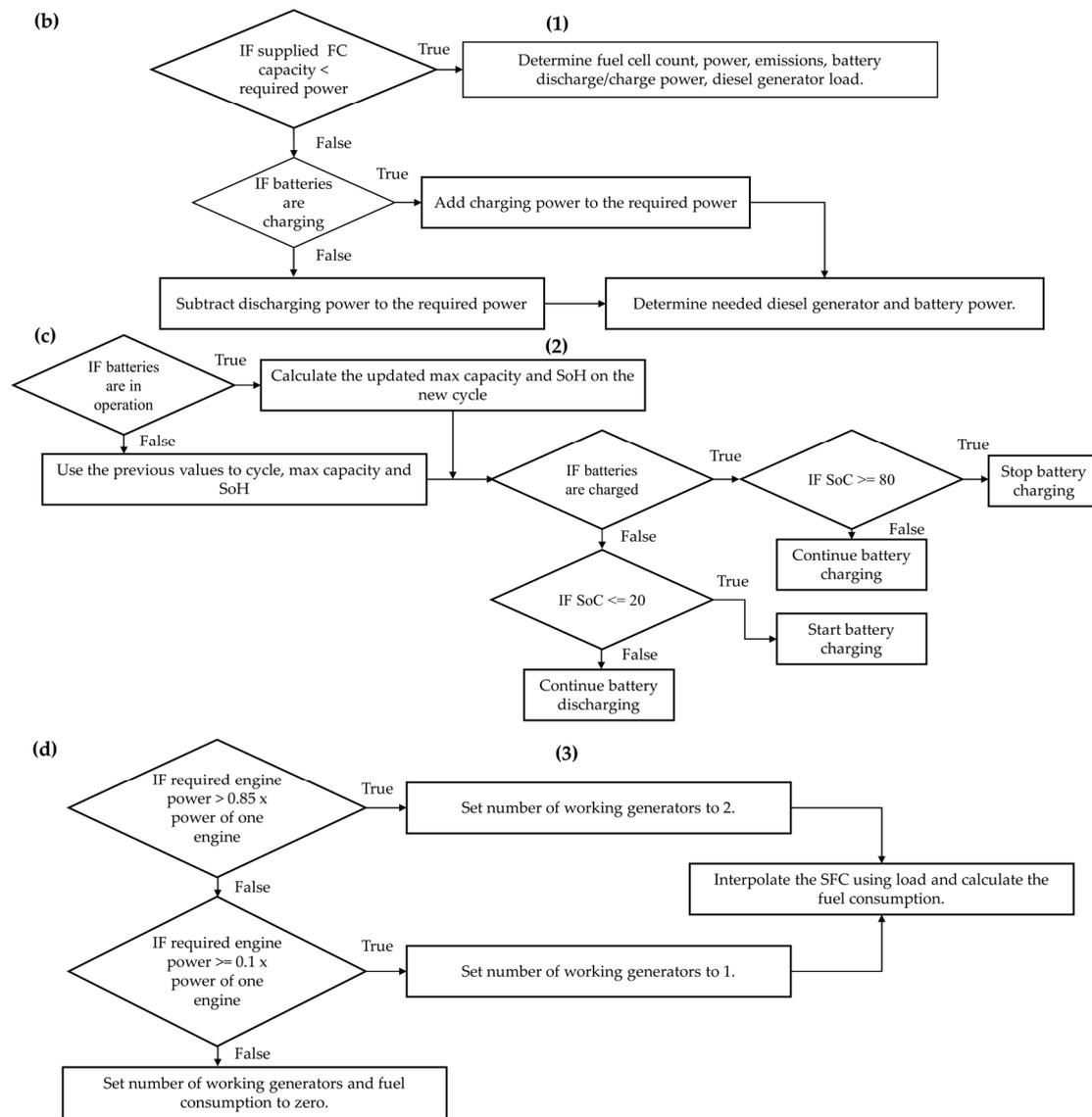


Figure A1. Simplified algorithm scheme: (a) general process, (b) FC determination details, (c) battery operation, and (d) D/G fuel consumption calculation.

Appendix B

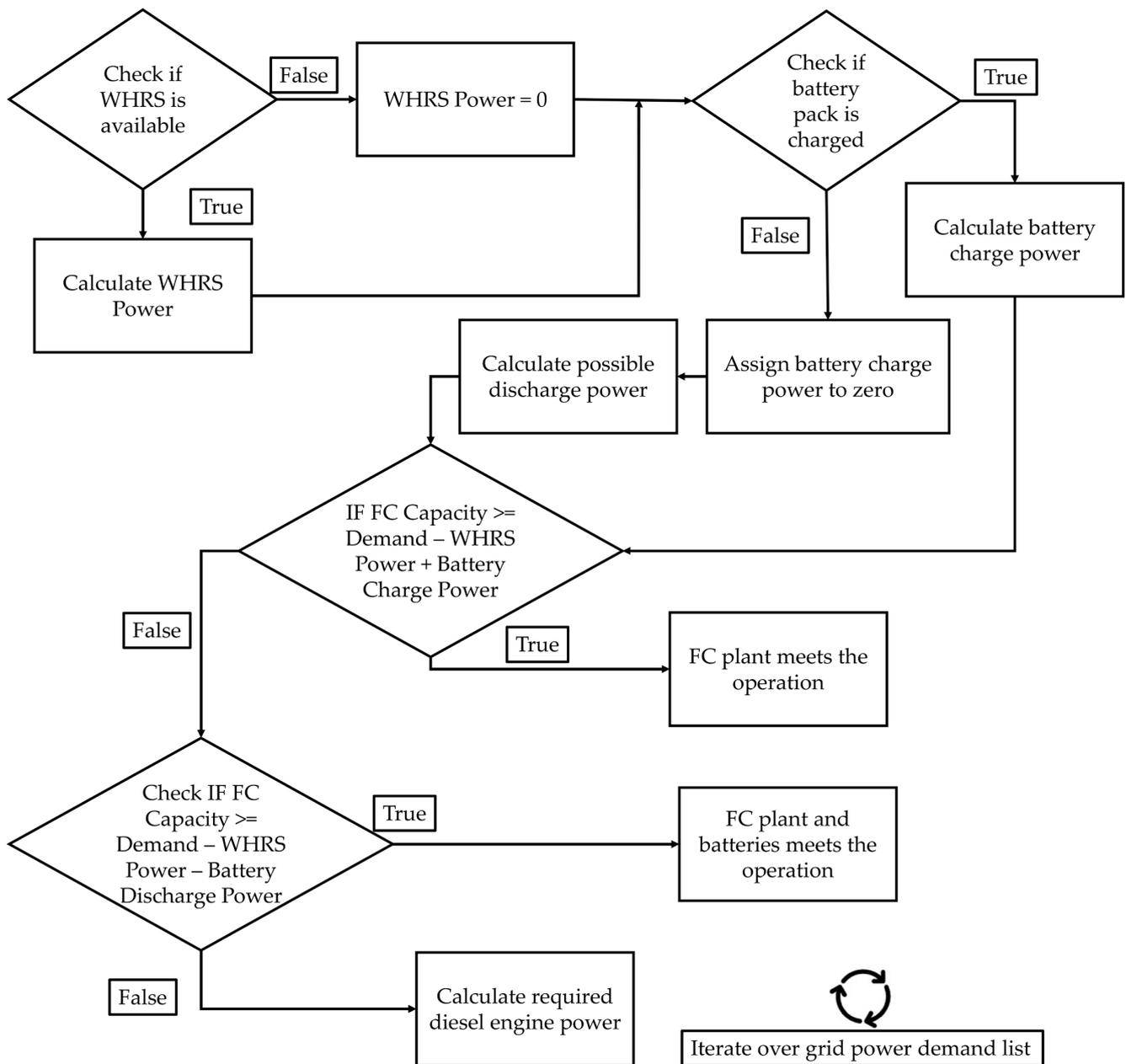


Figure A2. Simplified energy management strategy for establishing the hierarchy among power equipment.

Appendix C

Table A1. Installation, operation, and fuel costs of the configurations.

C _{plant}	FC		Battery		NH ₃ Storage P/T		H ₂ Storage P/T		HFO Storage		NH ₃ Cracker		WHRS		MDE		Total	
	2025	2040	2025	2040	2025	2040	2025	2040	2025/2040	2025	2040	2025/2040	2025/2040	2025	2040			
C1	6,097,104	3,129,703	685,840	442,367	9,435,707	6,086,031	8,860,660	8,283,109	0	0	2,648,824	1,708,491	1,348,579	1,348,579			622,987	
C2	5,080,920	2,608,086	234,401	151,189	9,053,576	5,839,556	6,625,170	8,205,986	14,530	14,530	2,541,550	1,639,300	1,348,579	1,348,579			1,245,974	
Base	N/A	N/A	N/A	N/A	N/A	N/A	N/A	N/A	700,157	N/A	N/A	N/A	1,868,962	2,569,119			2,569,119	
C _{O&M}	FC		Battery		NH ₃ Storage P/T		H ₂ Storage P/T		HFO Storage		NH ₃ Cracker		WHRS		MDE		Total	
	2024	2040	2024	2040	2024	2040	2024	2040	2024/2040	2024	2040	2024/2040	2024/2040	2024	2040			
C1	91,457	46,946	6858	4424	94,357	60,860	217,439	211,663	0	0	31,083	20,049	20,229	20,229			9345	
C2	76,214	39,121	2344	1512	90,536	58,396	130,668	125,004	145	145	31,083	20,049	20,229	20,229			18,690	
Base	N/A	N/A	N/A	N/A	N/A	N/A	N/A	N/A	7002	N/A	N/A	N/A	N/A	28,034	35,036			35,036
C _{fuel}	H ₂ —BS		Grey NH ₃		Blue NH ₃		Pink NH ₃		Green NH ₃									
	2024	2040	2024	2040	2024	2040	2024	2040	2024	2040	2024	2040	2024	2040	2024	2040—Low	2040—High	
C1	581,633	362,008	487,641	193,182	193,182	193,182	313,815	84,359	210,897	774,414	63,269	189,807	889,985	187,277			404,922	
C2	587,792	380,817	574,189	189,971	194,686	262,603	305,719	90,270	279,601	747,664	70,034	259,365	858,555	189,020			465,769	
Base	N/A	N/A	N/A	N/A	N/A	N/A	N/A	N/A	N/A	N/A	N/A	N/A	838,992	1,243,776			4,624,963	

Appendix D

Table A2. Normalised matrix and entropy results.

N _{ij}	CO _{2e}	Other Emissions	LCOE 2025	LCOE 2040 Low	LCOE 2040 High
Base	0.336	0.870	0.023	0.041	0.115
C1 H ₂ -BS	0.028	0.003	0.156	0.170	0.154
C1 Grey NH ₃	0.159	0.005	0.083	0.081	0.072
C1 Blue NH ₃	0.104	0.005	0.085	0.078	0.072
C1 Green NH ₃ -N	0.015	0.003	0.094	0.080	0.074
C1 Green NH ₃ -R	0.015	0.003	0.096	0.081	0.077
C2 H ₂ -BS	0.034	0.021	0.149	0.164	0.150
C2 Grey NH ₃	0.160	0.023	0.072	0.077	0.070
C2 Blue NH ₃	0.107	0.023	0.074	0.074	0.070
C2 Green NH ₃ -N	0.021	0.021	0.083	0.076	0.071
C2 Green NH ₃ -R	0.021	0.021	0.085	0.077	0.074

Table A2. Cont.

e_j	CO _{2e}	Other Emissions	LCOE 2025	LCOE 2040 Low	LCOE 2040 High
Base	-0.37	-0.12	-0.09	-0.13	-0.25
C1 H ₂ -BS	-0.10	-0.02	-0.29	-0.30	-0.29
C1 Grey NH ₃	-0.29	-0.03	-0.21	-0.20	-0.19
C1 Blue NH ₃	-0.24	-0.03	-0.21	-0.20	-0.19
C1 Green NH ₃ -N	-0.06	-0.02	-0.22	-0.20	-0.19
C1 Green NH ₃ -R	0.00	-0.02	-0.22	-0.20	-0.20
C2 H ₂ -BS	-0.11	-0.08	-0.28	-0.30	-0.28
C2 Grey NH ₃	-0.29	-0.09	-0.19	-0.20	-0.19
C2 Blue NH ₃	-0.24	-0.09	-0.19	-0.19	-0.19
C2 Green NH ₃ -N	-0.08	-0.08	-0.21	-0.20	-0.19
C2 Green NH ₃ -R	-0.08	-0.08	-0.21	-0.20	-0.19

Appendix E

Table A3. Sensitivity analysis results.

Comparison	CO _{2e}	SO _x + NO _x	LCOE—2025	CO _{2e}	SO _x + NO _x	LCOE-2040 L	CO _{2e}	SO _x + NO _x	LCOE 2040 H
A1-A2	-0.02	0.00	-0.10	-0.10	0.00	0.00	-1.19	-3.25	2.27
A1-A3	-0.27	0.00	0.09	-0.08	-0.23	0.20	-0.01	0.00	-0.09
A1-A4	-0.08	0.20	-0.11	-0.08	-0.20	0.16	-0.18	-0.46	0.24
A1-A5	0.04	0.12	-0.20	-0.19	0.00	0.01	-0.04	-0.04	-0.13
A1-A6	0.04	0.11	-0.19	-0.19	-0.04	0.04	-0.03	0.00	-0.16
A1-A7	-0.25	0.00	0.03	-0.03	0.00	-0.18	-0.15	-0.03	0.02
A1-A8	-0.20	0.04	-0.05	-0.04	-0.04	-0.14	-0.14	0.00	0.00
A1-A9	-0.30	0.00	0.03	-0.25	0.00	0.00	-0.20	-0.03	0.02
A1-A10	-0.26	0.03	-0.03	-0.25	-0.03	0.02	-0.20	0.00	0.00
A1-A11	-0.40	-1.07	0.08	-0.35	-0.95	0.05	-0.37	-1.01	-0.04
A2-A3	-0.45	0.01	0.24	-0.08	-0.24	0.21	0.30	0.85	-0.70
A2-A4	-0.10	0.26	-0.12	-0.08	-0.21	0.17	-0.01	0.00	-0.09
A2-A5	0.05	0.13	-0.21	-0.20	0.00	0.01	-0.03	0.00	-0.16
A2-A6	0.04	0.12	-0.20	-0.19	-0.04	0.04	-0.01	0.04	-0.19
A2-A7	-0.26	0.00	0.04	-0.03	0.00	-0.18	-0.14	0.00	0.00
A2-A8	-0.21	0.04	-0.04	-0.04	-0.04	-0.14	-0.13	0.03	-0.02
A2-A9	-0.31	0.00	0.03	-0.25	0.00	0.00	-0.19	0.00	0.00
A2-A10	-0.27	0.03	-0.03	-0.25	-0.03	0.02	-0.19	0.02	-0.01
A2-A11	-0.41	-1.10	0.08	-0.35	-0.95	0.05	-0.37	-0.99	-0.05

Table A3. Cont.

Comparison	CO _{2e}	SO _x + NO _x	LCOE—2025	CO _{2e}	SO _x + NO _x	LCOE-2040 L	CO _{2e}	SO _x + NO _x	LCOE 2040 H
A3-A4	0.12	0.41	-0.34	-0.03	0.00	-0.10	-0.51	-1.41	0.91
A3-A5	0.18	0.16	-0.32	-0.22	0.04	-0.04	-0.04	-0.04	-0.14
A3-A6	0.15	0.15	-0.30	-0.22	0.00	0.00	-0.03	0.00	-0.16
A3-A7	-0.25	-0.01	0.02	-0.02	0.04	-0.25	-0.16	-0.03	0.03
A3-A8	-0.19	0.04	-0.06	-0.03	0.00	-0.20	-0.15	0.00	0.01
A3-A9	-0.30	0.00	0.02	-0.27	0.03	-0.03	-0.21	-0.03	0.02
A3-A10	-0.26	0.03	-0.05	-0.27	0.00	0.00	-0.21	0.00	0.01
A3-A11	-0.41	-1.14	0.08	-0.38	-1.01	0.04	-0.38	-1.04	-0.04
A4-A5	0.21	0.00	-0.31	-0.22	0.04	-0.03	-0.03	0.00	-0.17
A4-A6	0.17	0.00	-0.28	-0.22	0.00	0.01	-0.02	0.04	-0.19
A4-A7	-0.31	-0.08	0.08	-0.02	0.04	-0.25	-0.15	0.00	0.01
A4-A8	-0.23	0.00	-0.03	-0.03	0.00	-0.20	-0.14	0.03	-0.01
A4-A9	-0.35	-0.05	0.06	-0.28	0.03	-0.03	-0.20	0.00	0.01
A4-A10	-0.30	0.00	-0.02	-0.28	0.00	0.00	-0.20	0.02	-0.01
A4-A11	-0.44	-1.24	0.10	-0.38	-1.03	0.04	-0.38	-1.03	-0.05
A5-A6	-0.02	0.00	-0.14	-0.17	-1.05	0.85	0.90	3.00	-2.22
A5-A7	-0.49	-0.10	0.22	2.30	0.05	-2.73	-0.59	-0.01	0.62
A5-A8	-0.34	-0.01	0.04	1.45	-0.33	-1.51	-0.53	0.12	0.50
A5-A9	-0.46	-0.06	0.14	-0.44	0.00	-0.02	-0.55	-0.01	0.34
A5-A10	-0.39	0.00	0.04	-0.43	-0.14	0.09	-0.53	0.07	0.28
A5-A11	-0.51	-1.36	0.15	-0.51	-1.90	0.10	-0.81	-2.27	0.08
A6-A7	-0.53	-0.11	0.25	4.96	1.23	-6.58	-0.66	-0.16	0.77
A6-A8	-0.35	-0.01	0.05	2.32	0.06	-2.78	-0.60	-0.01	0.63
A6-A9	-0.49	-0.07	0.15	-0.48	0.17	-0.16	-0.58	-0.09	0.41
A6-A10	-0.41	0.00	0.04	-0.46	0.00	-0.02	-0.56	-0.01	0.35
A6-A11	-0.52	-1.39	0.15	-0.52	-1.93	0.07	-0.83	-2.35	0.12
A7-A8	0.04	0.23	-0.39	-0.34	-1.13	1.05	0.39	2.36	-1.51
A7-A9	-0.43	0.00	0.02	-1.43	-0.02	0.96	-0.50	0.00	0.00
A7-A10	-0.29	0.10	-0.15	-1.35	-0.20	1.04	-0.46	0.16	-0.10
A7-A11	-0.51	-1.89	0.12	-0.73	-2.05	0.32	-0.92	-3.40	-0.19
A8-A9	-1.22	-0.39	0.71	-1.66	0.21	0.94	-0.57	-0.18	0.12
A8-A10	-0.54	0.00	0.02	-1.54	-0.02	1.04	-0.52	0.00	0.00
A8-A11	-0.63	-2.34	0.22	-0.74	-2.08	0.29	-0.96	-3.58	-0.15
A9-A10	0.01	0.32	-0.54	-0.11	-2.75	2.25	0.10	2.50	-1.60
A9-A11	-0.54	-2.61	0.15	-0.53	-2.60	0.14	-1.23	-5.89	-0.33
A10-A11	-0.66	-3.23	0.30	-0.54	-2.60	0.10	-1.30	-6.33	-0.26

Table A3. Cont.

Comparison	CO _{2e}	SO _x + NO _x	LCOE -2025	CO _{2e}	SO _x + NO _x	LCOE-2040 L	CO _{2e}	SO _x + NO _x	LCOE 2040 H
A5-A11	F	F	F	F	F	F	F	F	F
A6-A7	F	F	F	NF	NF	F	F	F	NF
A6-A8	F	F	F	NF	F	F	F	F	NF
A6-A9	F	F	F	F	NF	F	F	F	F
A6-A10	F	F	F	F	F	F	F	F	F
A6-A11	F	F	F	F	F	F	F	F	F
A7-A8	F	NF	F	F	F	NF	NF	NF	F
A7-A9	F	F	F	F	F	NF	F	F	F
A7-A10	F	F	F	F	F	NF	F	NF	F
A7-A11	F	F	F	F	F	F	F	F	F
A8-A9	F	F	NF	F	NF	NF	F	F	F
A8-A10	F	F	F	F	F	NF	F	F	F
A8-A11	F	F	F	F	F	F	F	F	F
A9-A10	F	NF	F	F	F	NF	F	NF	F
A9-A11	F	F	F	F	F	F	F	F	F
A10-A11	F	F	F	F	F	F	F	F	F
Comparison	CO _{2e}	SO _x + NO _x	LCOE -2025	CO _{2e}	SO _x + NO _x	LCOE-2040 L	CO _{2e}	SO _x + NO _x	LCOE 2040 -H
A1-A2	-4.16	0.00	-21.04	-26.52	0.00	-0.48	-309.56	-2433.07	N/F
A1-A3	-68.91	2.43	19.45	-21.95	-172.92	42.36	-3.31	0.00	-18.02
A1-A4	-20.57	N/F	-23.81	-19.88	-149.30	33.79	-45.67	-342.08	49.74
A1-A5	10.97	86.42	-41.15	-50.43	-2.73	1.07	-10.18	-27.36	-27.94
A1-A6	9.51	78.97	-40.13	-50.22	-30.63	7.38	-7.25	-0.10	-33.16
A1-A7	-64.74	-3.50	6.22	-7.22	-0.10	-37.03	-39.31	-23.91	4.54
A1-A8	-50.98	28.82	-9.82	-9.70	-26.13	-29.14	-37.60	-2.03	0.61
A1-A9	-76.77	-2.58	5.63	-64.02	-2.15	-0.15	-52.60	-20.40	3.99
A1-A10	-68.19	23.22	-6.96	-63.63	-24.74	5.01	-51.82	-1.73	0.64
A1-A11	-103.13	-800.74	16.36	-90.91	-706.10	10.42	-97.16	-752.76	-8.21
A2-A3	-117.76	4.27	49.99	-21.74	-180.60	44.26	77.16	N/F	-147.46
A2-A4	-25.29	N/F	-24.60	-19.63	-154.99	35.09	-2.50	0.00	-19.77
A2-A5	13.23	99.34	-44.16	-50.61	-2.75	1.08	-6.75	0.24	-33.70
A2-A6	11.35	89.63	-42.70	-50.39	-30.84	7.43	-3.83	27.47	-38.91
A2-A7	-68.51	-3.72	7.92	-7.09	-0.10	-37.28	-36.85	-1.99	0.26
A2-A8	-53.33	30.27	-9.26	-9.59	-26.30	-29.32	-35.15	19.82	-3.65
A2-A9	-80.04	-2.69	6.84	-64.23	-2.16	-0.15	-50.61	-1.69	0.34
A2-A10	-70.76	24.15	-6.39	-63.84	-24.88	5.05	-49.84	16.93	-3.00
A2-A11	-105.72	-821.69	17.34	-91.15	-708.65	10.46	-95.79	-741.92	-11.32
A3-A4	31.76	N/F	-70.62	-6.84	0.00	-20.39	-133.92	-1054.73	N/F
A3-A5	45.61	N/F	-67.43	-56.21	31.78	-7.30	-10.58	-28.93	-28.51
A3-A6	39.47	N/F	-62.89	-55.71	-3.01	0.59	-7.48	-0.10	-34.02
A3-A7	-64.08	-4.44	4.13	-4.47	32.16	-51.85	-40.94	-25.00	5.56

Table A3. Cont.

Comparison	CO _{2e}	SO _x + NO _x	LCOE -2025	CO _{2e}	SO _x + NO _x	LCOE-2040 L	CO _{2e}	SO _x + NO _x	LCOE 2040 -H
A3-A8	-48.73	32.13	-13.48	-7.50	0.26	-41.99	-39.13	-2.12	1.45
A3-A9	-77.64	-3.13	4.09	-70.46	24.00	-6.66	-54.49	-21.18	4.83
A3-A10	-68.12	25.27	-9.56	-69.93	-2.35	-0.63	-53.66	-1.80	1.35
A3-A11	-105.29	-851.36	16.16	-97.29	-755.37	7.47	-100.15	-776.69	-7.90
A4-A5	54.86	0.73	-65.29	-57.85	32.83	-6.86	-7.07	0.26	-34.75
A4-A6	43.67	0.59	-58.66	-57.26	-3.11	1.25	-3.93	29.52	-40.34
A4-A7	-80.47	-57.44	16.91	-4.40	33.14	-52.80	-38.88	-2.10	1.44
A4-A8	-59.26	-3.64	-6.01	-7.52	0.27	-42.63	-37.05	20.98	-2.71
A4-A9	-90.23	-38.63	12.68	-72.04	24.59	-6.32	-53.00	-1.78	1.34
A4-A10	-78.12	-2.79	-3.44	-71.48	-2.40	-0.15	-52.17	17.77	-2.17
A4-A11	-113.76	-922.90	21.53	-98.63	-766.58	7.88	-99.62	-772.38	-10.98
A5-A6	-5.98	0.00	-29.28	-44.58	-781.22	N/F	N/F	N/F	-463.50
A5-A7	-126.99	-77.44	45.17	N/F	36.61	-568.53	-152.49	-10.53	N/F
A5-A8	-87.03	-4.71	8.42	N/F	-247.13	-314.36	-138.03	90.77	N/F
A5-A9	-120.37	-46.81	28.88	-114.42	0.00	-4.69	-142.12	-5.72	71.37
A5-A10	-101.54	-3.41	7.45	-110.15	-102.34	18.91	-137.06	50.72	59.15
A5-A11	-130.95	-1017.04	30.38	-131.70	-1414.73	19.84	-210.25	-1695.82	17.44
A6-A7	-137.15	-83.94	51.43	N/F	N/F	-1371.30	-171.62	-122.21	N/F
A6-A8	-91.74	-4.98	10.61	N/F	41.67	-580.05	-155.35	-9.52	N/F
A6-A9	-126.00	-49.11	31.74	-125.59	N/F	-33.75	-152.01	-64.92	85.43
A6-A10	-105.49	-3.55	8.97	-120.04	0.00	-4.92	-146.52	-5.21	72.47
A6-A11	-133.89	-1040.97	31.78	-135.09	-1439.38	13.72	-217.37	-1759.02	25.15

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